

PFAS

PER- AND POLY-FLUOROALKYL SUBSTANCES

2^e Congrès International
2nd International Congress

Gestion des Risques Environnementaux & Sanitaires
Management of Environmental & Health Risks

13 & 14/06/2023 - Paris

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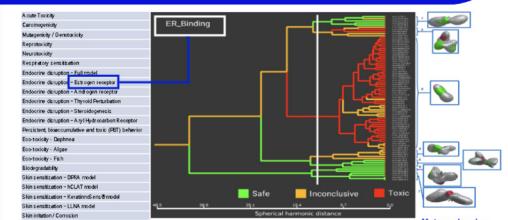
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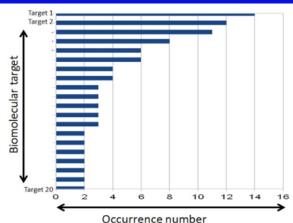


Predictive toxicology



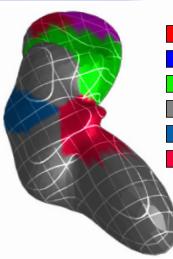
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⁽¹⁾New Alternative Method - ⁽²⁾Quantitative Structure Activity Relationship - ⁽³⁾Mechanism of Action - ⁽⁴⁾Format based on the mathematical methods from which the company takes its name - ⁽⁵⁾Substances e.g. ingredients, reaction intermediates, and APIs - ⁽⁶⁾The Occupational Exposure Band (OEB) classes



Programme

Mardi 13 juin 2023

Présidents :

- Dr. Frank Karg, SFSE & HPC International
- Me Laurence Lanoy, Avocate en Droit de l'Environnement

08h45

Accueil des participants

Contexte réglementaire - Investigations - Risques toxicologiques

09h10

Composés per- et polyfluoroalkylés (PFAS) dans les milieux aquatiques français : distribution spatiale et profils dans les sédiments & facteurs de contrôle écologiques chez les poissons

Pierre Labadie, Directeur de Recherche - CNRS

09h30

Les évolutions du cadre juridique et réglementaire européen et français concernant les substances PFAS

Me Laurence Lanoy, Avocate Associée - Laurence Lanoy Avocats

09h55

PFAS : Chimie Environnementale, Diagnostics & Identification des Sources, Toxicologie et Evaluation des Risques (EQRS)

Dr. Frank Karg / Président HPC International SAS/ Directeur Scientifique Groupe HPC AG - ARET, SFSE et INOGEN & IBS - Expert Judiciaire (France & Allemagne)

10h25

Sites pollués par des PFAS : état des problèmes et des solutions en développement en Suisse sur la base des expériences du canton de Genève

- *Christiane Wermeille, cheffe de la section Sites contaminés - Office Fédéral de l'Environnement (Suisse)*
- *Alain Davit, chef du secteur Sites pollués au service de Géologie, Sols et Déchets de la République et canton de Genève (Suisse)*

10h45

Pause-café / Thé

11h00

Surveillance des PFAS dans les eaux souterraines en France : Etat des lieux et enjeux

Julie Lions, Hydrogéologue - Bureau de Recherches Géologiques et Minières, BRGM

11h20

Occurrence des PFAS dans l'eau potable au Canada et à travers le monde

Sébastien Sauvé, Professeur Titulaire - Université de Montréal (Canada)

11h40

Analyse avancée des empreintes des PFAS dans les eaux souterraines : utilisation de statistiques multivariées avancées et des techniques d'apprentissage automatique

Dr. Carlo Monti, Directeur Général, Pratique médico-légale - TIG Environmental (Suisse)

12h05

Pollution des eaux souterraines par les PFAS dans la région de Vénétie (Italie du Nord) : implications réglementaires sur les atteintes à l'environnement et jugement

Dr. Paolo Ronco, Responsable de la recherche, de l'innovation et de la durabilité - Centro RIVE, Centre des ressources en eau de Vénétie (Italie)

12h30

Déjeuner

14h00

Estimation de l'exposition des jeunes enfants à deux composés perfluorés (PFOS et PFOA) à l'aide de la modélisation PBPK et des mesures de biomarqueurs

Aude Ratier, Ingénieur Modélisation Toxicocinétique - INERIS

14h20

Les PFAS et leurs impacts sur la santé humaine : exemple avec une étude sur l'exposition aux PFAS et contrôle de la glycémie chez les adultes américains âgés atteints de diabète de type 2

Eloïse Brosset, Doctorante - Centre de Recherche du Centre Hospitalier Universitaire de Sherbrooke, CRCHUS (Canada)

14h40

Combinaison des modèles QSAR et du clustering chimique – application à l'évaluation toxicologique des PFAS

- *Philippe Oudin, Gérant - Semaco Environnement & Professeur à l'Ecole des Mines de Nancy*
- *Sinan Karaboga, Directeur scientifique et commercial - Harmonic Pharma*

15h00

Exposition aux PFAS chez les enfants canadiens et ses déterminants

Nolwenn Noisel, Professeure adjointe - Université de Montréal (Canada)

15h20

Pause-café – Thé

15h40

Ecotoxicité des mélanges de composés perfluorés sur la faune du sol et possibles solutions de traitement

Hugo Carronnier, Chef de Projet Evaluation des risques - Valgo

16h00

Évaluation de la mobilité et de la remédiation des PFAS dans les sols contaminés

Muhammad Usman, Chercheur Post-doc - Ecole Nationale Supérieure de Chimie de Rennes

16h20

Relever les défis des PFAS dans l'eau : retour d'expérience de Veolia en Amérique du Nord et en Australie

Anna Catafau, Chef de projet Innovation - Veolia

16h40

Élaboration d'un plan de remédiation soutenu par la société pour un important hotspot de PFAS en Belgique : défis rencontrés et leçons apprises à ce jour

Jan Van Linden, Associé - ERM (Belgique)

17h00

Mécanismes de sorption des PFAS - principes et problèmes d'évaluation des risques de migration

Lars Van Passel, Directeur Technique - RSK Benelux (Belgique)

17h20

Guide et fiches de gestion des PFAS de la SFSE

- *Dr. Frank Karg, section ERIS : Méthodologie d'Evaluation des Risques et Impacts Sanitaires - SFSE & HPC International (France)*
- *Marie Jailler, Secrétaire de la section ERIS : Méthodologie d'Evaluation des Risques et Impacts Sanitaires - SFSE & Expert senior au sein du service des évaluations des risques sanitaires et environnementaux - SPAQuE (Belgique)*

17h40

Screening de PFAS semi quantitatifs dans les eaux : retour d'expérience

Sylvain Mourard, Responsable Commercial France Eau & Environnement - Mérieux NutriSciences

18h00

Discussion & Conclusions:

Dr. Frank Karg, HPC International - Inogen, Expert judiciaire (France & Allemagne)

18h15

Fin de la première journée



Program

Tuesday, June 13, 2023

Chairs:

- Dr. Frank Karg, SFSE & HPC International
- Me Laurence Lanoy / Environmental Lawyer

08:45

Welcoming participants

Legal Context - Investigations - Toxic Risks

09:10

Per- and polyfluoroalkyl substances (PFAS) in French aquatic environments: spatial distribution and patterns in sediments & ecological drivers in fish

Pierre Labadie, Research Manager - CNRS

09:30

PFAS: Regulatory and legal management of pollution in European Community and in France

Me Laurence Lanoy, Associate Lawyer - Laurence Lanoy Avocats

09:55

PFAS: Environmental Chemistry Investigations, Source Identification, Toxicology and TERQ Risk Assessments, including FTOH

Dr. Frank Karg / President HPC International SAS / Scientific Director HPC AG Group - ARET& SFSE and INOGEN & IBS, Judicial Expert (France & Germany)

10:25

Sites polluted by PFAS: state of the problems and solutions under development in Switzerland based on the experiences of the canton of Geneva

- *Christiane Wermeille, Head of the Contaminated Sites Section - Federal Office for the Environment (Switzerland)*
- *Alain Davit, Head of the Polluted Sites Sector in the Geology, Soil and Waste Department of the Republic and Canton of Geneva (Switzerland)*

10:45

Break for Coffee and Tea

11:00

Groundwater monitoring of PFAS in France: Current status and challenges

Julie Lions, Hydrogeologist - French Geological Survey, BRGM

11:20

Occurrence of PFAS in drinking water in Canada and across the world

Sébastien Sauvé, Professor - University of Montreal (Canada)

11:40

Advanced fingerprint analysis of PFAS in groundwaters: the use of advanced multivariate statistics and machine learning techniques

Dr. Carlo Monti, Managing Director, Forensics Practice - TIG Environmental (Switzerland)

12:05

PFAS groundwater pollution in Veneto Region (North Italy): regulatory implications on environmental harm and judgment case

Dr. Paolo Ronco, Head of Research, Innovation and Sustainability - Centro RIVE, Center for Water Resources of Veneto (Italy)

12:30

Lunch Break

14:00

Estimating the early-life exposure to two perfluorinated compounds (PFOS and PFOA) using PBPK modeling and biomarker measurements

Aude Ratier, Toxicokinetic Modeling Engineer - INERIS

14:20

PFAS and their Impacts on human health: An example of a study of PFAS exposure and blood glucose control in older american adults with type 2 diabetes

Eloïse Brosset, Doctoral student - Sherbrooke University Hospital Research Center, CRCHUS (Canada)

14:40

Combination of QSAR models and chemical clustering – application to the PFAS' toxicology assessment

- *Philippe Oudin, CEO - Semaco Environnement & Professor at the Ecole des Mines de Nancy*
- *Sinan Karaboga, Scientific and Commercial Director - Harmonic Pharma*

15:00

Exposure to PFAS among canadian children and its déterminants

Nolwenn Noisel, Assistant Professor - University of Montréal (Canada)

15:20

Break for Coffee and Tea

15:40

Ecotoxicity of mixtures of perfluorinated compounds on soil fauna and possible treatment solutions
Hugo Carronnier, Risk Assessment Project Manager - Valgo

16:00

Evaluating the mobility and remediation of PFAS in contaminated soils

Muhammad Usman, Post-doc Researcher - Ecole Nationale Supérieure de Chimie de Rennes

16:20

Addressing PFAS challenges in water: Veolia return of experience in North America and Australia

Anna Catafau, Innovation Project Manager - Veolia

16:40

Developing a societal supported remedial plan for an important PFAS hotspot in Belgium: challenges faced and lessons learned to date

Jan Van Linden, Associate Partner - ERM (Belgium)

17:00

PFAS sorption mechanisms - migration risk assessment principles and issues

Lars Van Passel, Technical Director - RSK Benelux (Belgium)

17:20

SFSE PFAS management guide and files

- *Dr. Frank Karg, Health Risk & Impact Assessment Methodology Section - SFSE & HPC International (France)*
- *Marie Jailler, Secretary of the Health Risk & Impact Assessment Methodology Section - SFSE and Senior Expert in the Health and Environmental Risk Assessment Department - SPAQuE (Belgium)*

17:40

Screening of semi-quantitative PFAS in waters: feedback

Sylvain Mourard, Sales Manager France, Water & Environment - Mérieux NutriSciences

18:00

Discussion & Conclusions:

Dr. Frank Karg, HPC International - Inogen, Judicial Expert (France & Germany)

18:15

End of day one

**Composés per- et polyfluoroalkylés (PFAS) dans les milieux aquatiques français :
distribution spatiale et profils dans les sédiments & facteurs de contrôle écologiques
chez les poissons**

Per- and polyfluoroalkyl substances (PFAS) in French aquatic environments: spatial distribution and patterns in sediments & ecological drivers in fish

Auteurs : Pierre Labadie¹, Nicolas Macorps¹, Karyn Le Menach¹, Patrick Pardon¹, Sabrina Guérin-Rechdaoui², Vincent Rocher², François Lestremau^{3,4}, Azziz Assoumani³, Hélène Budzinski¹

1 : CNRS/Université de Bordeaux, UMR 5805 EPOC, Talence, France

2 : SIAAP, Direction Innovation, Colombes, France

2 : SIAAP, Direction Innovation, Colombes, France

3 : INERIS, Unité Méthodes et développements en Analyses pour l'Environnement, 60550 Verneuil-en-Halatte, France

4 : Hydrosciences Montpellier, Université de Montpellier, IMT Mines Ales, IRD, CNRS, Ales, France

Contact : Pierre LABADIE, UMR 5805 EPOC, Equipe LPTC, Bâtiment A12, Université de Bordeaux, 351 crs de la Libération, 33405 Talence ; E-mail : pierre.labadie@u-bordeaux.fr

Caractère innovant :

Ces travaux ont dressé un état des lieux de la contamination du réseau hydrographique (sédiment) par les PFAS à l'échelle nationale, en ciblant des composés historiques ou d'intérêt très émergent et en ayant recours à la méthode TOP pour estimer la contribution de composés non identifiés. Un focus a également été fait sur la bioaccumulation des PFAS chez une espèce de poissons sentinelle dans la Seine : le chevaine *Squalius cephalus*.

Mots clés :

PFAS ; milieux aquatiques ; sédiment ; poisson ; biosurveillance

Objectifs : documenter les niveaux et profils de contamination dans deux compartiments intégrateurs au sein des hydrosystèmes (étude comparative PFAS historiques vs émergents) : sédiment et poissons. Pour les sédiments, il s'agissait de déterminer les niveaux et la distribution spatiale d'une large gamme de PFAS, la présence de PFAS non identifiés dans les sédiments de surface et l'influence des activités anthropiques sur les niveaux et profils de contamination. Concernant le volet biosurveillance, nous avons cherché à étudier les facteurs de contrôle potentiels de la bioaccumulation des PFAS et de la variabilité intraspécifique chez une espèce de poisson modèle fréquemment utilisé pour estimer la qualité chimique des masses d'eau continentales.

Résumé

Les composés per- et polyfluoroalkyles (PFAS) sont désormais omniprésents dans les environnements aquatiques. L'apparition de nouveaux PFAS requiert l'acquisition de nouvelles connaissances sur leur présence et leur devenir.

Dans cette étude, 37 PFAS historiques et d'intérêt émergent ont été recherchés dans des sédiments collectés à l'échelle nationale française ($n = 43$) ; cette sélection couvrait une large gamme de composés : molécules neutres, anioniques et zwittérioniques. En outre, une procédure de type TOP (Total Oxidizable Precursor Assay) a été mise en œuvre pour estimer la fraction non identifiée des PFAS dans ces échantillons. Les PFAS ont été détectés dans 86 % des échantillons et la teneur totale en PFAS ciblés était comprise entre la limite de détection et 23 ng/g de poids sec (médiane : 1,3

ng/g). Le profil moléculaire était souvent dominé par le PFOS mais des composés d'intérêt émergent tels que les 6:2 FTAB et 8:2 FTAB (fluorotélomères sulfonamidoalkyl-bétaïnes) ont respectivement été détectés dans 38 % et 24 % des échantillons, à des niveaux similaires à ceux du PFOS. Une analyse par ascendance hiérarchique a révélé des similitudes entre groupes de sites d'échantillonnage. Par exemple, une contribution élevée des FTAB a été associée à la proximité d'activités aéroportuaires où des mousses anti-incendie de type AFFF ont pu être utilisées. En outre, il est apparu que les teneurs en PFAS inconnus estimées via la méthode TOP étaient fortement corrélées avec la teneur totale en PFAS ciblés, avec des niveaux généralement plus élevés près des zones industrielles et urbaines. La médiane de leur contribution à la teneur totale en PFAS était de 58 %, ce qui souligne les efforts d'identification qui demeurent nécessaires à une caractérisation plus complète de la contamination des hydrosystèmes par les PFAS.

Pour appréhender cette dernière, des approches de biosurveillance peuvent être mise en œuvre grâce à des organismes sentinelles, notamment le chevaine *Squalius cephalus*. Nos travaux ont ici porté sur la compréhension des déterminants de la bioaccumulation des PFAS chez cette espèce de poisson, fondamentale pour l'évaluation des risques. Or, très peu d'études ont abordé l'influence combinée de traits de vie ou facteurs écologiques sur ce processus (ex : âge, sexe ou écologie trophique). Ce travail visait ainsi à combler ces lacunes à travers une étude de cas dans le bassin de la Seine (France), fortement anthropisé. Les poissons collectés en amont de Paris sont apparus significativement moins contaminés que ceux collectés en aval, en lien avec l'intensité des apports anthropiques. Les PFAS d'intérêt émergent n'ont que peu voire pas été détectés, en bon accord avec leur caractère plus faiblement bioaccumulable ou leur plus faible émission cumulée. Le profil de contamination était largement dominé par des composés historiques : acides carboxyliques et sulfoniques. La norme de qualité environnementale définie pour le PFOS dans le cadre de la Directive Européenne sur l'Eau a été ponctuellement dépassée, mais chez quelques individus seulement. L'influence de l'écologie trophique et de la biométrie sur la variabilité interindividuelle des niveaux de PFAS chez les poissons a démontré, qu'outre le site d'échantillonnage, le régime alimentaire était également influent : les sources de carbone exploitées à l'échelle individuelle et le niveau trophique expliquent en partie la variabilité de la contamination.



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PFAS : Regulatory and legal management of pollution

Per- and polyfluoroalkylated substances (PFAS) belong to a large family of endocrine disruptors used in many industrial processes and in the manufacture of everyday objects. Due to their non-stick, heat-resistant and waterproofing properties, these unique substances are persistent and harmful pollutants to the environment and human health, as they are found in drinking, ground and surface water, soil, air and food.

Awareness of these substances and their effects is both recent and significant: measures to monitor and limit PFAS in the environment are needed to restrict the use of these compounds and minimize exposure and risks to human health.

Regulations on PFAS are developing more or less rapidly depending on the sector.

In the water sector, as part of the recent transposition into French law of the European Directive of 16 December 2020, known as the "Drinking Water Directive", which reinforces its monitoring and sets limit values for around twenty of these substances in water intended for human consumption, the government published its action plan against PFAS on 17 January 2023. This document will have an impact on the obligations of the operators of ICPEs affected by these issues, through the various lines of action it proposes.

In this context, a ministerial decree currently under consultation provides for the analysis and monitoring of PFAS substances in the aqueous effluents of ICPEs.

In parallel, a proposal for a REACH restriction of all PFAS substances was recently submitted to ECHA by five European countries (Germany, Denmark, the Netherlands, Sweden and Norway), supported by France.

Laurence Lanoy, a lawyer specializing in environmental law, will give an overview of the changes in the regulatory framework concerning these substances of concern, presenting the issues relating to their management for the industrial sector.

* * *

Attorney at law since 1990 and Ph.D in law, Laurence Lanoy has developed an in-depth practice in environmental law before founding the firm Laurence Lanoy Avocats in 2005. She advises and assists national and international companies, public authorities, and international law firms, particularly in environmental and sustainable development law, mining law and energy law.



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Les évolutions du cadre juridique et réglementaire concernant les substances PFAS

Les substances per- et polyfluoroalkylées (PFAS) appartiennent à une grande famille de perturbateurs endocriniens utilisés dans de nombreux processus industriels et la fabrication d'objets de la vie quotidienne. En effet, par leurs propriétés antiadhésives, résistantes aux fortes chaleurs et imperméabilisantes, ces substances uniques constituent des polluants persistants et nocifs pour l'environnement et la santé humaine puisqu'elles se retrouvent dans les eaux potables, souterraines ou de surface, les sols, l'air ou encore l'alimentation.

La prise de conscience relative à ces substances et leurs effets est à la fois récente et majeure : des mesures de surveillance et de limitation des PFAS dans l'environnement s'imposent afin de restreindre l'usage de ces composés et de réduire au maximum l'exposition et les risques pour la santé humaine.

La réglementation des PFAS évolue plus ou moins rapidement selon les secteurs.

En matière d'eau, dans le contexte de la récente transposition en droit français de la directive européenne du 16 décembre 2020 dite « eau potable », qui renforce leur suivi et fixe des valeurs limites dans l'eau destinées à la consommation humaine pour une vingtaine de ces substances, le Gouvernement a publié le 17 janvier 2023 son plan d'action contre les PFAS. À travers les différents axes d'action qu'il déploie, ce document va avoir un impact sur les obligations des exploitants des ICPE qui sont concernés par ces problématiques. À cet égard, un arrêté ministériel actuellement en consultation prévoit l'analyse et la surveillance des substances PFAS dans les rejets aqueux des ICPE.

En parallèle, une proposition de restriction REACH de toutes les substances PFAS a été récemment présentée à l'ECHA par cinq pays européens (Allemagne, Danemark, Pays-Bas, Suède et Norvège) avec le soutien de la France.

Laurence Lanoy, avocat spécialiste en droit de l'environnement, effectuera un tour d'horizon des évolutions du cadre réglementaire concernant ces substances préoccupantes, en présentant les enjeux relatifs à leur gestion pour le secteur industriel.

* * *

Avocat depuis 1990 et Docteur en droit, Laurence Lanoy a développé une pratique approfondie en droit de l'environnement avant de fonder en 2005 le cabinet Laurence Lanoy Avocats. Elle conseille et assiste des entreprises nationales et internationales, des collectivités publiques et des cabinets d'avocats internationaux notamment en droit de l'environnement et du développement durable, en droit minier et en droit de l'énergie.

PFAS: Environmental Chemistry Investigations, Contamination Source Identification, Toxicology and TERQ Risk Assessments, including FTOH

PFAS : Chimie Environnementale, Diagnostics & Identification des Sources de Pollution, Toxicologie et Evaluation des Risques (EQRS), incluent les FTOH

Frank KARG¹, Ulrike HINTZEN², Lucie ROBIN-VIGNERON³, Philippe GIRARD⁴

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1. Introduction

Les PFAS : Per- & Polyfluoro-Alkyl Substances sont devenues progressivement depuis des années 1960 un problème environnemental majeur, également pour la Santé Publique, en raison de leur application multiple et vaste (historique et toujours actuelle). Cette menace environnementale et pour la Santé Publique commence à être prise en compte petit à petit depuis les années 2010 et fortement en 2022 & 2023. Par conséquent, les PFAS se retrouvent aujourd’hui dans les sols, les eaux souterraines, les aliments et l’eau potable ainsi que dans les Gaz du sol et de l’Air ambiant pour une famille de PFAS volatils, les FTOH : Fluorotélomère-Alcools. Entre 9 000 à 12 000 polluants synthétiques de PFAS ont été produits.

Les PFAS polymères du type « Teflon » (ou PFTE) etc. ne disposent pas d'une bonne biodisponibilité et sont donc par conséquent beaucoup moins toxiques que les PFAS monomères. Ces PFAS monomères font l'objet des travaux présentés ci-joint. Les PFAS sont connus notamment pour leurs effets toxicologiques de perturbateurs endocriniens, d'hépatotoxicité, d'Immunotoxicité, leurs effets sur le développement des fœtus et pour certains, la cancérogénicité (par ex. le PFOA) [1 - 84].

Une caractéristique importante des PFAS est leur comportement dans la Chimie Environnementale, car seuls les PFAS polyfluorés sont modifiés par bio-transformation microbiologique en PFAS perfluorés, qui demeurent totalement stables et non-dégradables, voir même bioaccumulables.

Les sources de Pollutions par PFAS sont multiples et notamment présentes sur les sites industriels, qui ont utilisés ces produits, les sites d'anciens incendies ou d'entraînement anti-incendie, où des mousses anti-incendie (**AFFF : Anti Fire Fighting Foams, par ex. sur des**

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aéroports) ont été utilisées. Les terrains agricoles sont aussi des sources de pollution par des PFAS, en raison de l'apport de boues de STEP : Stations d'Epuration qui comportent des PFAS accumulés.

Les activités (historiques) suivantes peuvent être à l'origine de pollutions par des PFAS :

- Entraînements anti-incendie,
- Aéroport ou base aérienne site militaire,
- Site d'incendie et utilisation des AFFF,
- Galvanisation électrochimique,
- Production de papiers ou cartons « cirés »,
- Production de Textiles imperméables,
- Sprays, peintures, laques d'imperméabilisation,
- Production et application de Teflons (PTFE, etc.),
- Sites pétroliers et de l'industrie chimique et/ou production et application des peintures, des teintures, des encres, des pigments, les cires chimiques et les produits de polissage,
- Applications de solvants (garages, pressings, blanchisseries, etc.,
- Décharges et anciennes décharges municipales, etc. (ISDD, ISDND, ISDD, etc.),
- Teintureries & Tanneries,
- Moquettes, tapis, tissus et plastiques avec des retardateurs de flammes,
- Production d'objets et meubles contenant des surfaces,
- Production de produits de nettoyage,
- Chimie photographique (laboratoires, et production des papiers et films, etc.),
- Production d'éléments électroniques,
- Production et applications de pesticides et biocides,
- Production de produits cosmétiques,
- Sites ayant reçus des Boues de STEP.

2. Chimie environnementale

La Chimie environnementale des PFAS est particulièrement importante et compliquée. Il n'existe aucun groupe de polluants montrant une chimie environnementale plus complexe que les PFAS. En particulier, il faut constater qu'il existe **plus que 9 000 substances PFAS**, divisées en **33 catégories de substances**. Les plus connues sont les Acides perfluoroalkane-sulfoniques (PFASs), les Acides perfluoroalkyliques-carboxyliques (PFCA), les Perfluoroalkyliques-phosphates & leurs esters, les Fluorotélomére-alcooles (**FTOH**), etc. (dont plus **32 autres groupes**...). Certains d'entre eux, comme par ex. le **PFOA** : Acide perfluoro-octanoïque et le **PFOS** : Perfluoro-octane-sulfonate (cf. Fig. 1) sont bannis (et **interdit en CE** et

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USA & Canada) par la **Convention de Stockholm** dans la catégorie des **POPs** : Persistent Organic Pollutants. Le PFOA est cancérogène. Les **produits commerciaux** contiennent principalement des **mélanges**.



Fig. 1 : Formules structurelles des PFOA & PFOS

La raison de la forte solubilité dans l'eau associée à une lipophilie est basée sur le fait qu'il existe des **PFAS** :

- **Anioniques** (par ex. les sulfonates, les sulfates, les carboxylates et les phosphates),
- **Cationiques** (par ex. ammonium quaternaire),
- **Amphotères** (par ex. betaines et sulfo-betaines) : base + acide et
- **Non ioniques** (par ex. polyéthylène glycols, oligomères d'acrylamide).

Il est très important de souligner, que les PFAS poly-fluorés non entièrement fluorés (» Précurseurs «) peuvent être convertis par bio-transformation en produits chimiques persistants et entièrement fluorés, les PFAS per-fluorés [87 - 94]. La dégradation complète microbiologique des PFAS n'a pas encore été démontrée.

Le schéma suivant montre un exemple de biotransformation des Alkylphosphates polyfluorés (PAP) dans les sols et eaux souterraines vers les Fluorotélomère alcohols (FTOH) volatils qui migrent par la suite dans les gaz du sol et vers l'air ambiant. Par la suite, les FTOH sont transformés microbiologiquement en PFAS per-fluorés stables. Par exemple ; le 6:2-FTOH est bio-transformé en PFHxA et PFPeA et le 8:2-FTOH en PFOA, PFHpA, PFHxA et 2H-PFOA (cf. les Fig. suivantes).

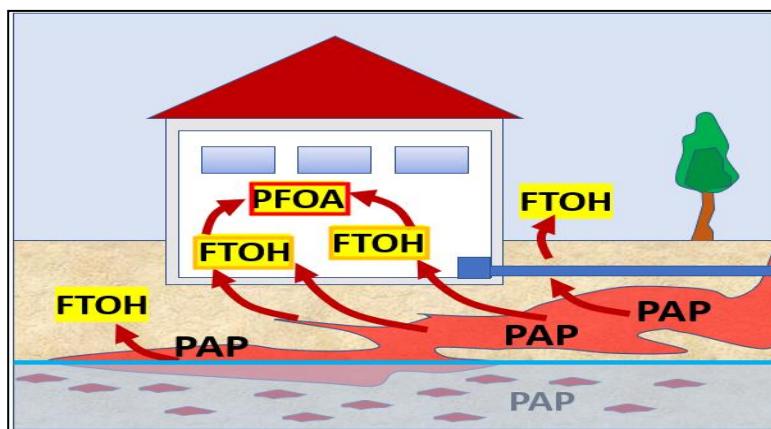


Fig. 2 : Exemple de biotransformation des Alkylphosphates polyfluorés (PAP) dans les sols et eaux souterraines vers les Fluorotélomère alcohols (FTOH) volatils et en PFAS per-fluorés stables, comme par ex. le PFOA cancérogène L. KOPF / HPC, 2017 et F. KARG, 2021 & 2022.

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Le Schéma suivant montre un exemple de la biotransformation du 8:2-FTOH ($\text{F}(\text{CF}_2)_8\text{CH}_2\text{CH}_2\text{OH}$) via des produits intermédiaires vers des PFAS perfluorés stables, comme par exemple le PFPA (Acide perfluoro-pantanonic), le PFHxA (Acide perfluoro-hexanonic), le PFHpA (Acide perfluoro-heptanonic), le 2H-PFOA, l'Acide 7:3 et le PFOA (Acide perfluoro-octanonic) cancérogène.

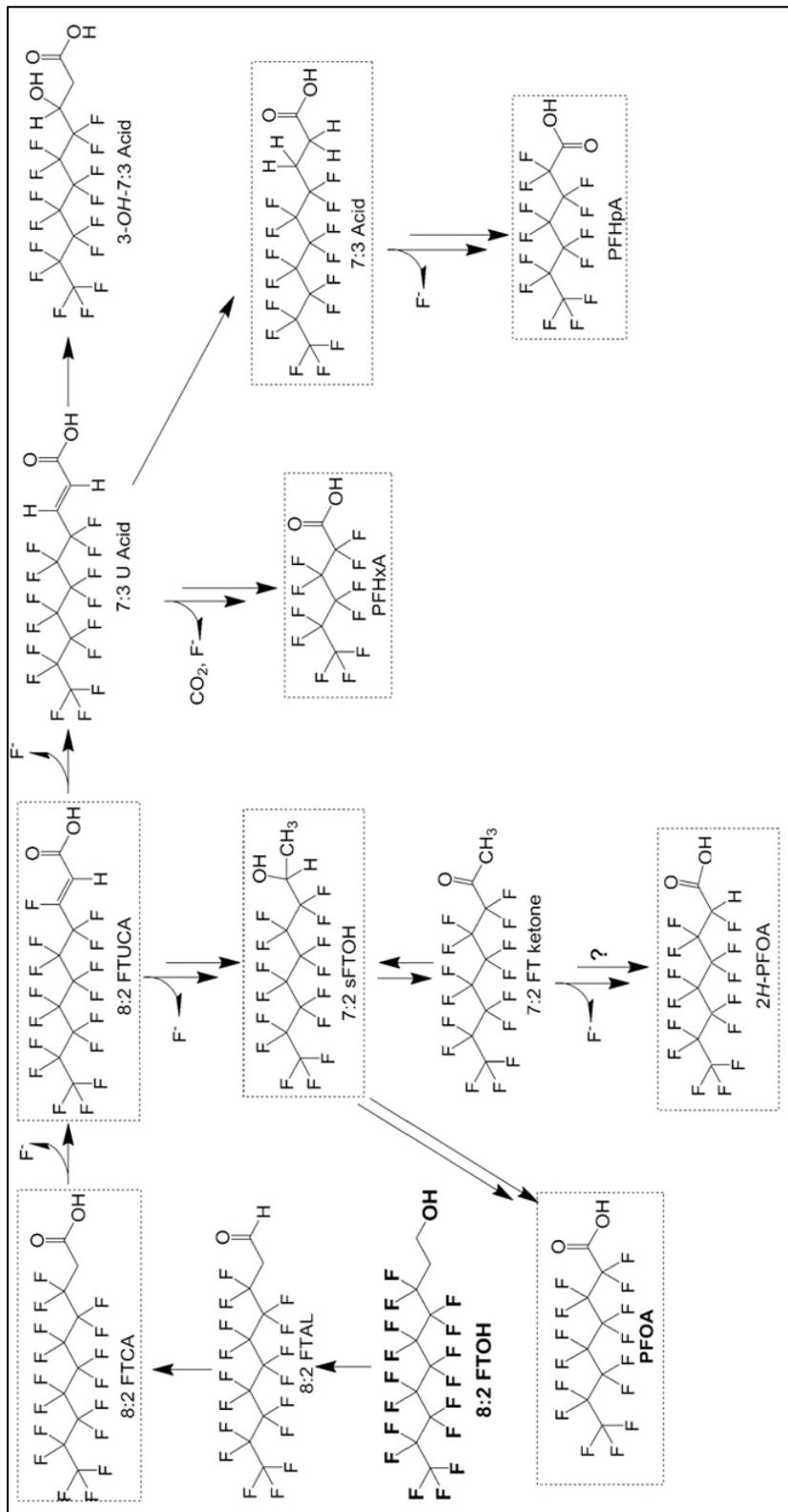


Fig. 3 : Exemple de biotransformation aérobie du 8:2 FTOH ($\text{F}(\text{CF}_2)_8\text{CH}_2\text{CH}_2\text{OH}$) dans les sols. Des flèches doubles indiquent la formation des substances perfluorées stables (Wang et al. 2009, modifié).

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En cas de changement de pH, certains PFAS pourraient devenir plus ou moins solubles, ce qui a également un impact sur les émanations de télomères fluorés volatils comme le FTOH, etc. dans les gaz du sol. Certains précurseurs pourraient modifier leurs solubilités (et leurs extractabilités lors des procédures d'analyse chimique). Par ex. l'intrusion d'eau de mer dans l'aquifère pourrait entraîner une augmentation du pH basique et donc de la solubilité de Capstone B. Cela a été observé en 2022 dans la zone portuaire de Hambourg / Allemagne suite aux inondations par l'eau de mer et l'intrusion dans les eaux souterraines, cf. la Fig. suivante. Ces effets pourraient entraîner des concentrations plus de 10 fois plus élevées dans les eaux souterraines qu'avant l'intrusion de l'eau de mer dans les sols et les eaux souterraines.

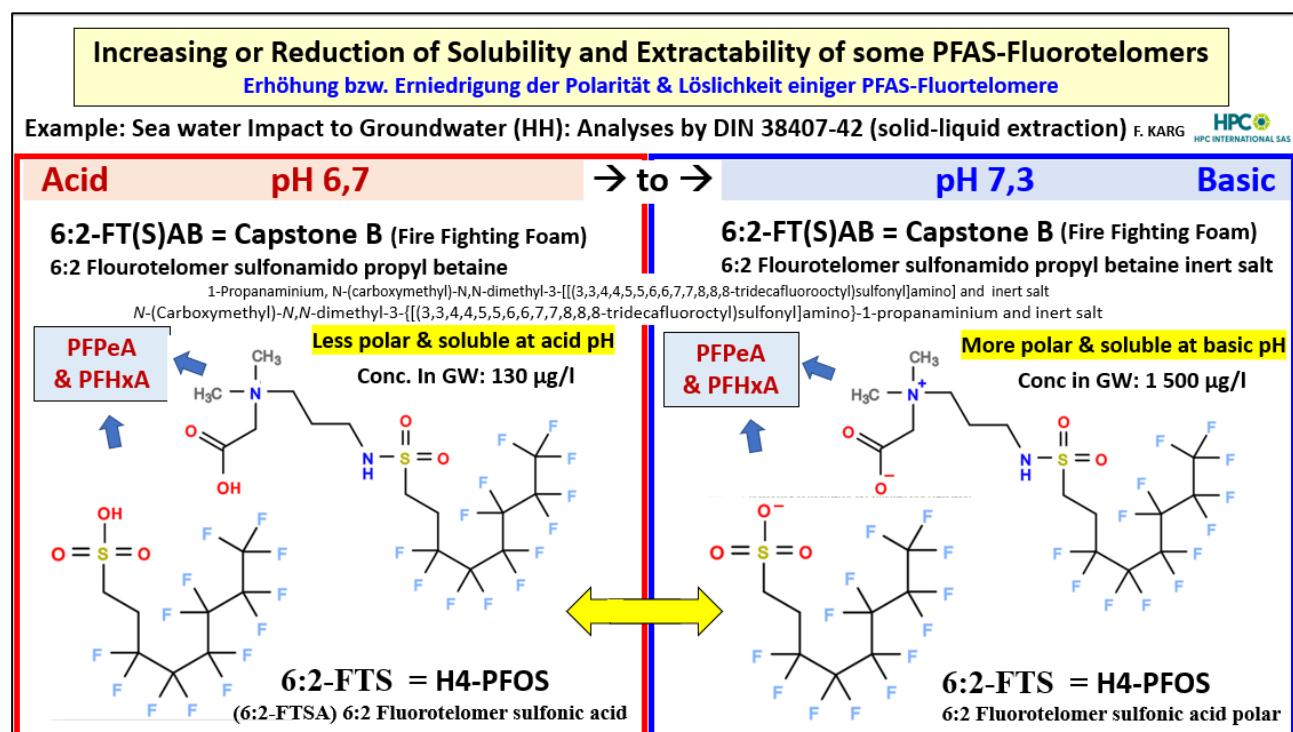


Fig. 4 : PFAS solubilités selon le pH (exemple : avant et après l'intrusion de l'eau de mer dans un aquifère

3. Identification des sources de pollution par les PFAS

Les pollutions des sols, eaux souterraines et eaux superficielles par les PFAS sont fréquemment des mélanges à l'origine de plusieurs sources de produits commerciaux et de sources de pollution. Il est possible de réaliser des screening complets des substances PFAS individuelles afin d'identifier entre 9 000 à 12 000 molécules, mais dans la gestion au quotidien des pollutions dans l'environnement ce n'est pas applicable du fait de la limite de faisabilité technico-économique.

Pour cette raison il est nécessaire de réduire le nombre de composés PFAS à analyser lors des diagnostics environnementaux grâce à la signature chimique des produits commerciaux suspectés d'être à l'origine de la contamination. En effet, les PFAS étant une

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famille de plus de 9 000 composés, il serait impossible de les quantifier tous.

Néanmoins il est nécessaire de rechercher l'identité du ou des produits commerciaux susceptibles d'être à l'origine d'une contamination du sous-sol, sur la base des teneurs en PFAS mesurées dans les sols ou les eaux (souterraines, de surface ou d'assainissement).

La solution de l'identification des produits commerciaux suspectés d'être à l'origine des pollutions par des PFAS, via une signature chimique, a été déjà développée par HPC avec des partenaires US-Américains et Suisses. Il s'agit du Fingerprinting par analyses des « Clusters PFAS » (Karg et al. & Monti 2022 & 2023) |85 – 86 & 95 - 97]. Il découle de ce travail qu'il est suffisant d'analyser 8 à 30 PFAS individuels afin d'identifier, via des relations entre les PFAS individuels, les produits et sources des pollutions d'origine.

Il a ainsi pu être identifié (en utilisant des résultats des analyses chimiques de plusieurs laboratoires d'analyses), parmi presque 60 000 Analyses de PFAS dans les eaux souterraines et superficielles en Italie du Nord-Est, 24 sources de pollution à l'origine d'un très grand panache de pollution par des PFAS. En comparaison de relations de seulement 8 PFAS entre eux, les produits commerciaux et activités à l'origine ont pu être identifiés, grâce aux banques de données (USA & Europe) concernant les Clusters PFAS. Ces sources sont par ex. les Imprégnations des textiles, les mousses anti-incendie, les galvanisations, les productions des platines électroniques, les traitements de surface des papiers, cartons, bois, etc.

Les origines des pollutions par des PFAS pourront être identifiées par les analyses des Clusters PFAS (Fingerprinting) (Fig. 1). Nous avons développé cette application avec des partenaires en Suisse et aux USA. Il s'agit principalement de la distribution de 8-30 PFAS perfluorés stables, afin d'identifier les produits d'origine (mousses anti-incendie, produits d'imprégnation des textiles, laques, Polymères fluorés (Teflon, PTFE, PVDF, etc.), mousses des procédés galvaniques, etc. Nous pouvons intégrer jusqu'à 500 PFAS individuels dans cette analyse de Clusters, mais la limitation à 20 – 30 PFAS individuels est moins couteuse et dans la plupart des cas suffisante. Cette approche a permis de dissocier dans le Nord-Est de l'Italie plusieurs origines différentes de grands panaches de pollutions par des PFAS mélangés dans les eaux souterraines et eaux superficielles (Fig. 5).

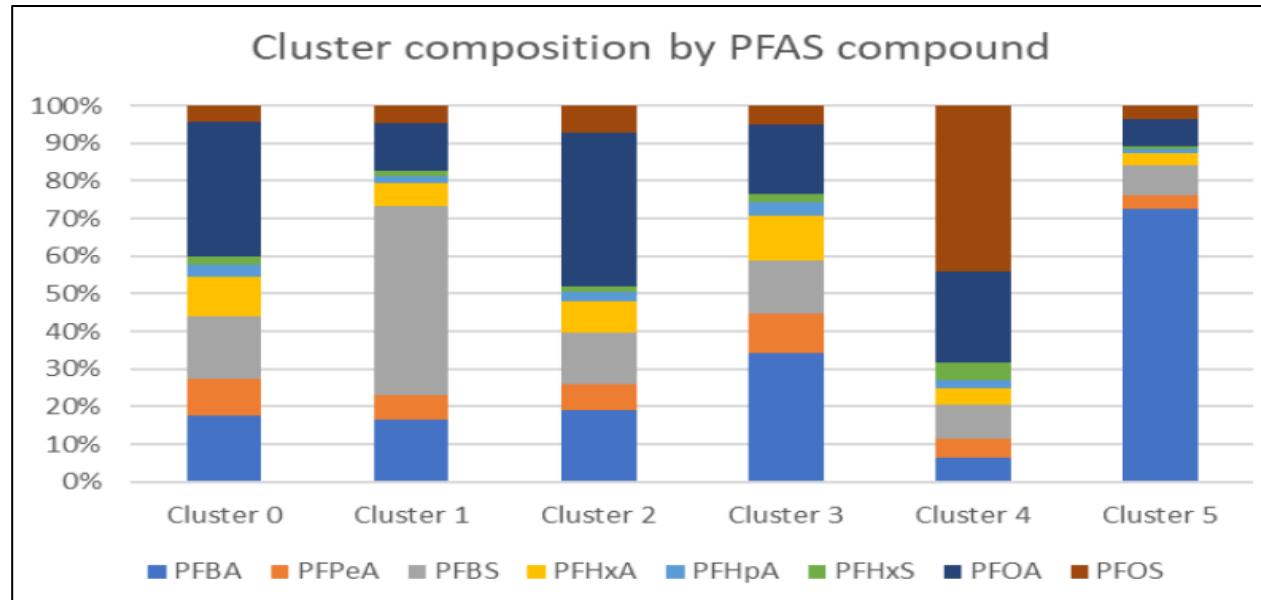


Fig. 5 : Clusters PFAS selon les produits d'origine.

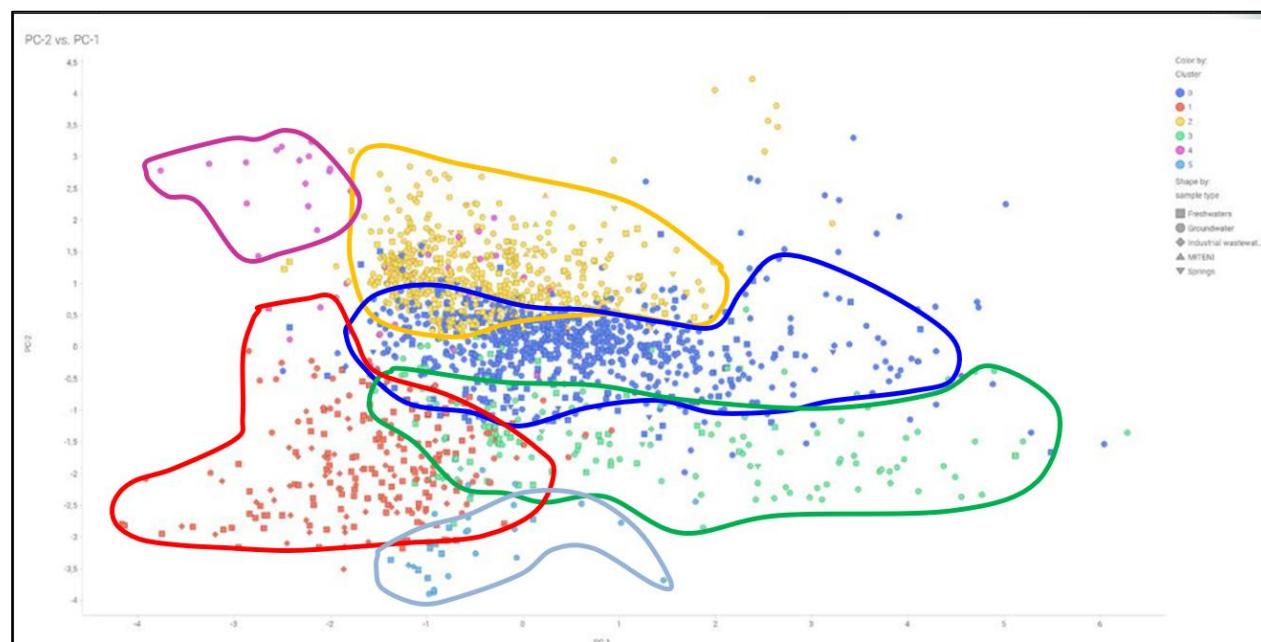


Fig. 6 : Identification de 6 produits d'origines via des analyses des Clusters PFAS (zone de 761 ha et 472 analyses des eaux souterraines et eaux superficielles dans le Nord-Est de l'Italie) [85 - 86].

La Recherche de l'identité du ou des produits commerciaux susceptibles d'être à l'origine d'une contamination du sous-sol, sur la base de teneurs en PFAS mesurées par un nombre de PFAS réduits (Liste PFAS signature) dans les sols ou les eaux (souterraines, de surface ou d'assainissement), se basera sur des enregistrements préalables de spectres de PFAS individuels (Liste des PFAS étendue), en comparaison des pourcentages de PFAS individuels par rapport du total des PFAS (TOF : Total Organofluorine), cf. Fig. 7.

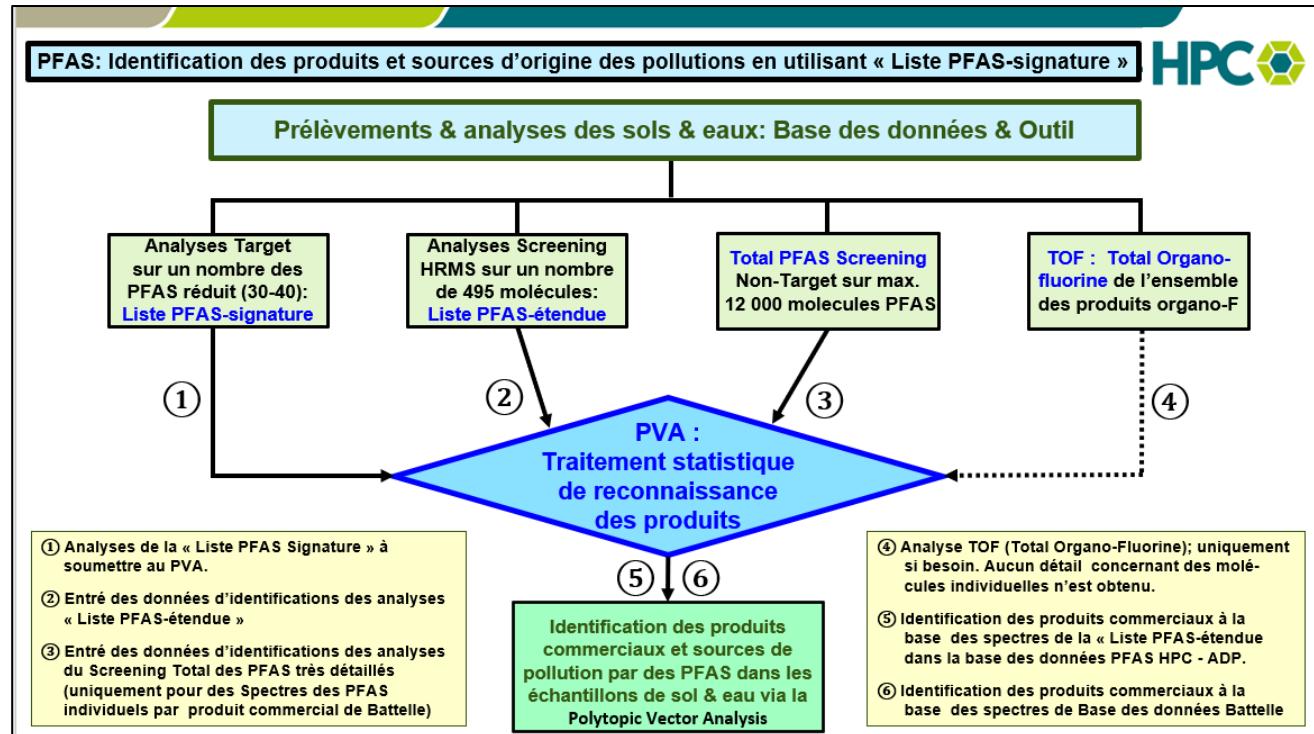


Fig. 7 : Principes de développement & fonctionnement de la Base des données HPC des PFAS

L'identification des produits commerciaux contenant les PFAS nécessite aussi des études historiques d'applications des produits commerciaux des PFAS (comme également prévu dans l'Arrêté Ministériel de 2023 concernant la surveillance des émissions des PFAS, par ex. dans les eaux usées des ICPE :

- Des recherches bibliographiques ;
- Des visites sur les sites (potentiellement) pollués concernés ;
- Des entretiens avec les différents témoins.

Les données suivantes seront notamment recherchées :

- La nature des produits utilisés ;
- La période d'utilisation des produits ;
- Les lieux d'utilisation / de stockage des produits ;
- La fréquence d'utilisation des produits ;
- Les quantités annuelles consommées ;
- Les éventuels incidents recensés (déversement accidentel, incendie...) ;
- L'éventuelle réalisation de travaux de dépollution ou d'aménagement sur les sites potentiellement exposés.

Afin de définir des spectres des PFAS individuels des produits commerciaux et sources des pollutions des PFAS, les analyses chimiques de caractérisation sont les suivantes :

- « Analyses Non-target » quantitatives sur environ 500 molécules PFAS individuelles par Liquid Chromatography High Resolution Mass Spectrometry (LC-HR-MS ou HRMS) pour identifier une liste "Liste PFAS-signature" à partir une "Liste PFAS-étendue"

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(jusqu'à 500 PFAS individuels).

- Constitution d'une signature chimique suffisamment représentative des produits commerciaux analysés par des analyses non ciblées. Les analyses de PFAS totaux (ci-après dénommées "Total PFAS") peuvent être réalisées soit par un « Total Organofluoré (TOF : Total Organofluorine) qui n'apporte aucune information sur des molécules individuelles de PFAS, soit par un screening total afin d'identifier jusqu'à 12 000 molécules PFAS (ces analyses présentent un coût très élevé et ne sont donc pas envisagées systématiquement).

La construction de la base de données des spectres individuels de produits commerciaux de PFAS est réalisée comme suit (cf. Fig. 8) :

- Une « Base de données HPC » de spectres des molécules de produits commerciaux de PFAS à partir de Microsoft Access (ou Excel).
- Pour des Produits commerciaux échantillonnés sur les sites ou obtenus auprès des fournisseurs des produits commerciaux des PFAS et analysés en complément, environ 500 molécules PFAS individuelles ("Liste PFAS-étendue") seront enregistrées dans la Base de Données PFAS d'HPC.
- Pour les spectres des analyses de Produits PFAS commerciaux dégradés (vieillissements des Poly-fluorés par biotransformations) dans les lysimètres dotés des bactéries de biotransformation), leurs spectres sur au maximum 500 molécules PFAS individuelles ("Liste PFAS-étendue") seront également enregistrés dans la Base de Données PFAS d'HPC.
- **Les Chromatogrammes des Produits commerciaux et des Produits soumis au vieillissement analysés et recensés sont transformés numériquement et stockés dans la « Base de données HPC » sur Microsoft Access.**

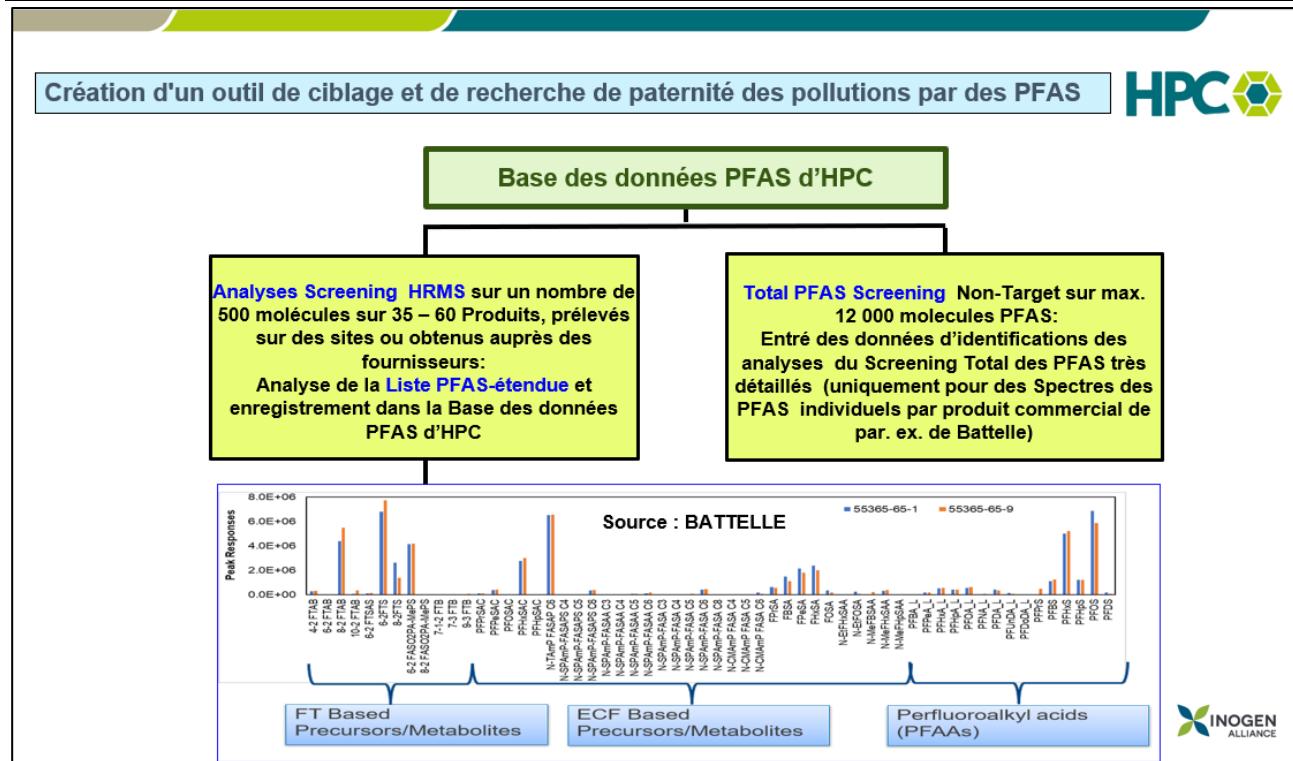


Fig. 8 : Base de données HPC des spectres moléculaires des produits commerciaux des PFAS d'HPC

- La « Base de données PFAS d'HPC » sur Microsoft Access est supérieure à celle d'autres Databanks, car elle intègre, en plus des spectres connus (Chromatogrammes des PFAS), ceux de Produits Commerciaux soumis au vieillissement par Lysimètres et Biotransformation bactériologique des PFAS polyfluorés).
- Une liste de Queries (Requêtes possibles ou Paramètres d'identification, comme par ex. la reconnaissance des produits commerciaux à partir des relations des PFAS individuels de la « Liste PFAS-signature ») sera créée pour identifier les produits commerciaux à partir de la liste standard d'analyses des PFAS réduits (30 – 40 substances individuels = « Liste PFAS-signature »).

Les principes de l'identification des produits commerciaux et sources de pollutions des PFAS à l'origine sont démontrés dans les Fig. 9 – 11.

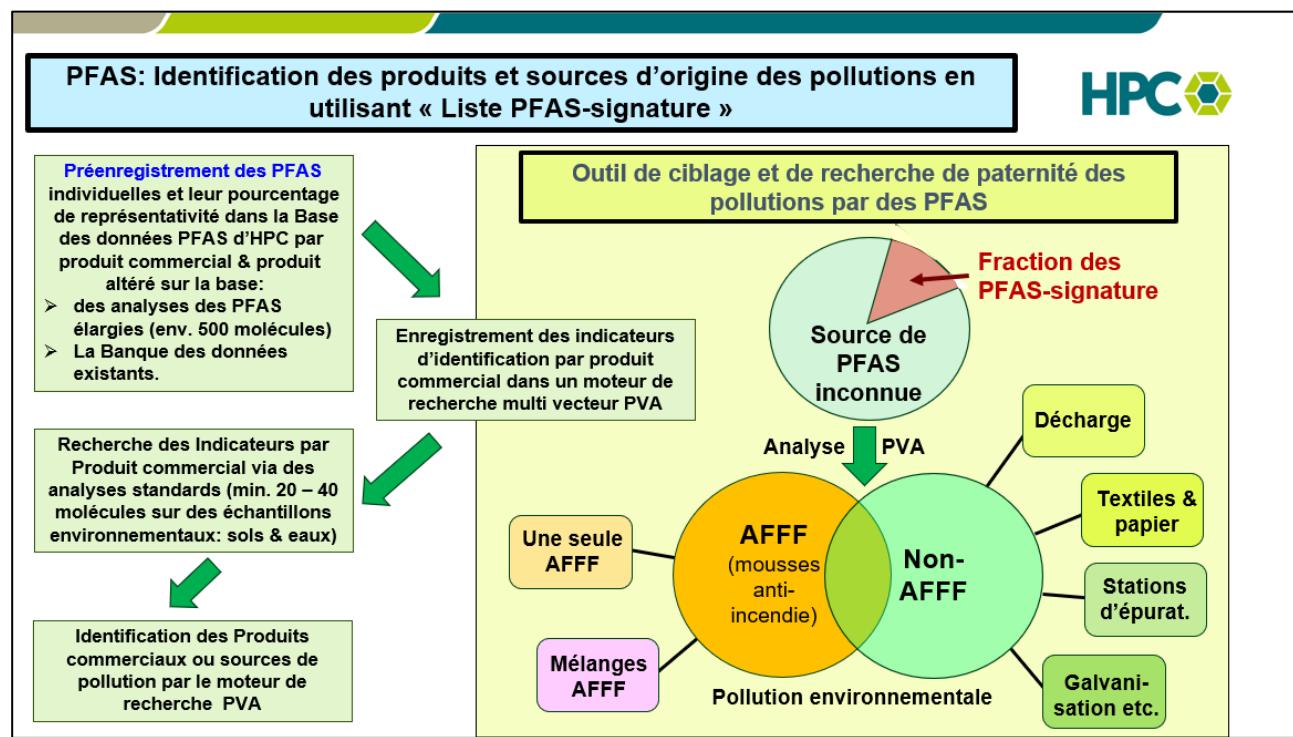


Fig. 9: Identification des produits PFAS initiaux dans les mélanges des produits commerciaux à l'origine des pollutions (1/3)

Le programme d'identification concernant l'origine des produits commerciaux :

Le programme d'identification permet de faire la recherche concernant l'origine des produits commerciaux sur la base d'un nombre de PFAS individuels réduit dans les analyses des échantillons environnementaux (sols et eaux) à partir de la « Base des données PFAS d'HPC » sur Microsoft Access (ou Excel). Ceci inclue :

- Les résultats de recherche de distribution des concentrations attendues seront exprimés en % de la concentration totale attendue dans les sols ou dans les eaux souterraines, de surface ou d'assainissement (Fonction A).
- Recherche de l'identité du ou des produits commerciaux susceptibles d'être à l'origine d'une contamination du sous-sol.
- Le programme d'identification indique (sous la forme d'un tableau) la probabilité d'origine des pollutions dans les sols ou dans les eaux souterraines, de surface ou d'assainissement par rapport aux Produits commerciaux recensés dans la « Base de données PFAS d'HPC ».
- Dans le cadre des identifications probabilistes des origines d'une contamination aux PFAS, les Spectres (distributions chromatographiques des PFAS) des Produits commerciaux altérés et enregistrés dans la « Base de Données PFAS d'HPC » sont inclus.
- Le programme d'identification sur la « Base de données PFAS d'HPC » sera supérieur à ceux existants, car il utilise en plus des spectres existants aussi des spectres (Chromatogrammes des PFAS) des Produits Commerciaux soumis au vieillissement par Lysimètres et de la Biotransformation bactériologique des PFAS polyfluorés, etc.).

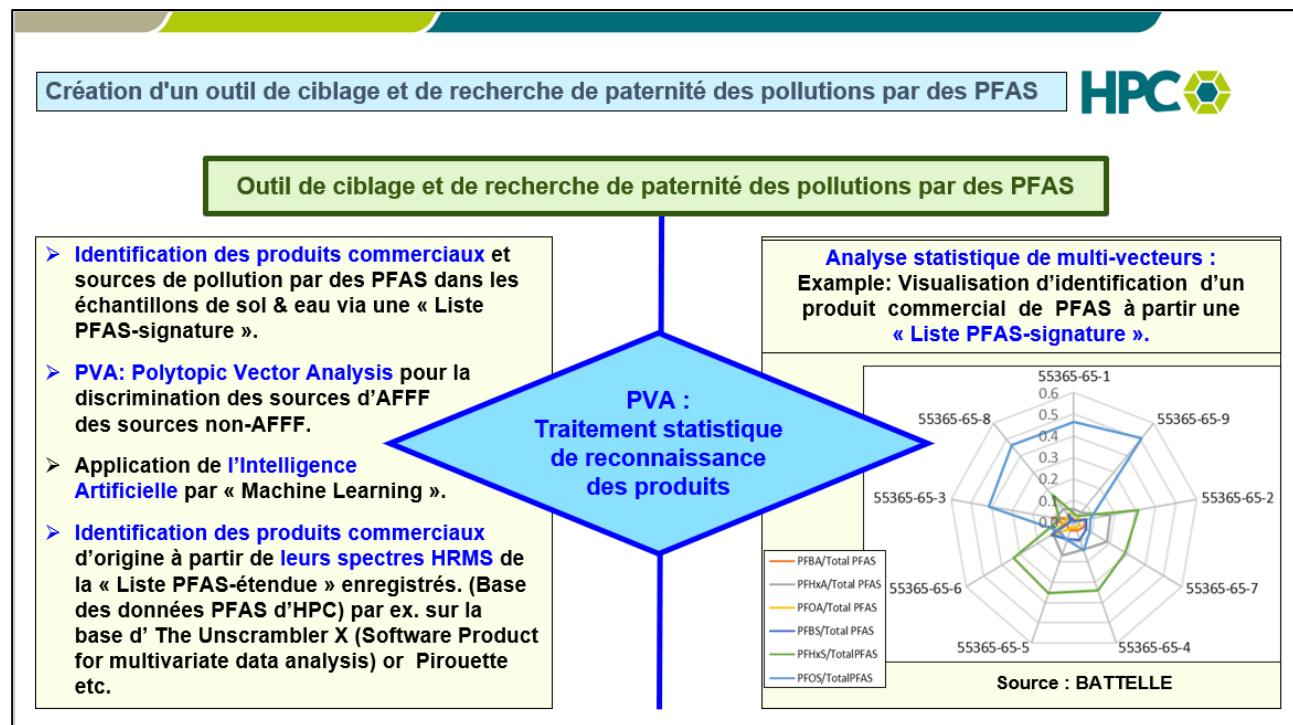


Fig. 10: Identification des produits PFAS initiaux dans les mélanges des produits commerciaux à l'origine des pollutions (2/3)

- Le programme d'identification des produits commerciaux sera un modèle de classification statistique comme celui qui a été développé à l'origine par Batelle pour la détection des AFFF (et autres produits) dans des mélanges de différentes sources de PFAS. Ce modèle, sera aussi basé sur l'Intelligence Artificielle et, en particulier, sur l'approche du Machine Learning.
- L'application de l'Outil sera basé sur la présence et la quantification dans les échantillons analysés de composés PFAS restreints extrêmement stables dans les conditions environnementales, et qui caractérisent l'empreinte de l'AFFF ou d'autres sources de PFAS en supplément des programmes d'identification existants sur des PFAS perfluorés stables formés par la biotransformation de l'altération des produits commerciaux.
- Les résultats de l'application de l'Outil sont les probabilités que le mélange de PFAS trouvé dans un échantillon soit lié à certains produits commerciaux (ou sources de pollution).
- En utilisant la méthode statistique appelée PVA (Polytopic Vector Analysis = Analyse Vectorielle Polytopique), il est possible d'attribuer un pourcentage d'appartenance statistique (ou ressemblance) des échantillons environnementaux à chaque Produit commercial (ou source de Pollution).
- Les logiciels utilisés pour le développement de l'Outil sont R, Mathematica®, The Unscrambler® et Pirouette®.

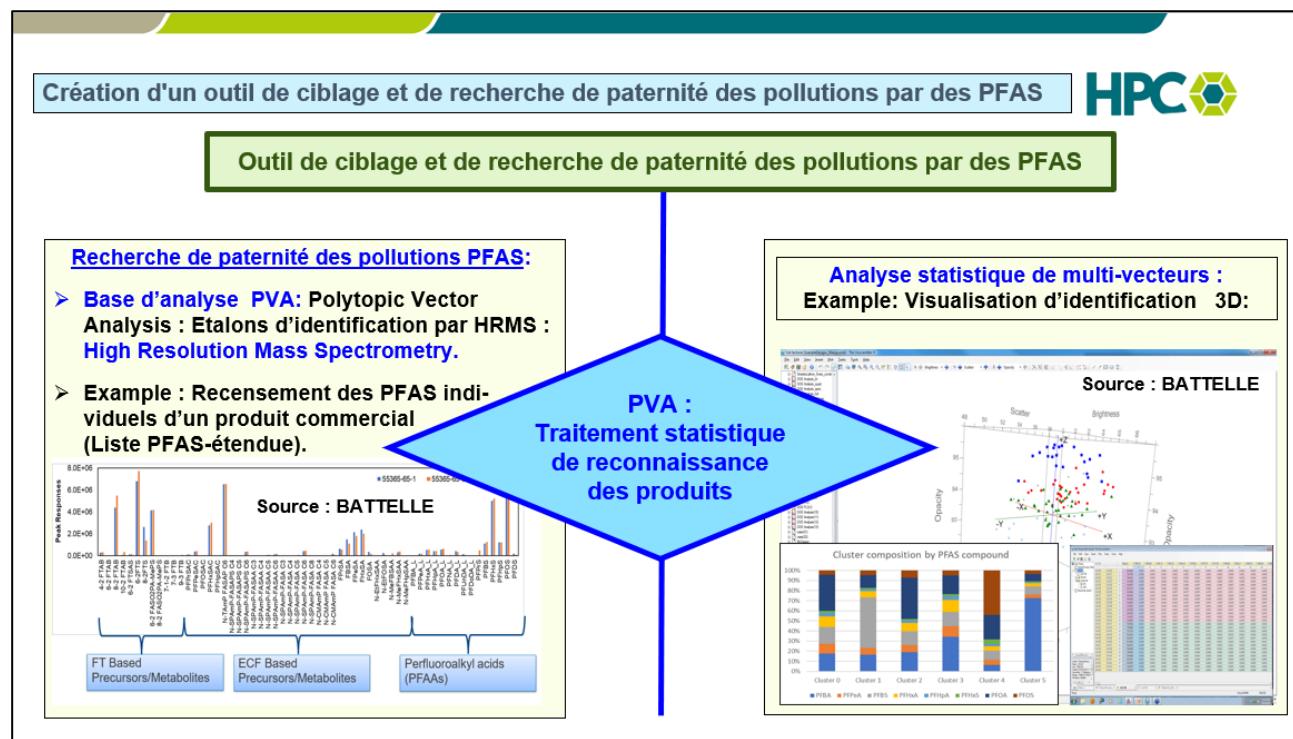


Fig. 11: Identification des produits PFAS initiaux dans les mélanges des produits commerciaux à l'origine des pollutions (3/3)

La difficulté existante sont les spectres (Chromatogrammes) des Produits commerciaux dégradés par biotransformation des poly-fluorés vers des per-fluorés, car ces spectres en PFAS individuels sont différents des produits « frais » et aussi de ceux intégrés dans la Base des données HPC à partir des Databanks existants.

Pour surmonter cette difficulté de l'identification des produits commerciaux altérés dans les échantillons environnementaux (sols & eaux, analysés pour un nombre réduit de PFAS), la « **Base des données PFAS-HPC** » contiendra des Chromatogrammes élargis sur 500 PFAS individuels, représentant des spectres PFAS de Produits commerciaux soumis au tests de vieillissement dans les Lysimètres Ces tests seront réalisés via un consortium bactériologique réputé de la bio-transformation en question (du type Acidimicrobium sp. Strain A6 (Huang & Jaffé 2019) et/ou Pseudomonas strain SYC et/ou Rhodococcus jostii RHA1 et/ou Pseudomonas oleovorans (Ying Shi, 2018). C'est ainsi que des étalons représentatifs de mélanges commerciaux de PFAS seront obtenus par la simulation d'un vieillissement de plusieurs années.

De manière générale préalable, il est recommandé d'intégrer dans les diagnostics des contaminations par les PFAS au minimum les polluants suivants :

N°	Polluant PFAS de base	Synonyme	VTRs existantes
1	Acide perfluoro-butanoïque	PFBA	Oui
2	Acide perfluoro-pentanoïque	PPPeA	Oui
3	Acide perfluoro-hexanoïque	PFHxA	Oui
4	Acide perfluoro-heptanoïque	PFHpA	Oui
5	Acide perfluoro-octanoïque	PFOA	Oui
6	Acide perfluoro-nonanoïque	PFNA	Oui
7	Acide perfluoro-decanoïque	PFDA	Oui
8	Acide perfluoro-butane-sulfonique	PFBS	Oui
9	Acide perfluoro-hexane-sulfonique	PFHxS	Oui
10	Acide perfluoro-heptane-sulfonique	PFHpS	Oui
11	Perfluoro-octane-sulfonate	PFOS	Oui
12	Acide H4-polyfluoro-octane-sulfonique	H4-PFOSA	associé au PFOS
N°	Polluant PFAS de base	Synonyme	VTRs existantes
13	Perfluoro-octane-sulfonamide	PFOSA	Oui
14	6:2 Fluorotélomère alcool	6:2-FTOH	Oui
15	8:2 Fluorotélomère alcool	8:2-FTOH	Oui

Si possible:

16	Perfuorobutane sulfonate	PFBS	Oui
17	Perfuoropentane sulfonate	PPPeS	Oui
18	Perfuorohexane sulfonate	PFHxS	Oui
19	Perfuoroheptane sulfonate	PFHpS	Oui
20	Perfuorodecane sulfonate	PFDS	Oui
21	Acide perfluoro-undecanoïque	PFUnDA	Oui
22	Acide perfluoro-dodecanoïque	PFDoDA	Oui
23	Acide perfluoro-tridecanoïque	PFTrDA	Oui
24	Acide perfluoro-tetradecanoïque	PFTeDA	Oui
25	Acide perfluoro-hexadecanoïque	PFHxDA	Oui
26	Acide perfluoro-octadecanoïque	PFODA	Oui
27	Acide Hexafluoro-propylèneoxydimer	HFPO-DA	Oui
28	Acide 3H-perfluoro-3-[(3-methoxy-propoxy) propanoïc	ADONA	Oui

Fig. 12 : Paramètres d'analyses PFAS en minimum recommandés

A l'état d'aujourd'hui, l'analyse des 20 PFAS individuels de la Directive Européenne 2020/2184, pourra être réalisée mais également élargie à 40 PFAS individuels. :

Selon la Directive Européenne 2020/2184, qui concerne la qualité des eaux de consommation humaine, les 20 PFAS suivants sont ciblés :

Somme des 20 PFAS de la Directive européenne Eau potable 2020 :

- Acide perfluorooctanoïque (PFOA) [5347]
- Acide perfluoroheptanoïque (PFHpA) [5977]
- Acide perfluorohexanoïque (PFHxA) [5978]
- Acide perfluoropentanoïque (PFPeA)[5979]
- Acide perfluorobutanoïque (PFBA) [5980]
- Acide perfluorobutane sulfonique (PFBS) [6025]
- Acide perfluorododécanoïque (PFDoDA) [6507]
- Acide perfluorononanoïque (PFNA) [6508]
- Acide perfluorodécanoïque (PFDA) [6509]
- Acide perfluoroundécanoïque (PFUnDA) [6510]
- Acide perfluoroheptane sulfonique (PFHpS) [6542]
- Acide perfluorotridecanoïque (PFTrDA) [6549]
- Acide perfluorodécane sulfonique (PFDS) [6550]
- Acide perfluorooctane sulfonique (PFOS) [6561]
- Acide perfluorohexane sulfonique (PFHxS) [6830]
- Acide perfluoropentane sulfonique (PFPeS) [8738]
- Acide perfluorononane sulfonique (PFNS) [8739]
- Acide perfluoroundécane sulfonique [8740]
- Acide perfluorododécanoïque [8741]
- Acide perfluorotridécane sulfonique [8742]

4. Evaluation des Risques Toxicologiques

Pour l'évaluation des risques sanitaires, les données toxicologiques (VTR : Valeurs Toxicologiques de Référence) sont à rechercher et à actualiser au niveau international quasiment de façon hebdomadaire. Les VTR les plus récentes sont disponibles surtout aux USA (EPA...), ATSDR et l'EFSA. L'ANSES avait publié aussi en 2017 des VTR des PFAS, mais au vu de l'avancement forcé des études toxicologiques, ces VTR sont pour une grande partie déjà dépassées.

Dans le cas des FTOH présents dans les sols, eaux souterraines et gaz du sol, il faudra impérativement réaliser aussi des investigations de FTOH dans l'Air ambiant des bâtiments aux usages sensibles (ERP : Ecoles, Crèches, etc.) ou résidentiel, sur la base de seuils de quantification (ou au moins pour les seuils de détection), de l'ordre de 4 – 8 ng/m³, afin de posséder une bonne base exploitable pour les EQRS : Evaluation Quantitative des Risques Sanitaires.

Une première approche d'évaluation des risques simplifiée est possible via des valeurs limites existantes, comme par ex. en Allemagne, ou publiées par la Communauté Européenne. En Allemagne il existe des Valeurs limites pour l'eau potable, pour les sols et pour les eaux souterraines.

L'US-EPA considère que la plus grande partie des expositions provient aujourd'hui de l'eau potable, en dehors des sites pollués, du fait de l'absence de surveillance et de traitement des PFAS.

Selon la Directive (UE) 2013/39/UE « Cadre sur l'eau » européenne (DCE), concernant le PFOS & dérivés (et d'autres substances prioritaires) une Norme de Qualité Environnementale (NQE-MA) de 0,65 ng/l pour les Eaux superficielles et de 0,13 ng/l pour les Eaux du milieu marin (et des NQE-CMA : Concentrations Maximales Admissibles) a été fixée.

Il est important d'annoncer, qu'une simple application des Valeurs limites, génériques et individuelles dans le cadre d'une évaluation simplifiée des risques ne prend pas en compte les scénarios d'expositions spécifiques et les expositions aux mélanges (« Cocktails ») des polluants avec, au minimum, une prise en compte de l'additivité des risques des polluants ayant les mêmes cibles et effets toxicologiques. Par conséquence, il est préférable de réaliser des EQRS (ou ARR, HRA, TERQ), ce qui correspond bien à la Méthodologie Française de Gestion des Sites Pollués, selon la note du Ministère en charge de l'Environnement, du 19/04/2017.

Un autre aspect important est que seule une EQRS (ou ARR, TERQ, HRA) permettra de définir des Valeurs de Contrôle de conformité sanitaire, sous forme de CMA (Concentration Maximale Admissible) pour des Excès des Risques Individuels (de cancer) maximalement acceptables : ERI < 10^{E-5} ou d'un Indice de Risque systémique de IR < 1 (= DJE / DJT : Dose Journalière d'Exposition sur la Dose Journalière Tolérable).

Concernant les EQRS : Evaluation Quantitative des Risques Sanitaires, la base est soit la mesure des concentrations dans les milieux d'exposition, soit la modélisation du transfert des polluants d'un compartiment à l'autre (par ex. des polluants dans les eaux souterraines ou du sol vers les gaz du sol et de l'air ambiant).

Une étape importante de l'EQRS est le choix des VTR (Valeurs Toxicologiques de Référence), car leur évolution est rapide. Par ex., le guide PFAS de l'ANSES de 2017 comporte certaines VTR pour les PFAS, mais vu les nombreuses publications de VTR jusqu'à aujourd'hui, ces valeurs sont partiellement dépassées, et notamment beaucoup plus contraignantes à ce jour. Une Dose Hebdomadaire Tolérable (DHT) de 4,4 ng/kg/Semaine (ou la Dose Journalier Tolérable (DJT) de 0,63 ng/kg/j pour des PFAS : PFOA, PFOS, PFNA & PFHxS) a été publiée par l'EFSA, le 17/09/2020. En 2020 des facteurs d'équivalence de toxicité par rapport au PFOA ont été aussi publiés par W. Bil et al. sous la forme de RPF : Relative Potency Factors.

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Afin d'assurer le bon choix des VTR des PFAS, il est recommandé d'appliquer des critères de choix scientifiques et non des critères nationaux. La Fig. suivante montre des critères de choix des VTR applicables, afin de prendre en compte les meilleures connaissances toxicologiques concernant les relations de dose à effet des PFAS.

No	TRD: Toxicological Reference Dose Choice Criteria	Appreciation			
		Favorable	Correct	Not favorable	Exclusion
1	Variability of indicated TRD	(+/- 0 %)	≤ (+/- 30 %)	> (+/- 30 %)	
2	Class (potential) Carcinogenic: EC: Class 3/ US-EPA: Class B2, C / IARC: Group 1	3 Organisms : CE, US-EPA, IARC, etc.	2 Organisms	1 Organisms	
3	Several Organisms shows similar TRD (+/- 50 %)	> 3 Organisms	2 Organisms	1 Organism	
4	Age of base Study	≤ 15 a	15 – 25 a	< 25 a	
5	Mechanistic toxicological basement Study (for ex. Genotoxicity):	Epidemiology	Mamifer	In-Vitro / In-silico	
6	Basement Study : Klimisch Quality Criteria	Class 1	Class 2	Class 3	Class 3
7	Verified Purity of Compound	Yes	< 95 %	No	
8	Excipient potentially toxic	Non		Yes	
9	Presence of population without exposure (test witness)	Yes		No	
10	General Quality Criteria (Klimisch) of toxicological effect studies	Standardized Study (OCDE, UE, US EPA, FDA, etc.)	Standardized Study without Details, but correctly documented	Document insufficient for evaluation, systematic deficiencies	
11	POD : Point of Departure	Quantified Epidemiological Data, BMLD, etc. (PBPK)	NOAEL sensitive NOAEL	LOAEL sensitive, LOAEL, Other	
12	Uncertainty (or Assessment) Factors	1 – 100	> 100 – 1000	> 1 000 – 10 000	> 10 000
13a	Transpositions: Between Exposure Pathways	Non		Yes	
13b	Transposition: Animal to Human	Non	Yes		
13c	Transpositions : From in-Vitro	Non		Yes	
13d	Transpositions : From in-Silico	Non		Yes	
14	Study time-representativity	≥ chronic (> 180 d)	sub-chronic (90 d) to chronic (180 d)	< sub-chronic (< 90 d)	
15	Integration of bio-disponibility / Bio-resorption capacity (ex.: DIN 19738)	Yes	Not known (100 %)	Known, but not considered	

Fig. 12 : Critères de choix des VTR (F. KARG 2022)

En se basant sur ces critères de choix, il est possible de définir par ex. les VTR indiquées en Fig. suivante (datant du début de l'année 2023). Il est important de prendre en compte, que les Administrations de certains états US-américains sont très avancées dans la recherche toxicologique et la publication des VTR concernant des PFAS, du fait de leur forte présence d'industries chimiques et pétrolières (avec les fortes utilisations historiques des AFFF) ou de la forte présence de grandes décharges industrielles (et les pollutions environnementales associées). Il s'agit notamment des états du Texas, du Michigan et de New Jersey. Ces VTR sont souvent les bases pour les publications des VTR des PFAS de la US-EPA fédérale.

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Substance	Cancerogen / not cancerogen	Chronic toxicological value			Species	Sigle	Security Factor	Organization
		Exposure path	Target organ	Value				
PFBA	NC	oral	Hepatic	2,9 µg/kg/d	Rate	RfD	NOAEL / 2400	TCEQ 2016
		inhalation	Hepatic	10 µg/m³	Rate	RfC	from oral value	TCEQ 2016
PPPeA	NC	oral	Hepatic	3,8 µg/kg/d	Rate	RfD	same than PFHxS LOAEL/(263*300)	TCEQ 2016
PFHxA	NC	oral	Hepatic	3,8 µg/kg/d	Rate	RfD	same than PFHxS LOAEL/(263*300)	TCEQ 2016
PFHpA	NC	oral	Hepatic	25 ng/kg/d	Rate	DJT	Extrapolation of DJT of Health Canada	ANSES 2017
PFOA	NC	oral	Hematologic	0,86 ng/kg/d	Rate	TDI	BMDL5	UBA 2020 BfR & EFSA 2018
			Hepatic, Mammar, Hematologic	12 ng/kg/d	Mice	RfD	LOAEL (81*100)	TECQ 2016
		inhalation	Hepatic	4,1 ng/m³	Rate	RfC	NOAEL / (81*3000)	TCEQ 2016
	C	oral	Testicular tumors	2,52 (mg/kg/d)⁻¹	Epidemio	SF	-	New Jersey 2017
PFNA	NC	oral	Hematologic	2,5 ng/kg/d	Mouse	RfD	NOAEL / 300	EPA IRIS 2019 New Hampshire DES 2019
		inhalation	Lung, respiratory system	28 ng/m³	Rate	RfC	NOAEL / (81*30 000)	EPA IRIS 2019 TCEQ 2018
PFDA	NC	oral	Hepatic	15 ng/kg/d	Rate	RfD	NOAEL / (81*1000)	TCEQ 2016
				53 ng/m³	Rate	RfC	from oral value	TCEQ 2016
PFBS	NC	oral	Hematologic and renal	1,4 µg/kg/d	Rate	RfD	NOAEL / (142*300)	TCEQ 2016
				4,9 µg/m³	Rate	RfC	from oral value	TCEQ 2016
PFHxS	NC	oral	Hematologic and thyroïdal	3,8 µg/kg/d	Rate	RfD	LOAEL / (263*300)	TCEQ 2016
		inhalation		13 ng/m³	Rate	RfC	from oral value	TCEQ 2016
PFHpS	NC	oral	Hepatic	0,43 ng/kg/d	Rate	TDI	Potency Factor : 0,6-2	UBA 2020, EFSA 2018, BfR 2018
PFOS	NC	oral	Hepatic	1,86 ng/kg/d	Monkey	TDI	NOAEL	UBA 2020 BfR & EFSA 2018
		inhalation	Thyroidal, neurological and foetal development	81 ng/m³	Rate	RfC	from oral value (23 ng/kg/j)	TCEQ 2016
PFOSA	NC	oral	Mammary glands	12 ng/kg/d	Mice	RfD	Same than PFOA NOAEL/(81*300)	TCEQ 2016
		inhalation		4,1 ng/m³	Rate	RfC	same than PFOA NOAEL/(81*3000)	TCEQ 2016

Fig. 13 : Choix de certains VTR selon les critères en Fig. 7 (F. KARG 2022) :

- ANSES: Agence nationale de sécurité sanitaire de l'alimentation, de l'environnement et du travail (2017)
ATSDR: Agency for Toxic Substances and Disease Registry
EFSA: European Food and Safety Authority
IRIS : Integrated Risk Information of Substances (U.S. - EPA)
UBA : Umweltbundesamt (Germany)
BfR: Bundesinstitut für Risikobewertung (Germany)
OEHHA : Office of Environmental Health Hazard Assessment
WHO: World Health Organization
RIVM : Netherlands Environmental & Health Institute
MDHHS: Michigan Department of Health and Human Services, Division of Environmental Health
TCEQ: Texas Commission on Environmental Quality
NJ-DWQIHES: New Jersey Drinking Water Quality Institute Health Effects Subcommittee
Bil et al. 2020 : Toxicological Equivalence Factors on PFOA RfD

Une étape complémentaire de l'EQRS (ou ARR, TERQ, HRA) est la définition des Valeurs de contrôle de conformité sanitaire, sous forme des CMA (Concentrations Maximales Admissibles) en intégrant une additivité des risques des polluants concernant les mêmes cibles et

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effets toxicologiques, pour des Excès des Risques Individuels (de cancer) maximalement acceptables : ERI < 10E-5 ou d'un Indice de Risque systémique de IR < 1 (= DJE / DJT : Dose Journalière d'Exposition sur Dose Journalière Tolérable). Les CMA sont couramment utilisées sous forme de Valeurs de contrôle sanitaire, afin de vérifier ou de co-élaborer les objectifs des mesures correctives, voire des objectifs de dépollution. Les mesures de Gestion, comme par ex. la dépollution sont basés en France sur un Plan de Gestion, une définition des Zones Sources des pollutions concentrées puis un Bilan Coût-Avantages des modalités et technologies différentes de gestion et des traitements.

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PFAS: Environmental Chemistry Investigations, Contamination Source Identification, Toxicology and TERQ Risk Assessments, including FTOH

PFAS : Chimie Environnementale, Diagnostics & Identification des Sources de Pollution, Toxicologie et Evaluation des Risques (EQRS), incluent les FTOH

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1. Introduction

PFAS: Per- & Polyfluoro-Alkyl Substances have gradually become since the 1960s a major environmental problem, also for Public Health, due to their multiple and vast application (historical and still current). This environmental and public health threat has started to be taken into account little by little since the 2010s and strongly in 2022 & 2023. Consequently, PFAS are now found in soils, groundwater, food and water, drinking water as well as in soil gases and ambient air for a family of volatile PFAS, the FTOHs: Fluorotelomer-Alcohols. Between 9,000 and 12,000 synthetic PFAS pollutants have been produced.

PFAS polymers of the "Teflon" type (or PFTE) etc. do not have good bioavailability and are therefore much less toxic than monomeric PFAS. These PFAS monomers are the subject of the work presented herewith. PFAS are known in particular for their toxicological effects of endocrine disruptors, hepatotoxicity, immunotoxicity, their effects on fetal development and for some, carcinogenicity (e.g. PFOA) [1 - 84].

An important characteristic of PFAS is their behavior in Environmental Chemistry, because only polyfluorinated PFAS are modified by microbiological bio-transformation into perfluorinated PFAS, which remain totally stable and non-degradable, or even bioaccumulable.

The sources of pollution by PFAS are multiple and particularly present on industrial sites, which have used these products, sites of former fires or firefighting training, where firefighting foams (AFFF: Anti Fire Fighting Foams, eg at airports) were used. Agricultural land is also a source of PFAS pollution, due to the input of sludge from STEP: Wastewater Treatment Plants which contain accumulated PFAS.

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The following (historical) activities can cause PFAS pollution: Entraînements anti-incendie,

- Airport or air base military site,
- Fire site and use of AFFF,
- Electrochemical galvanizing,
- Production of “waxed” paper or cardboard,
- Production of Waterproof Textiles,
- Sprays, paints, waterproofing lacquers,
- Production and application of Teflons (PTFE, etc.),
- Petroleum and chemical industry sites and/or production and application of paints, dyes, inks, pigments, chemical waxes and polishing products,
- Solvent applications (garages, dry cleaners, laundries, etc.,
- Landfills and former municipal landfills, etc. (ISDD, ISDND, ISDD, etc.),
- Dyeings & Tanneries,
- Carpets, rugs, fabrics and plastics with flame retardants,
- Production of objects and furniture containing surfaces,
- Production of cleaning products,
- Photographic chemistry (laboratories, and production of papers and films, etc.),
- Production of electronic elements,
- Production and applications of pesticides and biocides,
- Production of cosmetic products,
- Sites having received Sludge from STEP.

2. Environmental Chemistry

The environmental chemistry of PFAS is particularly important and complicated. There is no group of pollutants showing more complex environmental chemistry than PFAS. In particular, it should be noted that there are more than 9,000 PFAS substances, divided into 33 substance categories. The best known are Perfluoroalkane-sulfonic acids (PFASs), Perfluoroalkyl-carboxylic acids (PFCA), Perfluoroalkyl-phosphates & their esters, Fluorotelomer-alcohols (FTOH), etc. (including more than 32 other groups...). Some of them, eg. PFOA: Perfluoro-octanoic acid and PFOS: Perfluoro-octane-sulfonate (see Fig. 1) are banned (and prohibited in the EC and USA & Canada) by the Stockholm Convention in the category of POPs: Persistent Organic Pollutants. PFOA is carcinogenic. Commercial products mainly contain mixtures.

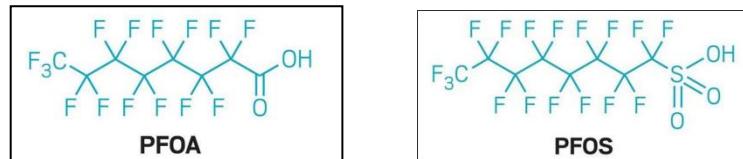


Fig. 1: Structural formulas of PFOA and PFOS

The reason for the high water solubility associated with lipophilia is based on the fact that there are PFAS:

- Anionics (e.g. sulfonates, sulfates, carboxylates and phosphates),
- Cationic (e.g. quaternary ammonium),
- Amphoteres (e.g. betaines and sulfo-betaines): base + acid and
- Non-ionic (eg polyethylene glycols, acrylamide oligomers).

It is very important to emphasize, that not fully fluorinated poly-fluorinated PFAS (“Precursors”) can be converted by bio-transformation into persistent and fully fluorinated chemicals, the per-fluorinated PFAS [87 – 94]. Complete microbiological degradation of PFAS has not yet been demonstrated.

The following diagram shows an example of the biotransformation of polyfluorinated alkyl phosphates (PAP) in soils and groundwater to volatile fluorotelomer alcohols (FTOH) which subsequently migrate into soil gases and into the ambient air. Subsequently, the FTOHs are transformed microbiologically into stable per-fluorinated PFAS. For example ; 6:2-FTOH is biotransformed into PFHxA and PFPeA and 8:2-FTOH into PFOA, PFHpA, PFHxA and 2H-PFOA (see the following Figs).

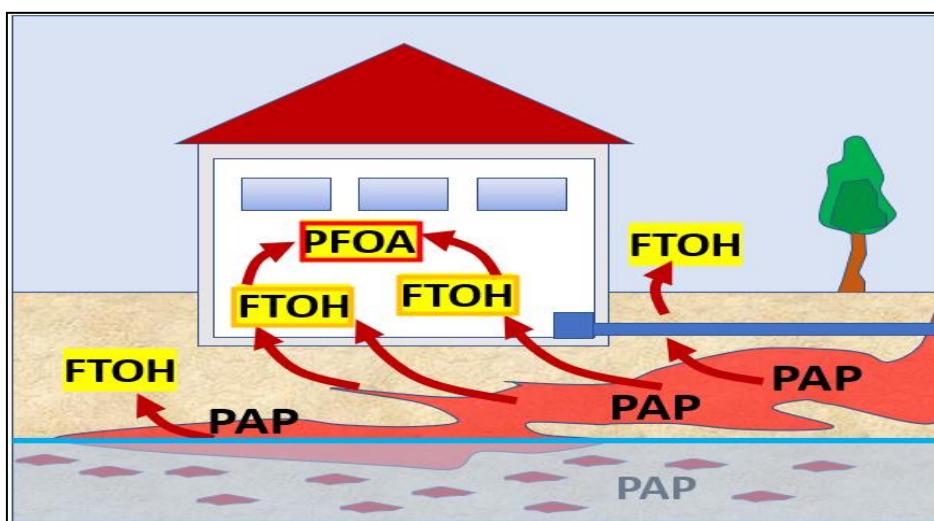


Fig. 2: Example of biotransformation of polyfluorinated alkyl phosphates (PAP) in soils and groundwater to volatile fluorotelomer alcohols (FTOH) and stable per-fluorinated PFAS, such as e.g. carcinogenic PFOA L. KOPF / HPC, 2017 and F. KARG, 2021 & 2022.

The following diagram shows an example of the biotransformation of 8:2-FTOH ($\text{F}(\text{CF}_2)_8\text{CH}_2\text{CH}_2\text{OH}$) via intermediate products to stable perfluorinated PFAS, such as PFPA (Perfluoro-pentanonic acid), PFHxA (Perfluoro-pentanonic acid), hexanonic), PFHpA (Perfluoro-heptanonic acid), 2H-PFOA, Acid 7:3 and carcinogenic PFOA (Perfluoro-octanonic acid).

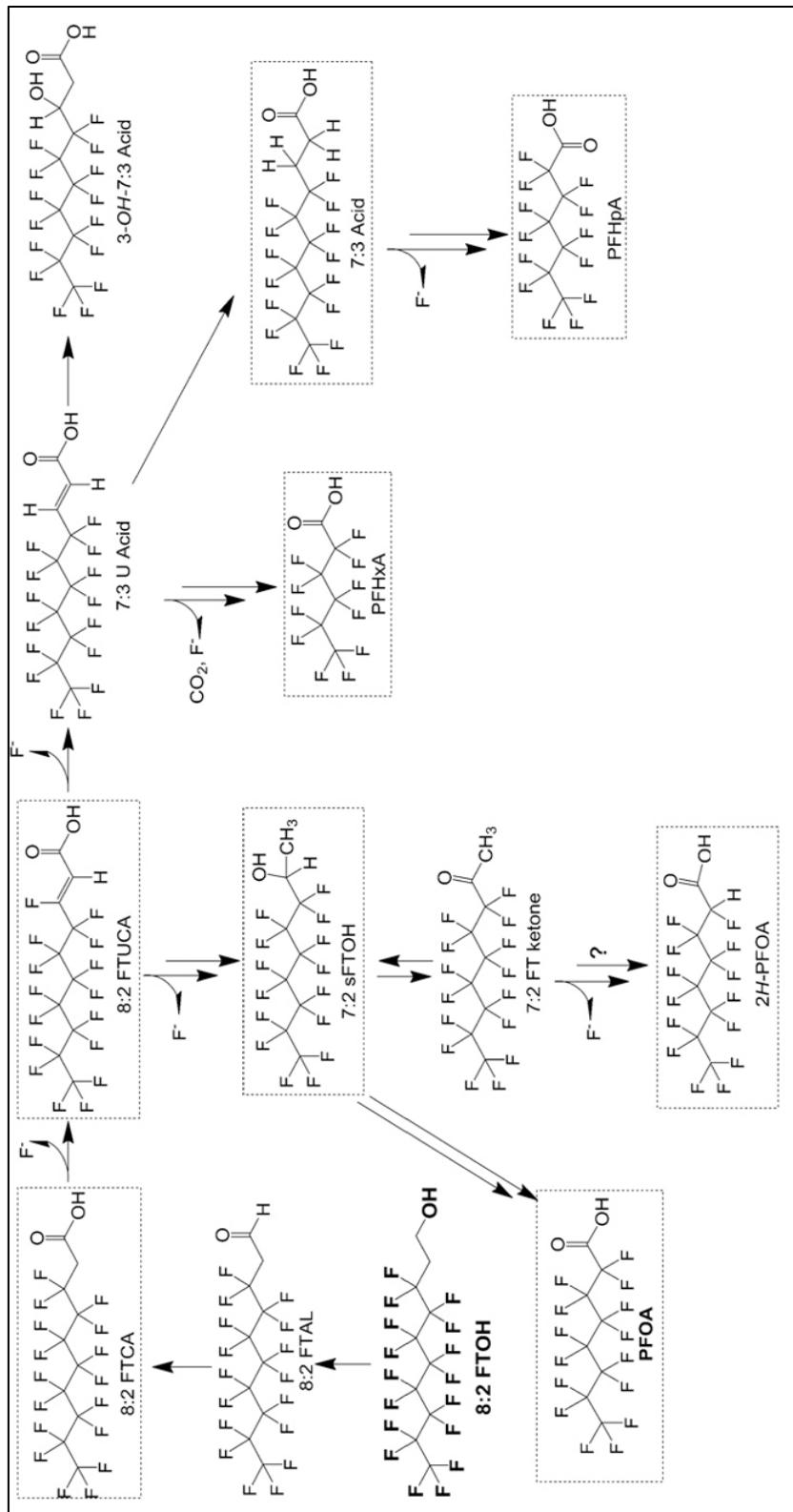


Fig. 3: Example of aerobic biotransformation of 8:2 FTOH ($\text{F}(\text{CF}_2)_8\text{CH}_2\text{CH}_2\text{OH}$) in soils. Double arrows indicate the formation of stable per-fluorinated substances (Wang et al. 2009, modified).

In the event of a change in pH, some PFAS could become more or less soluble, which also has an impact on the emanations of volatile fluorinated telomeres like FTOH, etc. in soil gases. Some precursors could modify their solubilities (and their extractabilities during chemical analysis procedures). Eg. the intrusion of seawater into the aquifer could lead to an increase in the basic pH and therefore the solubility of Capstone B. This was observed in 2022 in the port area of Hamburg / Germany following flooding by water from sea and groundwater intrusion, cf. Fig. next. These effects could result in concentrations more than 10 times higher in groundwater than before seawater intrusion into soils and groundwater.

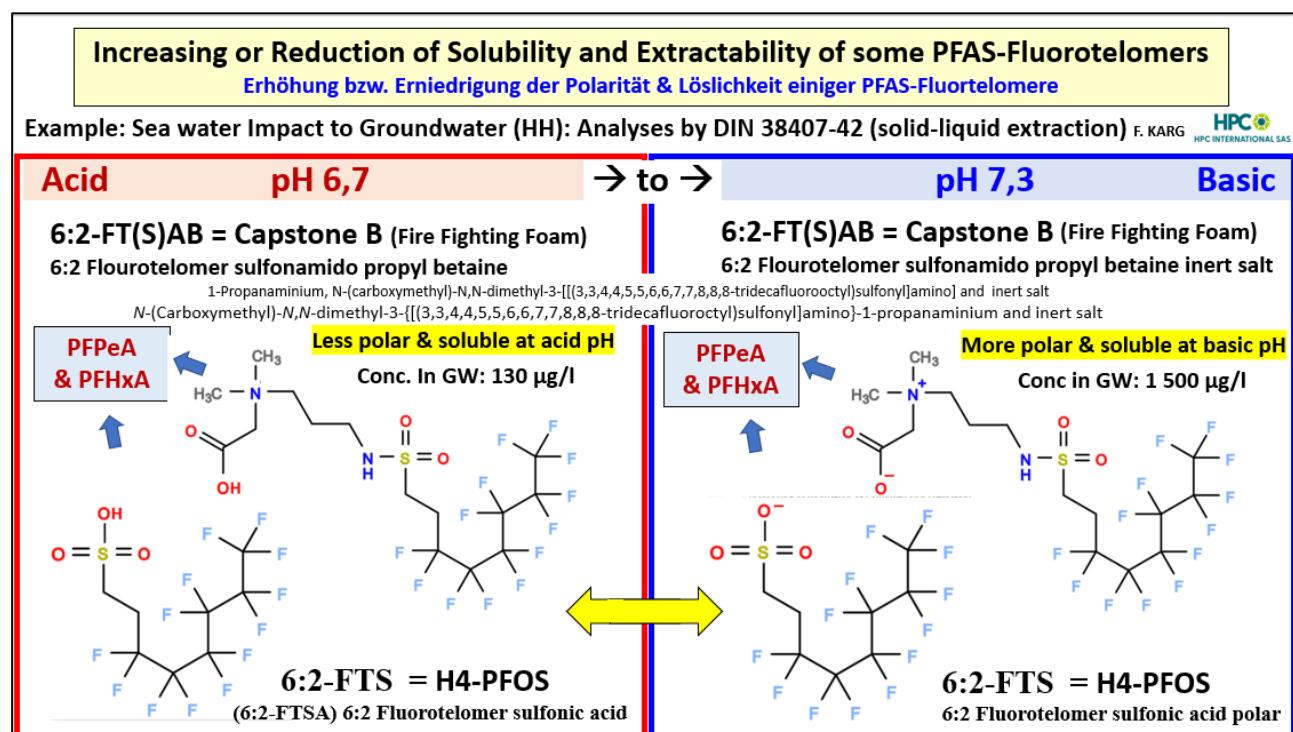


Fig. 4: PFAS solubilities according to pH (example: before and after seawater intrusion into an aquifer)

3. Identification of PFAS Pollution Sources

Pollution of soil, groundwater and surface water by PFAS are frequently mixtures at the origin of several sources of commercial products and sources of pollution. It is possible to carry out complete screening of individual PFAS substances in order to identify between 9,000 and 12,000 molecules, but in the daily management of pollution in the environment this is not applicable due to the technical-economic feasibility limit.

For this reason it is necessary to reduce the number of PFAS compounds to be analyzed during environmental diagnostics thanks to the chemical signature of the commercial products suspected of being the source of the contamination. Indeed, PFAS being a family of more than 9,000

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compounds, it would be impossible to quantify them all. Nevertheless, it is necessary to seek the identity of the commercial product(s) likely to be the source of contamination of the subsoil, on the basis of the PFAS contents measured in the soil or water (underground, surface or sanitation).

The solution for identifying commercial products suspected of being the source of PFAS pollution, via a chemical signature, has already been developed by HPC with US-American and Swiss partners. This is Fingerprinting by analysis of “PFAS Clusters” (Karg et al. & Monti 2022 & 2023) [85 – 86 & 95 - 97]. It follows from this work that it is sufficient to analyze 8 to 30 individual PFAS in order to identify, via the relationships between the individual PFAS, the products and sources of the original pollution.

It has thus been possible to identify (using the results of chemical analyzes from several analysis laboratories), among almost 60,000 Analyzes of PFAS in ground and surface waters in North-East Italy, 24 sources of pollution at the origin of a very large PFAS pollution plume. In comparison of the relationships of only 8 PFAS between them, the commercial products and activities at the origin could be identified, thanks to the databases (USA & Europe) concerning the PFAS Clusters. These sources are e.g. Textile impregnations, fire-fighting foams, galvanizing, production of electronic boards, surface treatments of paper, cardboard, wood, etc.

The origins of PFAS pollution can be identified by analyzes of PFAS Clusters (Fingerprinting) (Fig. 5). We have developed this application with partners in Switzerland and the USA. This mainly concerns the distribution of 8-30 stable perfluorinated PFAS, in order to identify the original products (fire-fighting foams, textile impregnation products, lacquers, fluorinated polymers (Teflon, PTFE, PVDF, etc.) .), foams from galvanic processes, etc. We can integrate up to 500 individual PFAS in this Cluster analysis, but the limitation to 20 – 30 individual PFAS is less expensive and in most cases sufficient. to dissociate in the North-East of Italy several different origins of large pollution plumes by PFAS mixed in groundwater and surface water (Fig. 6).

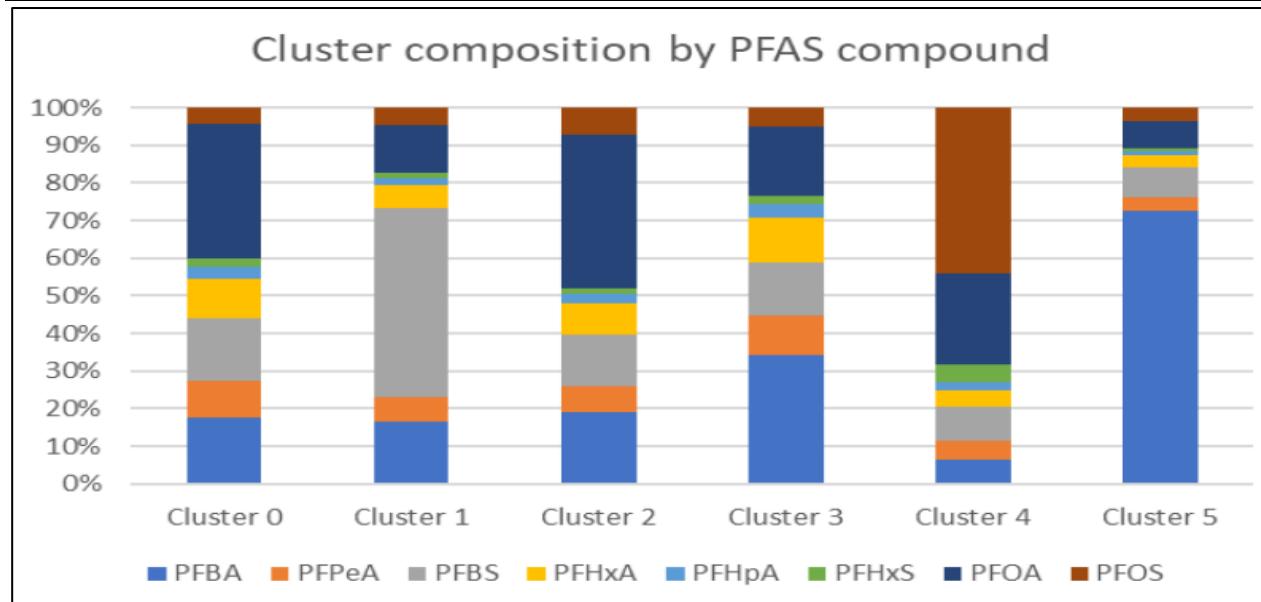


Fig. 5 : PFAS Clusters depending on commercial products

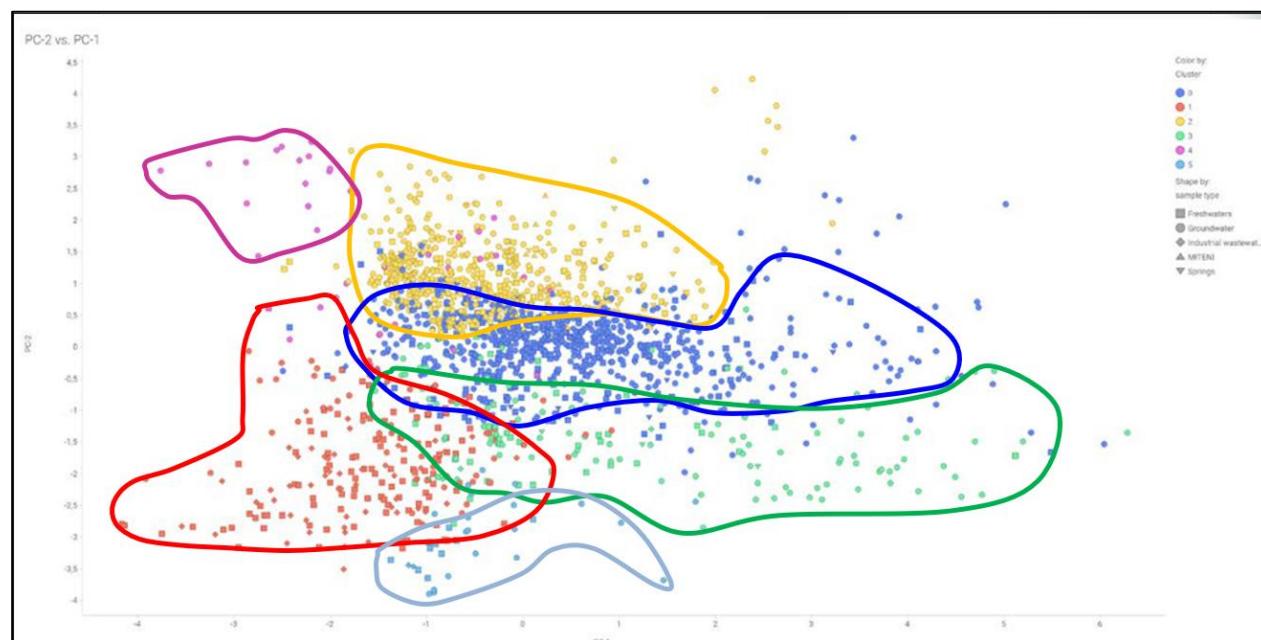


Fig. 6: Identification of 6 products of origin via analyzes of the PFAS Clusters (area of 761 ha and 472 analyzes of groundwater and surface water in the North-East of Italy 85 - 86].

Finding the identity of the commercial product(s) likely to be the source of subsoil contamination, based on PFAS levels measured by a reduced number of PFAS (signature PFAS list) in the soil or waters (ground, surface or sewage), will be based on prior recordings of individual PFAS spectra (Extended PFAS List), in comparison with the percentages of individual PFAS in relation to total PFAS (TOF: Total Organo Fluorite), cf. Fig. 7.

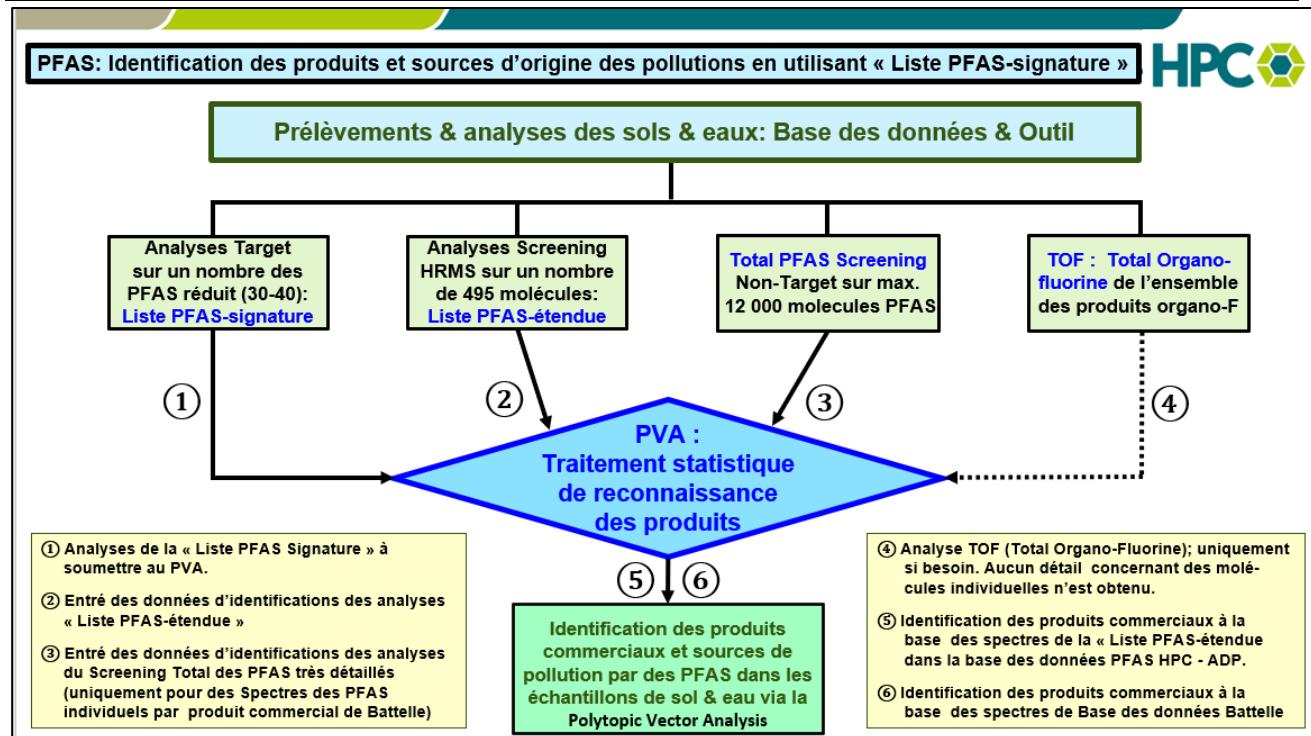


Fig. 7: Principles of development and operation of the PFAS HPC Database

The identification of commercial products containing PFAS also requires historical studies of applications of commercial PFAS products (as also provided for in the Ministerial Order of 2023 concerning the monitoring of PFAS emissions, e.g. in wastewater from ICPEs :

- Bibliographic research;
- Visits to the (potentially) polluted sites concerned;
- Interviews with various witnesses.

The following data will be sought in particular:

- The nature of the products used;
- The period of use of the products;
- The places of use / storage of the products;
- The frequency of use of the products;
- The annual quantities consumed;
- Any identified incidents (accidental spillage, fire, etc.);
- The possible carrying out of depollution or development work on potentially exposed sites.

In order to define individual PFAS spectra of commercial products and sources of PFAS pollution, the chemical characterization analyzes are as follows:

- Quantitative "Non-target analyses" on approximately 500 individual PFAS molecules by Liquid Chromatography High Resolution Mass Spectrometry (LC-HR-MS or HRMS) to identify a "PFAS-signature List" from a "PFAS-Extended List" (up to 500 individual PFAS).

- Constitution of a chemical signature sufficiently representative of the commercial products analyzed by non-targeted analyses. Total PFAS analyzes (hereinafter referred to as "Total PFAS") can be carried out either by a "Total Organo-Fluorine (TOF: Total Organofluorine) which does not provide any information on individual molecules of PFAS, or by a total screening in order to identify up to 12,000 PFAS molecules (these analyzes have a very high cost and are therefore not systematically considered).

The construction of the database of individual spectra of commercial PFAS products is carried out as follows (cf. Fig. 8):

- An "HPC Database" of spectra of molecules of commercial PFAS products from Microsoft Access (or Excel).
- For Commercial Products sampled at sites or obtained from suppliers of commercial PFAS products and further analyzed, approximately 500 individual PFAS molecules ("PFAS-Extended List") will be recorded in the HPC PFAS Database.
- For the spectra of the analyzes of degraded commercial PFAS Products (aging of Poly-fluorinated by biotransformations) in lysimeters equipped with biotransformation bacteria), their spectra on a maximum of 500 individual PFAS molecules ("Extended PFAS List") will also be recorded in the HPC PFAS Database.
- Chromatograms of analyzed and recorded Commercial Products and Products subject to aging are digitally transformed and stored in the "HPC Database" on Microsoft Access.

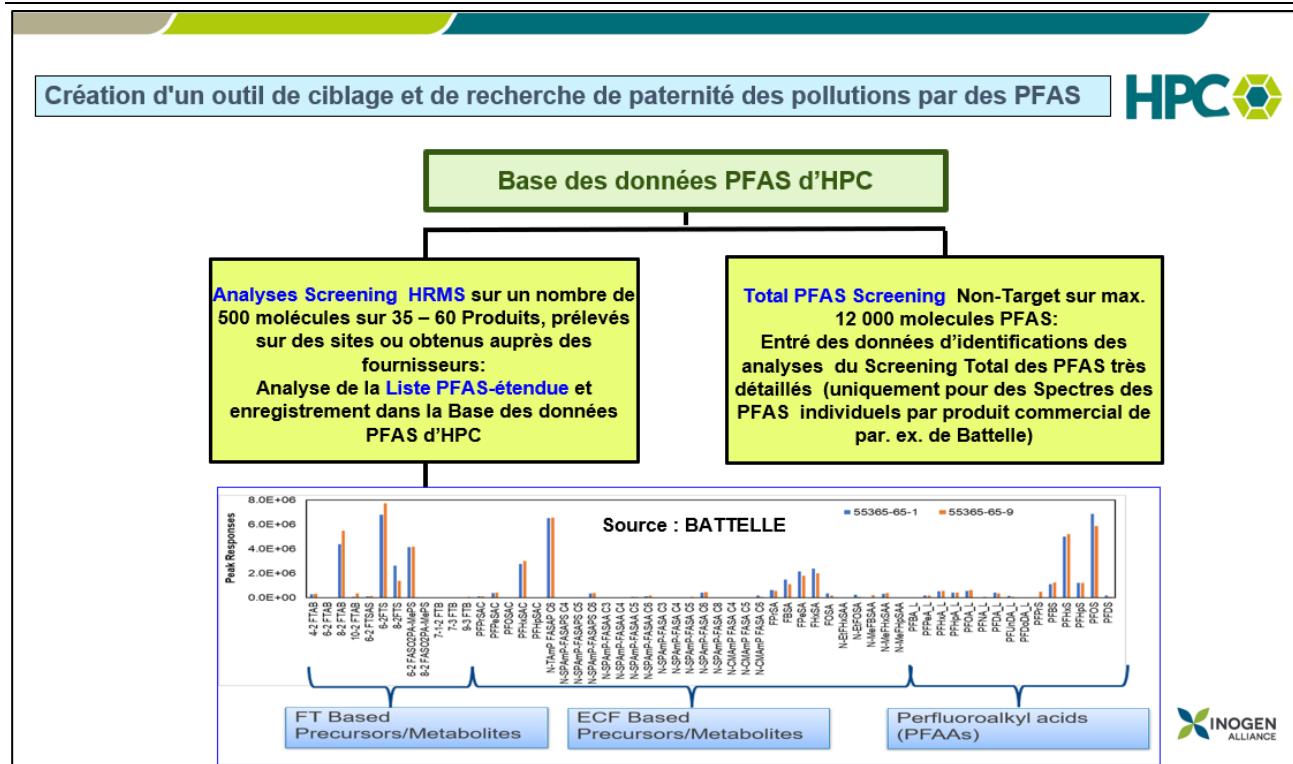


Fig. 8: HPC database of molecular spectra of commercial HPC PFAS products

- The "PFAS HPC Database" on Microsoft Access is superior to that of other Databanks, because it integrates, in addition to the known spectra (PFAS Chromatograms), those of Commercial Products subjected to aging by Lysimeters and Bacteriological Biotransformation polyfluorinated PFAS).
- A list of Queries (Possible Queries or Identification Parameters, e.g. the recognition of commercial products from the relationships of the individual PFASs of the "PFAS-signature list") will be created to identify the commercial products from the standard list of reduced PFAS analyzes (30 – 40 individual substances = "PFAS-signature list").

The principles of identifying commercial products and sources of PFAS pollution at the origin are demonstrated in Figs. 9 – 11.

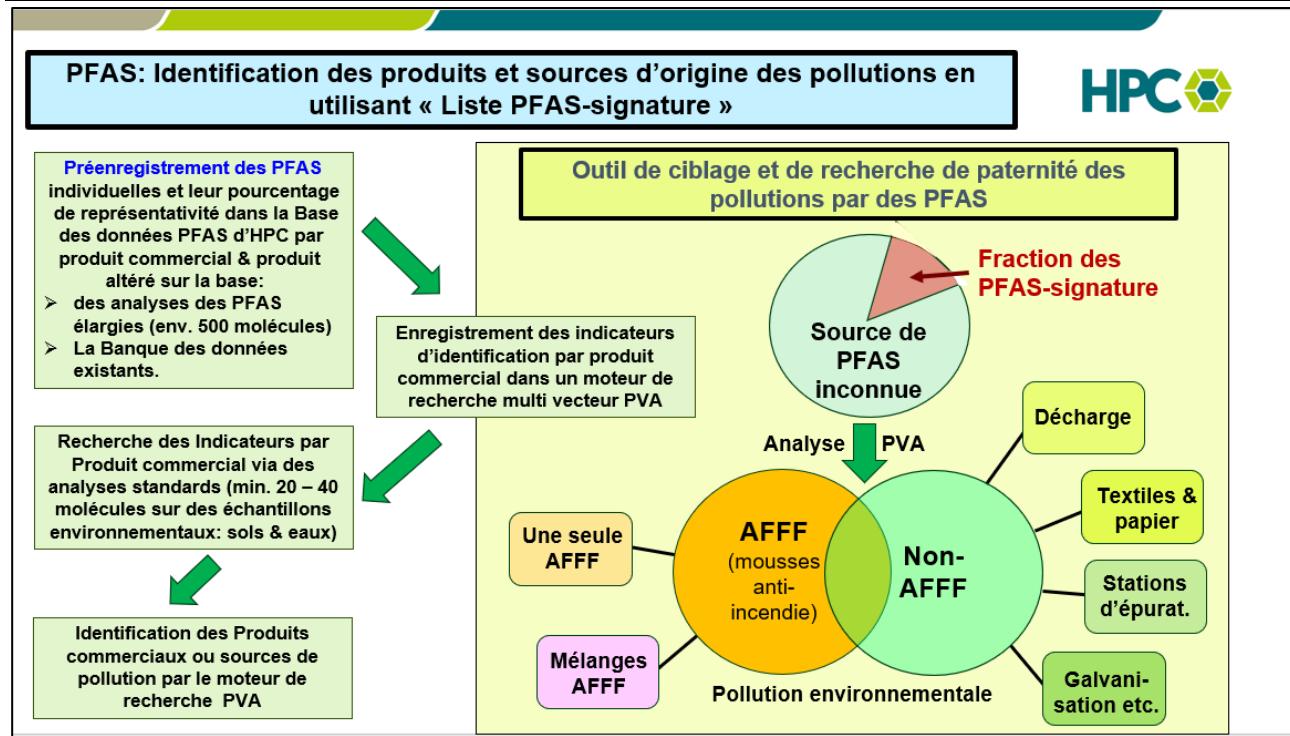


Fig. 9: Identification of initial PFAS products in mixtures of commercial products causing pollution (1/3)

The identification program concerning the origin of commercial products: The identification program makes it possible to search for the origin of commercial products on the basis of a reduced number of individual PFAS in the analyzes of environmental samples (soil and water) from the "PFAS database of HPC" on Microsoft Access (or Excel). This includes:

- The expected concentration distribution research results will be expressed as a % of the total expected concentration in soils or in groundwater, surface water or sewerage (Function A).
- Search for the identity of the commercial product(s) likely to be the source of contamination of the basement.
- The identification program indicates (in the form of a table) the probability of origin of pollution in the soil or in groundwater, surface water or sanitation compared to the commercial Products listed in the "Database PFAS of HPC".
- As part of the probabilistic identifications of the origins of PFAS contamination, the Spectra (chromatographic distributions of PFAS) of commercial Products altered and recorded in the "HPC PFAS Database" are included.
- The identification program on the "PFAS HPC Database" will be superior to those existing, because it uses, in addition to the existing spectra, also spectra (Chromatograms of PFAS) of Commercial Products subjected to aging by Lysimeters and of the Bacteriological biotransformation of polyfluorinated PFAS, etc.).

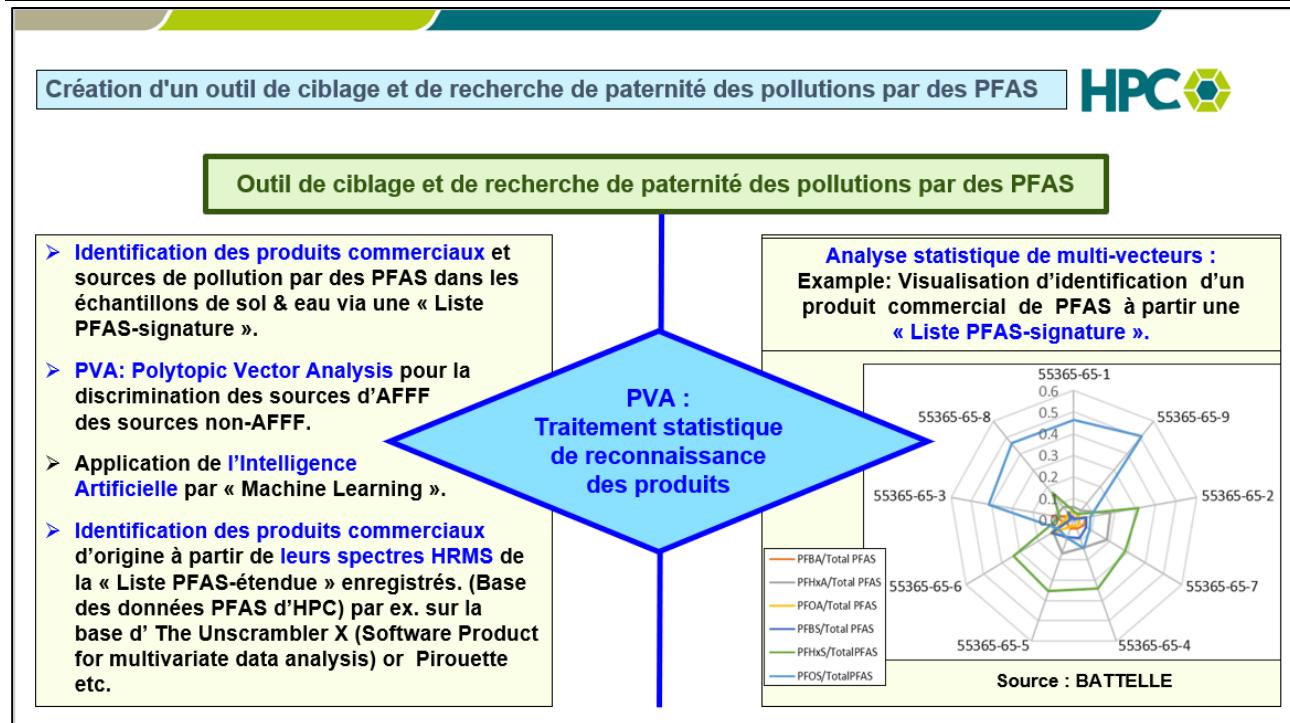


Fig. 10: Identification of initial PFAS products in mixtures of commercial products causing pollution (2/3)

- The commercial product identification program will be a statistical classification model like the one originally developed by Batelle for the detection of AFFF (and other products) in mixtures of different PFAS sources. This model will also be based on Artificial Intelligence and, in particular, on the Machine Learning approach.
- The application of the Tool will be based on the presence and quantification in the samples analyzed of restricted PFAS compounds that are extremely stable under environmental conditions, and which characterize the footprint of AFFF or other sources of additional PFAS existing identification programs on stable perfluorinated PFAS formed by the biotransformation of the weathering of commercial products.
- The results of applying the Tool are the probabilities that the mixture of PFAS found in a sample is related to certain commercial products (or pollution sources).
- Using the statistical method called PVA (Polytopic Vector Analysis), it is possible to assign a percentage of statistical membership (or resemblance) of environmental samples to each Commercial Product (or source of Pollution).
- The software used for the development of the Tool are R, Mathematica®, The Unscrambler® and Pirouette®.

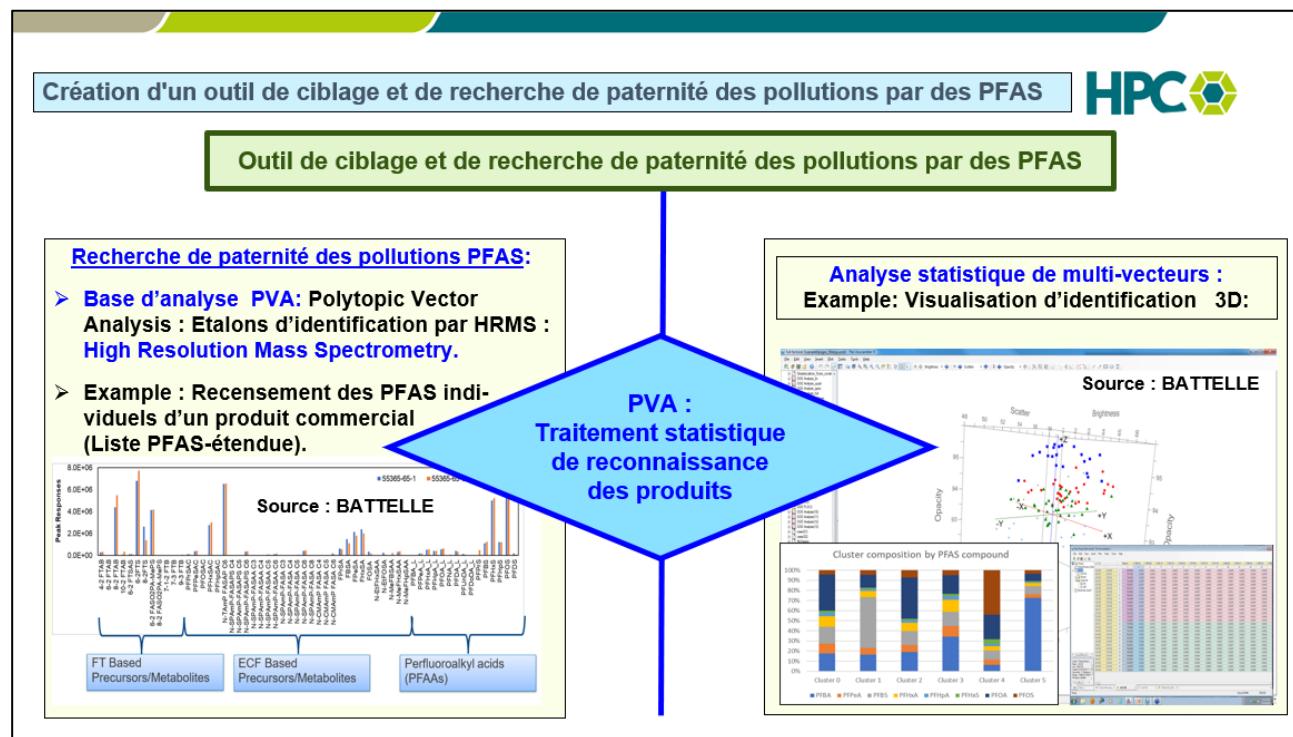


Fig. 11: Identification of initial PFAS products in mixtures of commercial products causing pollution (3/3)

The existing difficulty are the spectra (chromatograms) of the commercial products degraded by biotransformation from poly-fluorinated to per-fluorinated, because these individual PFAS spectra are different from the "fresh" products and also from those integrated into the HPC database at from existing databanks.

To overcome this difficulty of identifying altered commercial products in environmental samples (soil & water, analyzed for a reduced number of PFAS), the "PFAS-HPC Database" will contain expanded Chromatograms on 500 individual PFAS, representing PFAS spectra of commercial Products subjected to aging tests in Lysimeters These tests will be carried out via a reputable bacteriological consortium of the bio-transformation in question (of the type Acidimicrobium sp. Strain A6 (Huang & Jaffé 2019) and/or Pseudomonas strain SYC and /or Rhodococcus jostii RHA1 and/or Pseudomonas oleovorans (Ying Shi, 2018).

This is how standards representative of commercial mixtures of PFAS will be obtained by simulating aging for several years. As a general rule, it is recommended to include in the diagnosis of contamination by PFAS at least the following pollutants:

N°	Polluant PFAS de base	Synonyme	VTRs existantes
1	Acide perfluoro-butanoïque	PFBA	Oui
2	Acide perfluoro-pentanoïque	PPPeA	Oui
3	Acide perfluoro-hexanoïque	PFHxA	Oui
4	Acide perfluoro-heptanoïque	PFHpA	Oui
5	Acide perfluoro-octanoïque	PFOA	Oui
6	Acide perfluoro-nonanoïque	PFNA	Oui
7	Acide perfluoro-decanoïque	PFDA	Oui
8	Acide perfluoro-butane-sulfonique	PFBS	Oui
9	Acide perfluoro-hexane-sulfonique	PFHxS	Oui
10	Acide perfluoro-heptane-sulfonique	PFHpS	Oui
11	Perfluoro-octane-sulfonate	PFOS	Oui
12	Acide H4-polyfluoro-octane-sulfonique	H4-PFOSA	associé au PFOS
N°	Polluant PFAS de base	Synonyme	VTRs existantes
13	Perfluoro-octane-sulfonamide	PFOSA	Oui
14	6:2 Fluorotélomère alcool	6:2-FTOH	Oui
15	8:2 Fluorotélomère alcool	8:2-FTOH	Oui

Si possible:

16	Perfluorobutane sulfonate	PFBS	Oui
17	Perfluoropentane sulfonate	PPPeS	Oui
18	Perfluorohexane sulfonate	PFHxS	Oui
19	Perfluoroheptane sulfonate	PFHpS	Oui
20	Perfluorodecane sulfonate	PFDS	Oui
21	Acide perfluoro-undecanoïque	PFUnDA	Oui
22	Acide perfluoro-dodecanoïque	PFDoDA	Oui
23	Acide perfluoro-tridecanoïque	PFTrDA	Oui
24	Acide perfluoro-tetradecanoïque	PFTeDA	Oui
25	Acide perfluoro-hexadecanoïque	PFHxDA	Oui
26	Acide perfluoro-octadecanoïque	PFODA	Oui
27	Acide Hexafluoro-propylèneoxydimer	HFPO-DA	Oui
28	Acide 3H-perfluoro-3-[(3-methoxy-propoxy) propanoïc	ADONA	Oui

Fig. 12: Minimum recommended PFAS analysis parameters

As it stands today, the analysis of the 20 individual PFAS of European Directive 2020/2184 can be carried out but also extended to 40 individual PFAS. : According to European Directive 2020/2184, which concerns the quality of water for human consumption, the following 20 PFAS are targeted:

Somme des 20 PFAS de la Directive européenne Eau potable 2020 :

- Acide perfluorooctanoïque (PFOA) [5347]
- Acide perfluoroheptanoïque (PFHpA) [5977]
- Acide perfluorohexanoïque (PFHxA) [5978]
- Acide perfluoropentanoïque (PPPeA)[5979]
- Acide perfluorobutanoïque (PFBA) [5980]
- Acide perfluorobutane sulfonique (PFBS) [6025]
- Acide perfluorododécanoïque (PFDoDA) [6507]
- Acide perfluorononanoïque (PFNA) [6508]
- Acide perfluorodécanoïque (PFDA) [6509]
- Acide perfluoroundécanoïque (PFUnDA) [6510]
- Acide perfluoroheptane sulfonique (PFHpS) [6542]
- Acide perfluorotridecanoïque (PFTrDA) [6549]
- Acide perfluorodécane sulfonique (PFDS) [6550]
- Acide perfluorooctane sulfonique (PFOS) [6561]
- Acide perfluorohexane sulfonique (PFHxS) [6830]
- Acide perfluoropentane sulfonique (PPPeS) [8738]
- Acide perfluorononane sulfonique (PFNS) [8739]
- Acide perfluoroundécane sulfonique [8740]
- Acide perfluorododécane sulfonique [8741]
- Acide perfluorotridécane sulfonique [8742]

4. Health Risk Assessments (TERQ : Toxicological Exposure Risk Quantification)

For the assessment of health risks, toxicological data (TRV: Toxicological Reference Values) must be sought and updated at international level almost on a weekly basis. The most recent TRVs are available mainly in the USA (EPA, etc.), ATSDR and EFSA. ANSES also published PFAS TRVs in 2017, but given the forced advancement of toxicological studies, these TRVs are for the most part already outdated.

In the case of FTOHs present in soils, groundwater and soil gases, it will also be imperative to carry out FTOH investigations in the ambient air of buildings with sensitive uses (ERP: Schools, Nurseries, etc.) or residential, on the basis of quantification thresholds (or at least for the detection thresholds), of the order of 4 – 8 ng/m³, in order to have a good exploitable basis for the EQRS: Quantitative Assessment of Health Risks.

A first simplified risk assessment approach is possible via existing limit values, e.g. in Germany, or published by the European Community. In Germany there are limit values for drinking water, soil and groundwater.

The US-EPA considers that the majority of exposure today comes from drinking water, outside of polluted sites, due to the lack of monitoring and treatment of PFAS. According to Directive (EU) 2013/39/EU "Water Framework" (WFD), concerning PFOS & derivatives (and other priority substances) an Environmental Quality Standard (EQS-MA) of 0.65 ng/l for surface waters and 0.13 ng/l for waters from the marine environment (and EQS-MAC: Maximum Allowable Concentrations) has been set.

It is important to note that a simple application of the Limit Values, generic and individual within the framework of a simplified risk assessment does not take into account specific exposure scenarios and exposures to mixtures ("Cocktails") pollutants with, at a minimum, taking into account the additivity of the risks of pollutants having the same targets and toxicological effects. Consequently, it is preferable to carry out EQRS (or ARR, HRA, TERQ), which corresponds well to the French Methodology for the Management of Polluted Sites, according to the note from the Ministry in charge of the Environment, of 04/19/ 2017.

Another important aspect is that only an EQRS (or ARR, TERQ, HRA) will make it possible to define Health Compliance Control Values, in the form of MAC (Maximum Admissible Concentration) for maximally acceptable Excess Individual Risks (of cancer): ERI < 10E-5 or a Systemic Risk Index of IR < 1 (= DJE / DJT: Daily Exposure Dose over the Tolerable Daily Intake). Regarding the EQRS: Quantitative Assessment of Health Risks, the basis is either the measurement of concentrations in the exposure media, or the modeling of the transfer of pollutants from one compartment to another (e.g. pollutants in groundwater or from the ground to the gases of the ground and the ambient air. An important step in the EQRS is the choice of TRVs (Toxicological Reference Values), because their evolution is rapid.

For example, the 2017 ANSES PFAS guide includes certain TRVs for PFAS, but given the many TRV publications to date, these values are partially outdated, and in particular much more restrictive to date. A Tolerable Weekly Intake (DHT) of 4.4 ng/kg/Week (or the Tolerable Daily Intake (TDI) of 0.63 ng/kg/d for PFAS: PFOA, PFOS, PFNA & PFHxS) has been published by EFSA, 09/17/2020. In 2020, toxicity equivalence factors with respect to PFOA were also published by W. Bil et al. in the form of RPF: Relative Potency Factors.

In order to ensure the correct choice of TRVs for PFAS, it is recommended to apply scientific selection criteria and not national criteria. Fig. The following shows criteria for choosing the applicable TRVs, in order to take into account the best toxicological knowledge concerning the dose-effect relationships of PFAS.

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No	TRD: Toxicological Reference Dose Choice Criteria	Appreciation			
		Favorable	Correct	Not favorable	Exclusion
1	Variability of indicated TRD	(+/- 0 %)	≤ (+/- 30 %)	> (+/- 30 %)	
2	Class (potential) Carcinogenic: EC: Class 3/ US-EPA: Class B2, C / IARC: Group 1	3 Organisms : CE, US-EPA, IARC, etc.	2 Organisms	1 Organisms	
3	Several Organisms shows similar TRD (+/- 50 %)	≥ 3 Organisms	2 Organisms	1 Organism	
4	Age of base Study	≤ 15 a	15 – 25 a	< 25 a	
5	Mechanistic toxicological basement Study (for ex. Genotoxicity):	Epidemiology	Mamifer	In-Vitro / In-silico	
6	Basement Study : Klimisch Quality Criteria	Class 1	Class 2	Class 3	Class 3
7	Verified Purity of Compound	Yes	< 95 %	No	
8	Excipient potentially toxic	Non		Yes	
9	Presence of population without exposure (test witness)	Yes		No	
10	General Quality Criteria (Klimisch) of toxicological effect studies	Standardized Study (OCDE, UE, US EPA, FDA, etc.)	Standardized Study without Details, but correctly documented	Document insufficient for evaluation, systematic deficiencies	
11	POD : Point of Departure	Quantified Epidemiological Data, BMLD, etc. (PBPK)	NOAEL sensitive NOAEL	LOAEL sensitive, LOAEL, Other	
12	Uncertainty (or Assessment) Factors	1 – 100	> 100 – 1000	> 1 000 – 10 000	> 10 000
13a	Transpositions: Between Exposure Pathways	Non		Yes	
13b	Transposition: Animal to Human	Non	Yes		
13c	Transpositions : From in-Vitro	Non		Yes	
13d	Transpositions : From in-Silico	Non		Yes	
14	Study time-representatively	≥ chronic (> 180 d)	sub-chronic (90 d) to chronic (180 d)	< sub-chronic (< 90 d)	
15	Integration of bio-disponibility / Bio-resorption capacity (ex: DIN 19 738)	Yes	Not known (100 %)	Known, but not considered	

Fig. 12: TRV selection criteria (F. KARG 2022)

Based on these selection criteria, it is possible to define e.g. the TRVs indicated in FIG. following (dating from the beginning of the year 2023). It is important to take into account that the Administrations of certain US States are very advanced in toxicological research and the publication of TRVs concerning PFAS, due to their strong presence in the chemical and oil industries (with heavy uses history of AFFF) or the strong presence of large industrial landfills (and the associated environmental pollution). These include the states of Texas, Michigan and New Jersey. These TRVs are often the basis for federal US-EPA PFAS TRV publications.

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Substance	Cancerogen / not cancerogen	Chronic toxicological value			Species	Sigle	Security Factor	Organization
		Exposure path	Target organ	Value				
PFBA	NC	oral	Hepatic	2,9 µg/kg/d	Rate	RfD	NOAEL / 2400	TCEQ 2016
		inhalation	Hepatic	10 µg/m³	Rate	RfC	from oral value	TCEQ 2016
PPPeA	NC	oral	Hepatic	3,8 µg/kg/d	Rate	RfD	same than PFHxS LOAEL/(263*300)	TCEQ 2016
PFHxA	NC	oral	Hepatic	3,8 µg/kg/d	Rate	RfD	same than PFHxS LOAEL/(263*300)	TCEQ 2016
PFHpA	NC	oral	Hepatic	25 ng/kg/d	Rate	DJT	Extrapolation of DJT of Health Canada	ANSES 2017
PFOA	NC	oral	Hematologic	0,86 ng/kg/d	Rate	TDI	BMDL5	UBA 2020 BfR & EFSA 2018
			Hepatic, Mammar, Hematologic	12 ng/kg/d	Mice	RfD	LOAEL (81*100)	TECQ 2016
		inhalation	Hepatic	4,1 ng/m³	Rate	RfC	NOAEL / (81*3000)	TCEQ 2016
	C	oral	Testicular tumors	2,52 (mg/kg/d)⁻¹	Epidemio	SF	-	New Jersey 2017
PFNA	NC	oral	Hematologic	2,5 ng/kg/d	Mouse	RfD	NOAEL / 300	EPA IRIS 2019 New Hampshire DES 2019
		inhalation	Lung, respiratory system	28 ng/m³	Rate	RfC	NOAEL / (81*30 000)	EPA IRIS 2019 TCEQ 2018
PFDA	NC	oral	Hepatic	15 ng/kg/d	Rate	RfD	NOAEL / (81*1000)	TCEQ 2016
				53 ng/m³	Rate	RfC	from oral value	TCEQ 2016
PFBS	NC	oral	Hematologic and renal	1,4 µg/kg/d	Rate	RfD	NOAEL / (142*300)	TCEQ 2016
				4,9 µg/m³	Rate	RfC	from oral value	TCEQ 2016
PFHxS	NC	oral	Hematologic and thyroïdal	3,8 µg/kg/d	Rate	RfD	LOAEL / (263*300)	TCEQ 2016
				13 ng/m³	Rate	RfC	from oral value	TCEQ 2016
PFHpS	NC	oral	Hepatic	0,43 ng/kg/d	Rate	TDI	Potency Factor : 0,6-2	UBA 2020, EFSA 2018, BfR 2018
PFOS	NC	oral	Hepatic	1,86 ng/kg/d	Monkey	TDI	NOAEL	UBA 2020 BfR & EFSA 2018
		inhalation	Thyroidal, neurological and foetal development	81 ng/m³	Rate	RfC	from oral value (23 ng/kg/j)	TCEQ 2016
PFOSA	NC	oral	Mammary glands	12 ng/kg/d	Mice	RfD	Same than PFOA NOAEL/(81*300)	TCEQ 2016
		inhalation		4,1 ng/m³	Rate	RfC	same than PFOA NOAEL/(81*3000)	TCEQ 2016

Fig. 13: Choice of certain TRVs according to the criteria in FIG. 7 (F. KARG 2022):

- ANSES: Agence nationale de sécurité sanitaire de l'alimentation, de l'environnement et du travail (2017)
 ATSDR: Agency for Toxic Substances and Disease Registry
 EFSA: European Food and Safety Authority
 IRIS : Integrated Risk Information of Substances (U.S. - EPA)
 UBA : Umweltbundesamt (Germany)
 BfR: Bundesinstitut für Risikobewertung (Germany)
 OEHHA : Office of Environmental Health Hazard Assessment
 WHO: World Health Organization
 RIVM : Netherlands Environmental & Health Institute
 MDHHS: Michigan Department of Health and Human Services, Division of Environmental Health
 TCEQ: Texas Commission on Environmental Quality
 NJ-DWQIHES: New Jersey Drinking Water Quality Institute Health Effects Subcommittee
 Bil et al. 2020 : Toxicological Equivalence Factors on PFOA RfD

A complementary step to the EQRS (or ARR, TERQ, HRA) is the definition of health compliance control values, in the form of MACs (Maximum Admissible Concentrations) by integrating an additivity of the risks of pollutants concerning the same targets and toxicological effects, for

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maximally acceptable Excess Individual Risks (of cancer): ERI < 10E-5 or a Systemic Risk Index of IR < 1 (= DJE / DJT: Daily Exposure Dose over Tolerable Daily Dose). MACs are commonly used in the form of Sanitary Control Values, in order to verify or co-develop corrective action objectives, or even depollution objectives. Management measures, e.g. depollution are based in France on a Management Plan, a definition of the Source Zones of concentrated pollution and then a Cost-Benefit Balance Sheet of the different management and treatment methods and technologies.

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Sites pollués par des PFAS : état des problèmes et des solutions en développement en Suisse sur la base des expériences du canton de Genève

Sites polluted by PFAS: state of the problems and solutions under development in Switzerland based on the experiences of the canton of Geneva

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La Suisse est confrontée, comme tous les autres pays, à des pollutions ponctuelles et diffuses par les PFAS. Ces dernières années, de nombreuses mesures ont mis en évidence ces substances à des concentrations très élevées dans des eaux souterraines et à des concentrations plus faibles mais néanmoins préoccupantes dans les sols et les eaux de surfaces.

La législation suisse sur les sites pollués et celle sur l'élimination des déchets ne contient pas encore de normes pour évaluer et traiter ces pollutions. Néanmoins, les autorités doivent régulièrement statuer sur le besoin et les buts d'assainissement d'un site contaminé par les PFAS ou autoriser les filières d'élimination pour des matériaux pollués par des PFAS. En attendant que les bases légales soient créées, des valeurs limite sont donc définies au cas par cas sur la base des connaissances toxicologiques actuelles et des expériences réalisées en Suisse et à l'étranger.

Le canton de Genève a recensé dans son cadastre des sites pollués, public depuis 2004, 877 sites, dont 475 aires d'exploitation, 378 sites de stockage définitif et 24 lieux d'accident. 85 % des investigations, surveillances et assainissements sont réalisés et 11 sites contaminés sont totalement assainis. Depuis 2017, quatre campagnes d'échantillonnage et d'analyses des PFAS ont été réalisées dans les eaux souterraines et de surface, ce qui représente environ 450 prélèvements et analyses sur 111 points. Les valeurs les plus élevées ont été retrouvées principalement dans des résidus de mousses d'extinction d'incendie (décharges ou industries importantes) mais également en lien avec certaines activités. Une problématique importante concerne l'assainissement complémentaire de sites déjà assainis pour d'autres substances.

Pour la suite, les travaux devront se concentrer sur la recherche des sites de protection civile, d'entraînement des pompiers et où l'historique indique des incendies importants. Les décisions d'investigation, d'assainissement et de surveillance devront être priorisées selon les concentrations en PFAS analysées. Finalement, le cadastre des sites pollués devra être actualisé sur la base des recommandations de l'OFEV afin d'intégrer les sites pollués par les PFAS.

Au niveau fédéral, l'office fédéral de l'environnement a mis sur pied plusieurs groupes de travail avec les cantons afin de définir les démarches pour l'inscription au cadastre des sites pollués par des PFAS, leur investigation, l'évaluation des risques ainsi que leur assainissement. Pour l'instant, l'évaluation de l'atteinte aux eaux souterraines et de surface est réalisée par l'analyse de 9 substances PFAS ; la valeur de concentration à appliquer est de 50 ng/l (somme des substances pondérée par les facteurs de puissance relative).

Les connaissances sur les pollutions existantes, comme celles du canton de Genève, sont des bases de travail indispensables afin de pouvoir fixer dans la législation les valeurs seuils. En complément, des analyses de sols et des eaux souterraines, réparties sur tout le territoire, doivent permettre de définir la pollution de fond aux PFAS. C'est donc une procédure itérative qui est en cours : les expériences réalisées par les cantons alimentent et font évoluer ce qui sera la base du système réglementaire qui est ensuite appliqué sur des cas réels.

Surveillance des PFAS dans les eaux souterraines en France : Etats des lieux et enjeux

Groundwater Monitoring of PFAS in France: Current status and challenges

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Parmi les milliers de composés appartenant à la large famille des poly et per-fluoroalkylés regroupés sous l'acronyme PFAS, 20 composés sont réglementés en France dans les eaux souterraines. Il existe cependant un éventail beaucoup plus large de PFAS.

Dans le monde, différents travaux ont révélé l'accumulation des PFAS dans les sols, en particulier sur les sites touchés par les activités de lutte contre les incendies, les industries utilisant ou fabriquant des composés PFAS mais aussi les sources agricoles telles que l'application de biosolides et l'irrigation avec des eaux usées traitées et les dépôts atmosphériques. Un nombre croissant d'études montrent que les zones non saturées (ZNS) des aquifères, c'est-à-dire la zone entre le sol et la surface de la nappe phréatique, représentent des réservoirs de PFAS, constituant une menace à long terme pour la contamination des eaux souterraines. De plus, en France, deux tiers des volumes d'eau prélevés pour l'alimentation en eau potable (AEP) proviennent des eaux souterraines (OFB, 2017¹), la surveillance des PFAS dans ce milieu apparaît comme essentielle, pour documenter la répartition spatiale, la dynamique d'évolution et anticiper les impacts potentiels sur la qualité des eaux.

Contexte réglementaire

En France, la surveillance des eaux souterraines est réglementée par l'arrêté du 25 janvier 2010 établissant le programme de surveillance de l'état des eaux. A partir de 2015, 6 composés de la famille des PFAS ont été inscrits dans la liste des micropolluants de l'analyse régulière du contrôle de surveillance (Annexe VIII). Cette liste a été complétée lors de la dernière révision, publiée le 26 avril 2022, et inclut dorénavant 20 composés, correspondant à ceux ciblés dans la Directive des eaux destinée à la consommation humaine (EDCH).

Peu de pays ont dérivé des valeurs seuils pour les eaux souterraines, généralement pour les composés individuels, PFOA et PFOS. En France, il n'existe actuellement pas de valeur seuil pour les PFAS individuels dans les eaux souterraines. Néanmoins, La refonte de la Directive européenne sur les eaux destinées à la consommation humaine (Directive 2020/2184) du 30 décembre 2020 et la proposition du 26 octobre 2022 de révision à la Directive 2006/118/CE proposent des valeurs seuils pour une somme de PFAS. Ainsi, l'arrêté du 30 décembre 2022 modifiant l'arrêté du 11 janvier 2007 relatif aux limites et références de qualité des eaux brutes et des eaux destinées à la consommation humaine régule les PFAS avec une approche par groupe avec (i) 0,1 µg/L pour la somme de 20 PFAS dans l'eau distribuée et (ii) 2 µg/l dans les eaux brutes (eaux souterraines et de surface). Bien que la Directive EDCH propose le seuil de 0,5 µg/L pour le paramètre « PFAS totaux », celui-ci reste en cours de définition et n'a pas été retroussé dans l'arrêté national.

De par les enjeux de performance et la multiplicité des molécules concernées, il existe un fort enjeu analytique autour des PFAS, que ce soit sur les paramètres à évaluer selon les objectifs réglementaires mais aussi le besoin de développer des méthodes indiciaires permettant une vision plus globale de la contamination. Les développements méthodologiques sont actuellement en cours notamment dans le cadre du projet H2020 PROMISCES (Grant Agreement n°101036449) pour le

¹ https://www.eaufrance.fr/sites/default/files/2018-06/prelevements_2013_201703.pdf

déploiement de ces méthodes mais aussi de leur contexte d'application, en vue de leur transfert vers des laboratoires en charge de la surveillance réglementaire.

Occurrence en PFAS dans les eaux souterraines

Une analyse des données disponibles dans la banque de données ADES a été réalisée pour l'ensemble des 20 PFAS prescrits dans l'arrêté surveillance pour les eaux souterraines.

L'analyse de ces composés est encore difficile et les très faibles concentrations généralement observées imposent un haut niveau d'expertise technique et les laboratoires ayant mené des campagnes de mesures doivent respecter des critères d'assurance qualité. Les données présentées ont été produites par les Réseaux opérés par les Agences de l'eau pour le rapportage de l'état des masses d'eau souterraines au niveau européen : le réseau de Contrôle de Surveillance (RCS) et le Réseau de Contrôle Opérationnels (RCO)) complété par le réseau des services Santé-Environnement Eau (SISE-Eau) opéré par les ARS.

Du point de vue spatial, les représentations cartographiques d'occurrences en PFAS ont permis de documenter leur répartition en France et de cibler les zones d'intérêt, tant d'un point de vue des secteurs géographiques mais également des contextes hydrogéologiques à cibler.

Du point de vue temporel, c'est à partir de la période 2009-2012 que les premières analyses à l'échelle nationale sont reportées dans ADES. C'est en particulier le résultat des campagnes exceptionnelles (CAMPEX) réalisées sur les eaux souterraines en 2011 et 2012.

A l'échelle de l'ensemble du jeu de données analysé, les paramètres les plus régulièrement suivis sont PFOA, PFHpA, PFHxA, PFOS et PFHxS. Il s'agit principalement des paramètres de la liste de l'arrêté du 7 août 2015. Le suivi des composés, PFPeS, PFNS, PFUnDS, PFDoDS, PFTrDS a débuté en 2022. En 2022, les données montrent que les principales substances quantifiées dans les eaux souterraines en France, sont PFOS, PFHxA, PFOA, PFHxS. Les substances suivantes sont également quantifiées PFHpA, PFPeA, PFBS, PFBA et PFHPeS. A l'inverse, PFTrDA, PFNS et PFDoDS n'ont pas encore été quantifiée d'après les données disponibles en mai 2023.

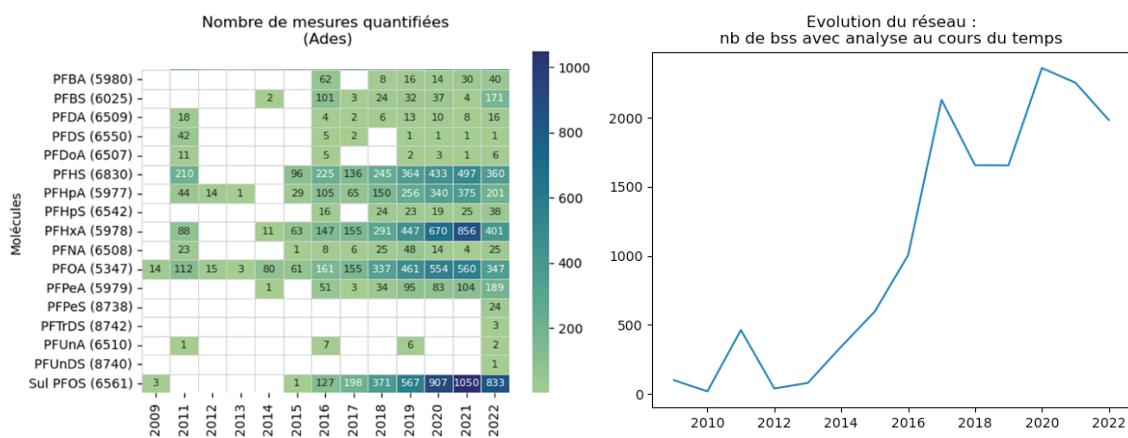


Figure : Nombre de mesures quantifiées par année et par paramètre (fig. à g.) et Evolution du nombre de point d'eau surveillé (fig. à d.) dans le cadre de la surveillance des eaux souterraines (données ADES extraites au 30 mai 2023 – bancarisation de l'année 2022 incomplète).

Les valeurs seuils actuellement en discussion pour les eaux souterraines et les eaux destinées à la consommation humaine auront ainsi une importance capitale dans l'évaluation du risque environnemental et sanitaire que représente ces composés. En utilisant les données encore parcellaires actuellement disponibles, l'application de la valeur de 4,4 ng/L proposée par le SCHEER (révision publiée le 26 octobre 2022) pour la somme pondérée des 24 PFAS aboutirait, par exemple, au déclassement de plus de 230 masses d'eau souterraine.

Les limites de quantification sont très variables selon les campagnes de mesures et les laboratoires. La couverture nationale étant pilotée à l'échelle des agences de l'eau et par différents prestataires, cela rend difficile l'analyse de la dynamique d'évolution de ces données.

Ainsi, il apparaît pertinent et indispensable de considérer avec attention la surveillance de ces composés, notamment en raison d'un possible effet de stock dans le sol et la ZNS, qui pourrait jouer un rôle d'effet retard au regard de la contamination des eaux souterraines.

Occurrence of PFAS in drinking water in Canada and across the world

Occurrence des PFAS dans l'eau potable au Canada et à travers le monde

Sébastien Sauvé et Termeh Teymoorian

We have analysed hundreds of drinking water samples for a large suite of target and non-target PFAS. Preliminary analysis has shown that PFAS are detectable in over 95% of the collected samples with the sum of PFAS ranging from below detection limits up to above 200 ng/L. Given the recent developments and turmoil in the determinations of drinking water recommendations for PFAS, this raises many questions on our current exposure to PFAS, how should we regulate their presence in drinking water and how should we implement restrictions or bans on the use of PFAS in the industries that contribute to their environmental release. It also raises questions on the means and options available to treat water to remove PFAS.

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Advanced fingerprinting analysis of PFAS in groundwaters: the use of advanced multivariate statistics and Machine learning techniques

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PFAS are omnipresent in environmental media and are distributed globally. Their potential to enter the food chain is high, thanks to their ability to migrate from water to soils and be taken up by plants. These substances can enter the human body through food, drinking water, and direct skin contact; moreover, they are not metabolized and can bio-accumulate. Contact with PFAS-containing materials (e.g., water-resistant materials, detergents, paints, fabrics) can also increase human exposure. The scope of this project is to demonstrate an approach to identify original PFAS sources using a multivariate method (PCA and cluster analysis) and PVA (polytopic vector analysis) algorithm developed in R by the authors.

Together with this example, we will present a new advanced analytical tool, PFAS Signature® (Figure 1), for PFAS source differentiation and tracking using high-resolution mass spectrometry (HRMS) techniques, in combination with PFAS targeted analysis and advanced statistical analysis. It is also an effective tool for establishing the background of PFAS at a site at the beginning of a site investigation. The identification of sources of contamination is based on chemical signature, isomeric profiles, manufacturing, age of release, fate and transport, and biotransformation products. By understanding the specific signature of these analytes in different matrices, information can be deduced about their sources. Artificial intelligence and machine learning (AI/ML) tools are then applied to the filtered data obtained from the suspect screening libraries to compare and differentiate the PFAS composition and trends seen in field-derived samples during a site investigation. This new method improves dramatically the more conventional chemical forensic approach based on PCA and PVA.

This case study is in the Veneto Region of Northern Italy. Data were obtained from a public water monitoring database¹ containing more than 8,000 samples analyzed for PFAS in groundwater, surface water, springs, and treated industrial wastewater discharged to surface water. After database review, nearly 1,400 samples and eight PFAS out of 18 (PFBA, PFPeA, PFBS, PFHxA, PFHpA, PFHxS, PFOA, PFOS) were deemed useable for fingerprinting. The distribution of the concentration of the three major compounds are reported in the following figures with reference to the limits enforced in the Veneto Region legislation and with regards to the geographic position of the alleged source of contamination (Figure 1, Figure 2 and Figure 3).

¹ <https://www.arpa.veneto.it/dati-ambientali/open-data/idrosfera/concentrazione-di-sostanze-perfluoroalchiliche-pfas-nelle-acque-prelevate-da-arpav>

PFOA

Exceedance Category ● <= 500 ng/L ● Exceeds 500 ng/L ○ Nondetect

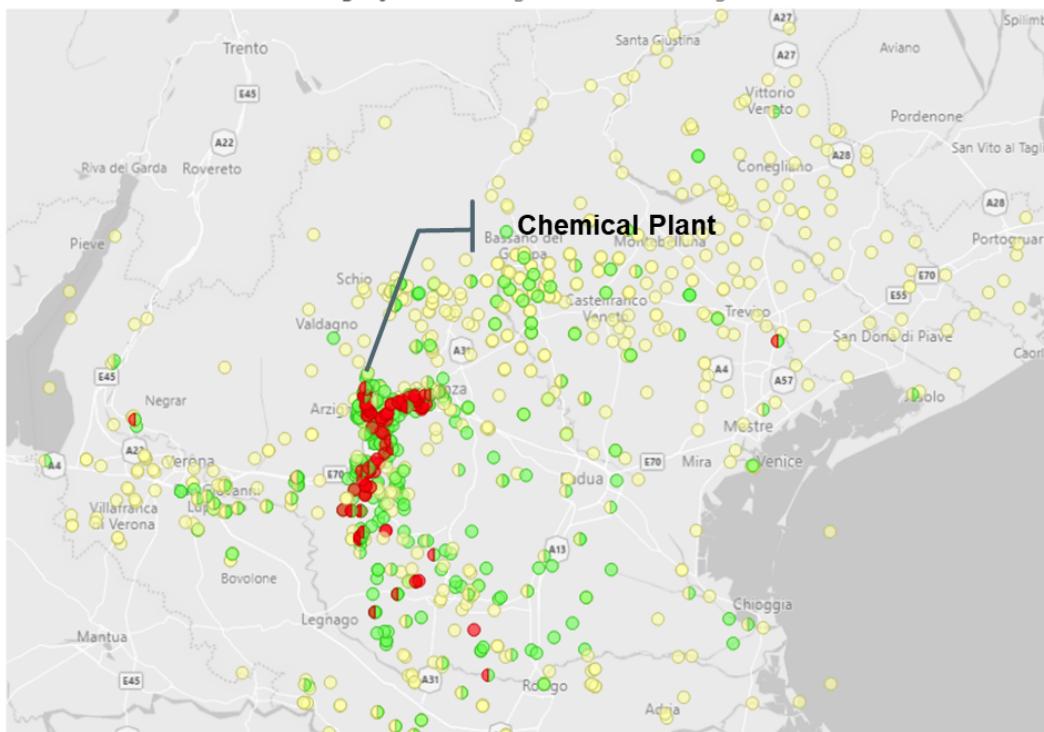


Figure 1: distribution of PFOA concentrations in the surface waters and groundwaters of the study area.

PFBS

Exceedance Category ● <= 500 ng/L ● Exceeds 500 ng/L ○ Nondetect



Figure 2: distribution of PFBS concentrations in the surface waters and groundwaters of the study area

PFBA



Figure 3: distribution of PFBA concentrations in the surface waters and groundwaters of the study area

The fingerprint of a nearby chemical plant manufacturing PFAS (the alleged source of PFAS contamination) was also derived from public data. The PCA (Figure 4), cluster analysis (Figure 5 and **Erreur ! Source du renvoi introuvable.**), and PVA results indicate that in certain parts of the Veneto region groundwater and surface water contamination is associated with the nearby PFAS manufacturer and industrial sources, while other sources are responsible for PFAS contamination elsewhere in the Region. The results highlight the ability of the approach to differentiate original sources of PFAS contamination, particularly when source fingerprints can be derived.

A classification model has been developed based on Machine learning approach. The model has been employed for the automatic classification of samples to a certain source with a certain probability. The performance of the model is very high and similar models should be developed for the analysis of other cases.

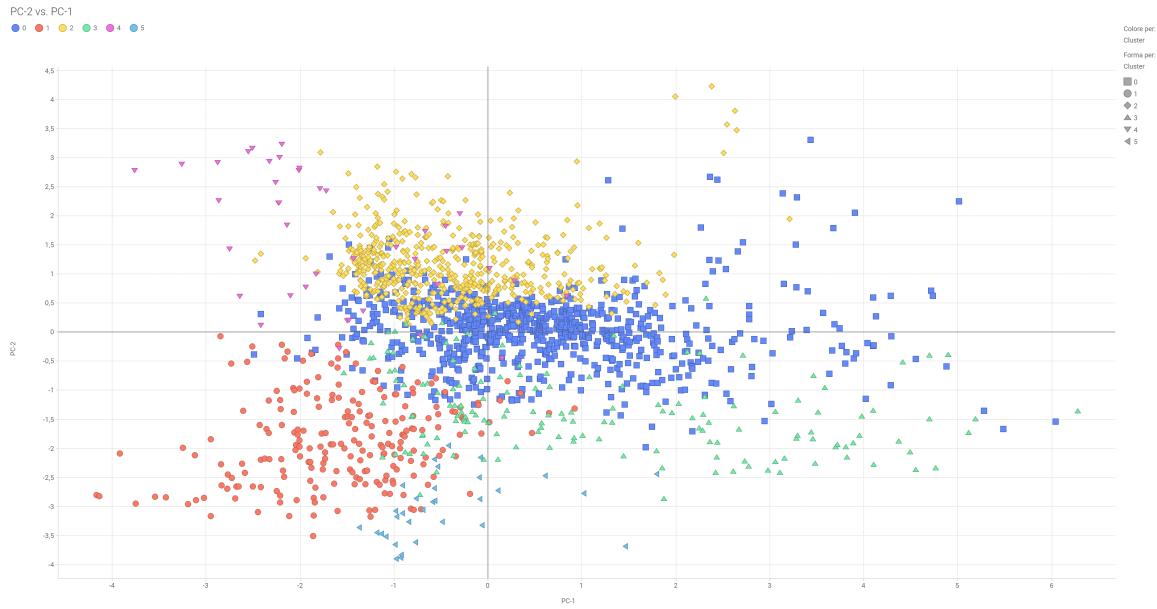


Figure 4: results of the PCA as applied to all the samples that remained in the database after the quality screening. The database contains samples of groundwater, freshwater, springs, industrial wastewaters and samples related directly to MITENI groundwater. The cluster that represents the MITENI fingerprint is the yellow one.

The main conclusions that should be derived from this preliminary analysis are:

- The final composition of the sources (clusters) derived by PVA and cluster analysis are reported in the following figures (Figure 5 and Table 1). The fingerprint of each cluster is represented in the first figure. Each cluster have a well defined fingerprint;
- Clearly cluster 2 (the alleged source one) and 4 (Industrial wastewater one) show a completely different fingerprint. The first is dominated by PFOA while the second by PFOS;
- Cluster 1 is dominated by PFBS and cluster 5 and 3 by PFBA. Different dominant PFAS compounds indicate, very often, a different origin of the contaminant.
- All the Alleged source samples belong to Cluster 2. A certain number of Groundwater and Freshwater samples belong to this cluster indicating a similar source of contamination.
- The principal limits of these analysis are also presented. A classification model has been developed based on Machine learning approach. The model has been employed for the automatic classification of samples to a certain source with a certain probability. The performance of the model is very high and similar models should be developed for the analysis of other cases.

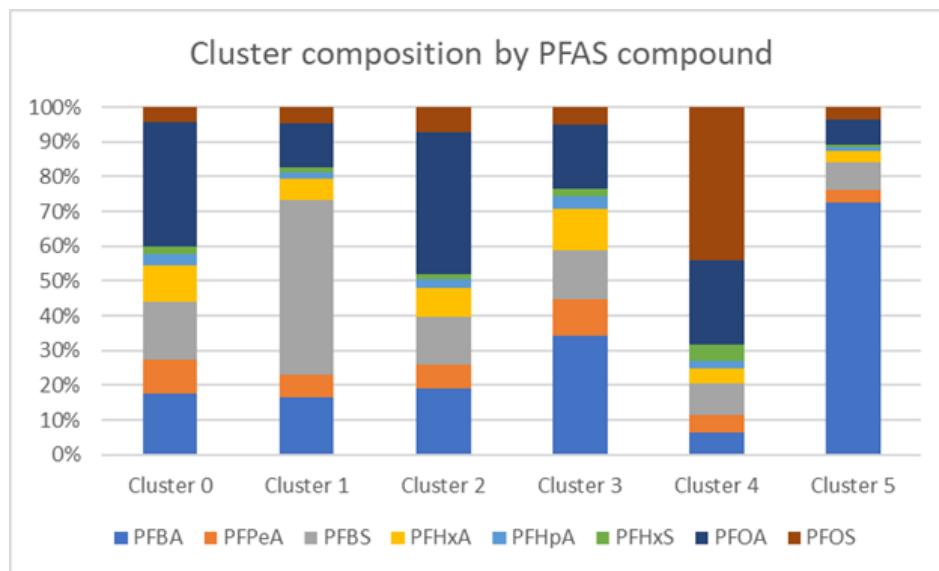


Figure 5: chemical composition of the clusters. Cluster 2 represents the fingerprint of the alleged source.

	Groundwater various sources	Freshwaters various sources	Industrial discharge	Alleged source	Springs	TOTAL samples
Cluster 0	527	241	29	0	18	815
Cluster 1	579	359	80	0	19	1037
Cluster 2	360	120	0	16	39	535
Cluster 3	106	45	4	0	3	158
Cluster 4	40	1	0	0	0	41
Cluster 5	20	10	2	0	0	32
TOT	1632	776	115	16	79	2618

Table 1: composition of clusters by source

PFAS groundwater pollution in Veneto Region (Northern Italy): regulatory implications on environmental harm and judgment case

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In 2013, a survey study undertook by the Italian National Research Council (IRSA-CNR), revealed the high presence of perfluoro-alkyl substances (PFAS) in groundwaters and sediments in the Agno-Fratta Gorzone basin, showed increasing concentrations of sum of PFAS well above 70.000 ng/l. The main source of contamination was identified near Trissino in the Province of Vicenza, where a fluoro-chemical plant was producing PFAS compounds since the early '60. The groundwater contamination plume extends over an area close to 200 km² and affects both public water supply and private wells. The area of impact was later classified according to the drinking water thresholds and it includes 30 Municipalities, where 140.000 people are directly exposed to the PFAS pollution. Consequently, a series of emergency and risk management undertakings were implemented, including the biomonitoring and environmental surveys and groundwater modelling, the drafting of water safety plan (WSP), the legislative and regulation measures with the settings of Environmental Quality Standards for ground and surface water. In particular, anticipating the implementation of the holistic risk analysis model as required by the Directive (EU) 2020/2184, the WSP allowed to better characterize the risk by identifying and implementing mitigation and reduction strategies at different time scales: at short term with the installation of more than 30 GAC filtration units and, at long term, through the interconnection between water distribution networks, the drilling of new, alternative water supply wells, the set up of state-of-the-art monitoring network, the implementation of R&D projects to support the development of more effective treatment devices, that currently sizes more than 110 M Euro.

Following these events, the Vicenza criminal prosecutor's office instituted a criminal proceeding for some crimes offences contained in the Italian penal code. Crimes of poisoning the groundwater, intentional environmental disaster, and culpable pollution. With a decree dated 26 April 2021, the Judge for Preliminary Investigations at the Court of Vicenza - following the preliminary hearing - ordered the indictment for all the defendants and for all the disputed crimes before the Assize Court. The trial, still ongoing, will have to ascertain: (1) the existence of the offense to public safety (as regards the crime of environmental disaster) and (2) the significance and measurability of groundwater pollution (for the dispute relating to the crime of environmental pollution). The Italian judicial case is particularly significant since, despite the absence of regulations establishing threshold values for perfluoro-alkyl (PFAS) substances in groundwater, very serious crimes are charged both against natural persons and against the company that produces them. Criminal proceedings are also an excellent point of observation for evaluating the proposal for a directive on the criminal protection of the environment, which, presented by the European Commission on 15 December 2021, is aimed at replacing directive 2008/99.

The first scientific publication proposing the analytical measurability of PFAS in environmental matrices dates back to 1999 (J. Field et al., 1999). The method included an enrichment through SPE (solid-phase extraction, SPE) and subsequent derivatization to allow the determination by GC-MS of only PFOA with limit of detection (LOD) of 18000 ng/L and limit of quantification (LOQ) of 36000 ng/l, respectively. Although the analytical methods by GC-MS have been improved, reaching LOD and LOQ for PFOA equal to 1300 and 3000 ng/L, because of the need of derivatization and the applicability of the method only to perfluoro-carboxylates (PFOA, PFHxA, PFBA, etc.), gas chromatography has been replaced by liquid chromatography coupled to mass spectrometry. Already in 2001 it was possible to quantify (LOQ) 9 ng/L and 17 ng/L of PFOA and PFOS using a triple quadrupole mass spectrometer. The standard procedure for PFAS analysis since then has been preconcentration by SPE and subsequent determination by LC-MS triple quadrupole. The best results were obtained with instruments equipped with Qtrap technology, i.e. triple quadrupoles with a linear trap for ion enrichment, which provided pg/mg/litre LOD in 2004. In the first half

of the 2000s several studies reported that even with a simple single quadrupole mass spectrometer, it was possible to analyze tens ng/L in relatively clean matrices such as groundwater. In 2009, the US Environmental Protection Agency (EPA, 2008) published the first version of the standardized method 537 for the analysis of 14 PFAS in drinking water by preconcentration and LC-MS/MS analysis. The scientific publications between 2010 and 2020 reported further analytical improvements expanding the application of the methodology to wastewater, sludge and landfill leachates.

The short-chain PFCAs (PFBA) in environmental matrices started to be analyzed effectively in the first half of the 2010s by new chromatographic columns with much higher performance than the previous ones. In fact, the former EPA 537 method did not include PFBA in the list of analyzed PFAS. Since 2012, there has been an increasing analytical interest for environmental matrices, which can be explained by both the greater accessibility of the analytical instrumentation, and the increased relevance of environmental issues related to PFAS.

As far as the sample cleaning, storage, and handling procedures, already in the studies of early 2000s the need of using PFAS free solvents and checking procedural blanks was recognized necessary to lower LOD and LOQ and to obtain satisfactory results.

Significance valuation related to the present case is the subject of a preliminary exploration effort: prudently the interpretation of phenomenologies is setting up from the aquifer's viewpoint. Some guidelines, derived from statistics, allow the identification of analysis criteria, such as the strength, the consistency, the specificity or the temporality of causal factors. The extent of the PFAS plume, which assumes longitudinal dimensions of more than 40 kilometers, and the amount of contaminant mass discharge required to achieve the experimental PFAS concentrations in the central Veneto aquifers, become primary indicators of geographic significance. These parameters delineate the hydrogeochemical magnitude of the observed findings. Likewise, the temporality of the emissive behaviours seems consistent with the transport mechanisms of pollutants within saturated porous systems and at the same time is coherent with the long-term diffusion scenarios available at the state of knowledge. Therefore the persistence of the pollutant source, active for several decades and with stationary characteristics due to specificity of local industrial production, introduces additional connotations of significance in terms of duration, increased by the finding of irreversibility of environmental effects. This consideration results from the ultimate impossibility of qualitative repair of the ex-ante use conditions of the groundwater asset, within a reasonable restoration time frame. Without entering the field of epidemiological specificity, which can relate the significative health effects on the population chronically exposed to "forever chemicals", a further socio-economic index of significance can be traced to the enormous loss of valuable water resource. Indeed, the environmental circumstances in question have severely compromised the natural quality of some of the most important groundwater production areas in Italy. The number and extent of corrective actions that Water Utilities had to deploy on an emergency basis to cope with the occurrences constitutes, albeit indirectly, further evidence of the existence of an environmental harm, caused by significant groundwater deterioration.

PFAS Contamination in Veneto Region is an extensive and massive event, with lots of consequences for the environment, the public health, the socio-economy, and the regulation. It represents a prominent case to demonstrate the practical application of the quadrupole helix approach involving, respectively, the policy and decision makers (at both local, regional and national level), the industry/business communities, the research/academia institutions, the media- and culture-based public and the civil society and, finally, the water utilities playing a cross-sectorial challenging role of being in charge of the overall management of the drinking water resources.

Estimating the early-life exposure to two perfluorinated compounds (PFOS and PFOA) using PBPK modeling and biomarker measurements

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Children exposure to per- and polyfluoroalkyl acids substances (PFAS) is a major concern in health risk assessment as early-life exposures are suspected to induce adverse effects such as cardiometabolic, neurodevelopmental, or impact on immune response. Our work aims to estimate children exposure to PFOA and PFOS using a Physiologically-Based Pharmacokinetic (PBPK) modelling approach.

A lifetime model for PFAS was updated to simulate the internal PFAS exposures during the *in utero* life and childhood, and to include individual characteristics and exposure scenarios. Our approach was applied on the HELIX cohort, involving 1,239 mother-child pairs with measured PFOA and PFOS plasma concentrations at two sampling times: during pregnancy and in childhood (6 to 10 y.o.).

Our model predicted an increase of the plasma concentrations during fetal development and childhood until 2 y.o when the maximum concentrations were reached. Higher PFOA plasma concentrations than PFOS ones were predicted until 2 y.o, and then PFOS levels gradually became higher than PFOA ones. From 2 to 8 y.o, mean concentrations decrease from 3.09 to 1.88 µg/L (PFOA) and from 4.77 to 3.56 µg/L (PFOS). The concentration-time profiles vary with the age and were mostly influenced by *in utero* exposure (0-4 months), breastfeeding (5 months – 2 (PFOA) or 5 (PFOS) y.o), and food intake (after 3 (PFOA) or 6 (PFOS) y.o). Similar measured biomarker levels can correspond to large differences in the simulated internal exposures, highlighting the importance to investigate the child exposure over the early life to improve exposure classification of the child. The simulated internal concentrations in several infants and toddlers exceeded the HBM-I value threshold compared to the ones in older children, especially for PFOA.

Our approach based on PBPK modelling and HBM data demonstrates the possibility to simulate individualized internal exposures that could be accounted to refine risk assessment in early life.

**PFAS and their impacts into human health:
Example with a study into exposure to PFAS and glycemic control in older US adults
with type 2 diabetes mellitus**

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Introduction

Diabetes is a chronic disease characterized by the presence of excess sugar in the blood. In 2021, it was 537 million in the world that were affected or one in 10 people. There are different types of diabetes, we will focus more specifically on type 2 diabetes mellitus (T2DM). An adult aged 60 years and older at diagnosis loses at least 5 years in life expectancy due to diabetes complications. Optimal glycemic control (targeted glycated hemoglobin or A1C levels <7%) has been recommended to prevent diabetes complications. Old age induces an alteration of glucose homeostasis, particularly using certain medications. But beyond treatments, environmental pollution plays an undeniable role in certain metabolic responses in humans. PFAS have been positively associated with impaired glucose homeostasis, incident kidney failure in T2DM, and inversely associated with coronary heart diseases in T2DM.

In this study, we aimed to examine associations of serum concentrations of PFAS with poor glycemic control in US adults aged ≥65 years with known T2DM.

Methods

We worked with data from the NHANES study. The NHANES (National Health and Nutrition Examination Survey) is a bi-annual survey which aims to evaluate the health and nutritional status of the American population. This study has been in existence since 1960 and has been an ongoing program since 1999. The study examines a sample of at least 5000 people each year. We identified 5 PFAS of interest: PFDeA, PFNA, PFOA, PFOS, PFHxS, which are increasingly present in our environment.

We used generalized linear mixed models (the GLIMMIX procedure in the SAS statistical package (version 9.4. SAS)). The PFAS distributions were categorized into quartiles, and we estimated the adjusted odds ratio (aOR) and 95% confidence intervals (95% CI) for the association of PFAS with poor glycemic control (defined as hemoglobin A1C ≥ 7%, (for aged <75 years) or ≥ 8%, (for aged ≥75 years).

Results

We included 4575 adults aged 65 years and older with known T2DM. The median in µg/L was 0.20 (0.14-0.40) for PFDeA, 0.74 (0.50-1.30) for PFNA, 3.00 (1.87- 4.30) for PFOA, 14.50 (7.70-24.90) for PFOS, and 1.70 (0.90-2.60) for PFHxS, and 5.80 (2.30-20.70) for ΣPFAS. Defining poor glycemic control as A1C ≥ 7.5% in those aged 65-74 years and ≥ 8% in those aged ≥ 75 years, the results were unchanged for PFDeA. However, poor glycemic control was more likely to be observed in adults in the fourth quartile of PFNA (aOR 2.30, 95% CI: 1.25-4.21) and PFHxS (aOR 2.87, 95% CI: 1.56 to 5.30), as well as in those in the second quartile of PFOS (aOR 2.81, 95% CI:

1.11-7.14), PFHxS (aOR 1.90, 95% CI: 1.09-3.32) and Σ PFAS (aOR 2.29, 95% CI: 1). Among men, those in the second quartile of PFHxS or the fourth quartile of PFDeA showed a higher probability of poor glycemic control. A lower probability of poor glycemic control was observed in those in the first PFOA quartile. PFOS and Σ PFAS were not associated with poor glycemic control. In women, the results were less consistent across sensitivity analyses.

Conclusion:

This is the first study to work on older populations and glycemic control in known diabetes. In addition, it has a large sample size and a heterogeneous population in terms of race/ethnicity. Our study suggests that, among U.S. adults aged \geq 65 years with known T2DM, poor glycemic control is more likely to be observed in those with elevated serum levels of NFP and PFHxS (regardless of sex) and PFDeA (in men), after controlling for age at interview, poverty, marital status, body mass index, smoking, hypertension, glomerular filtration rates, albumin/creatinine ratio, insulin therapy, NHANES cycle, and cohort effect.

Les PFAS et leurs impacts sur la santé humaine : Exemple avec une étude sur l'exposition aux PFAS et contrôle de la glycémie chez les adultes américains âgés atteints de diabète de type 2

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Introduction

Le diabète est une maladie chronique caractérisé par la présence d'un excès de sucre dans le sang. En 2021, c'était 537 millions de personnes dans le monde qui étaient touchées soit une personne sur 10 (Atlas 2021 de la International Diabetes Federation). Un adulte âgé de 60 ans et plus au moment du diagnostic perd au moins 5 ans d'espérance de vie en raison de complications du diabète. Un contrôle optimal de la glycémie (taux ciblés d'hémoglobine glyquée ou d'A1C < 7 %) a été recommandé pour prévenir les complications du diabète. La vieillesse induit une altération de l'homéostasie du glucose, notamment via la prise de certains médicaments. Mais au-delà des traitements, la pollution environnementale joue un rôle indéniable dans certaines réponses métaboliques chez l'humain. Les PFAS ont été positivement associés à l'altération de l'homéostasie du glucose, à l'insuffisance rénale incidente dans le Diabète de Type 2 (DT2) et inversement associés aux maladies coronariennes dans le DT2. Dans cette étude, nous avions pour objectif d'examiner l'association entre les concentrations sériques de PFAS et le contrôle glycémique chez les adultes américains âgés de \geq 65 ans vivant avec un DT2 connu.

Méthodologie

Nous avons travaillé avec les données de l'étude NHANES. L'étude NHANES pour *National Health and Nutrition Examination Survey* est une enquête bi-annuelle ayant pour but d'évaluer l'état de santé et le statut nutritionnel de la population américaine. Cette étude existe depuis 1960 et est devenue un programme continu à partir de 1999. L'étude examine un échantillon d'au moins 5000 personnes chaque année. Nous avons identifié 5 PFAS d'intérêt : le PFDeA, PFNA, PFOA, PFOS, PFHxS, qui sont de plus en plus présent dans notre environnement.

Nous avons choisi les modèles linéaires généraux mixtes (procédure GLIMMIX dans le logiciel SAS 9.4). Les distributions de PFAS ont été catégorisés en quartiles, et nous avons estimé le rapport de cotes ajusté (aOR) et les intervalles de confiance à 95 % (IC à 95 %) pour l'association de PFAS à un contrôle glycémique inadéquat (défini comme une hémoglobine A1C $\geq 7,0\%$ (pour ceux âgés <75 ans) or $\geq 8,0\%$ (pour ceux âgés ≥ 75 ans).

Résultats

Nous avons inclus 4575 adultes âgés à partir de 65 ans atteints de DT2 connu. La médiane en µg/L était de 0,20 (0,14–0,40) pour le PFDeA, 0,74 (0,50–1,30) pour le PFNA, 3,00 (1,87–4,30) pour le PFOA, 14,50 (7,70–24,90) pour le PFOS et 1,70 (0,90–2,60) pour le PFHxS, et 5,80 (2,30–20,70) pour le ΣPFAS. En définissant le mauvais contrôle glycémique comme étant A1C $\geq 7,5\%$ chez les 65–74 ans et $\geq 8\%$ chez les ≥ 75 ans, les résultats sont restés inchangés pour le PFDeA. Cependant, un mauvais contrôle glycémique était plus susceptible d'être observé chez les adultes du quatrième quartile de PFNA (aOR 2,30, IC à 95 % : 1,25 à 4,21) et de PFHxS (aOR 2,87, IC à 95 % : 1,56 à 5,30), ainsi que chez ceux dans le deuxième quartile de PFOS (aOR 2,81, IC 95 % : 1,11–7,14), PFHxS (aOR 1,90, IC 95 % : 1,09–3,32) et ΣPFAS (aOR 2,29, IC 95 % : 1). Chez les hommes, ceux du deuxième quartile de PFHxS ou du quatrième quartile de PFDeA ont montré une probabilité plus élevée de mauvais contrôle glycémique. Une probabilité plus faible de mauvais contrôle glycémique a été observée chez les personnes du premier quartile de l'APFO. Le PFOS et la ΣPFAS n'étaient pas associés à un mauvais contrôle glycémique. Chez les femmes, les résultats étaient moins cohérents d'une analyse de sensibilité à l'autre.

Conclusion et Perspective :

C'est la première étude à travailler sur les populations âgées et le contrôle glycémique dans des cas de diabète connu. De plus elle a une grande taille d'échantillon et une population hétérogène en termes de race/ethnicité. Notre étude suggère que, chez les adultes américains âgés de ≥ 65 ans atteints de DT2 connu, un mauvais contrôle glycémique est plus susceptible d'être observé chez ceux qui présentent des taux sériques élevés de PFNA et de PFHxS (indépendamment du sexe) et de PFDeA (chez les hommes), après contrôle de l'âge au moment de l'entretien, la pauvreté, l'état matrimonial, l'indice de masse corporelle, le tabagisme, l'hypertension, les taux de filtration glomérulaire, le rapport albumine/créatinine, l'insulinothérapie, le cycle NHANES et l'effet de cohorte.

Combination of chemical clustering, QSAR models and known experimental data into a New Approach Method (NAM) – application to the PFA's toxicology assessment

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Abstract

PFA's representing around 10 000 molecules, a specific individual toxicological investigation is not feasible. Therefore, we set up a NAM (New Alternative Method) combining chemical clustering, QSAR models and known experimental data (e.g Toxicological Reference Value (TRV)).

We applied this NAM on subset of representative PFAS published by the EPA containing 74 unique PFAs compounds prioritized for testing by researchers to generate data to inform read-across approaches [4].

Firstly, the subset molecules were grouped by chemical similarity using the proprietary spherical harmonic (SH) descriptors taking into account the shape and the physico-chemical properties of the molecules. Interestingly, the resulting clustering obtained pointed out several clusters characterized by "representative compounds" that could be further used as molecular templates to screen novel compounds with the aim of toxicological homogenous PFA's groups.

Second, we applied QSAR based methodologies to assess the toxicological profile of each compound on selected endpoints (carcinogenic, mutagenic, reprotoxic, ...) used for classification on REACH SVHC list.

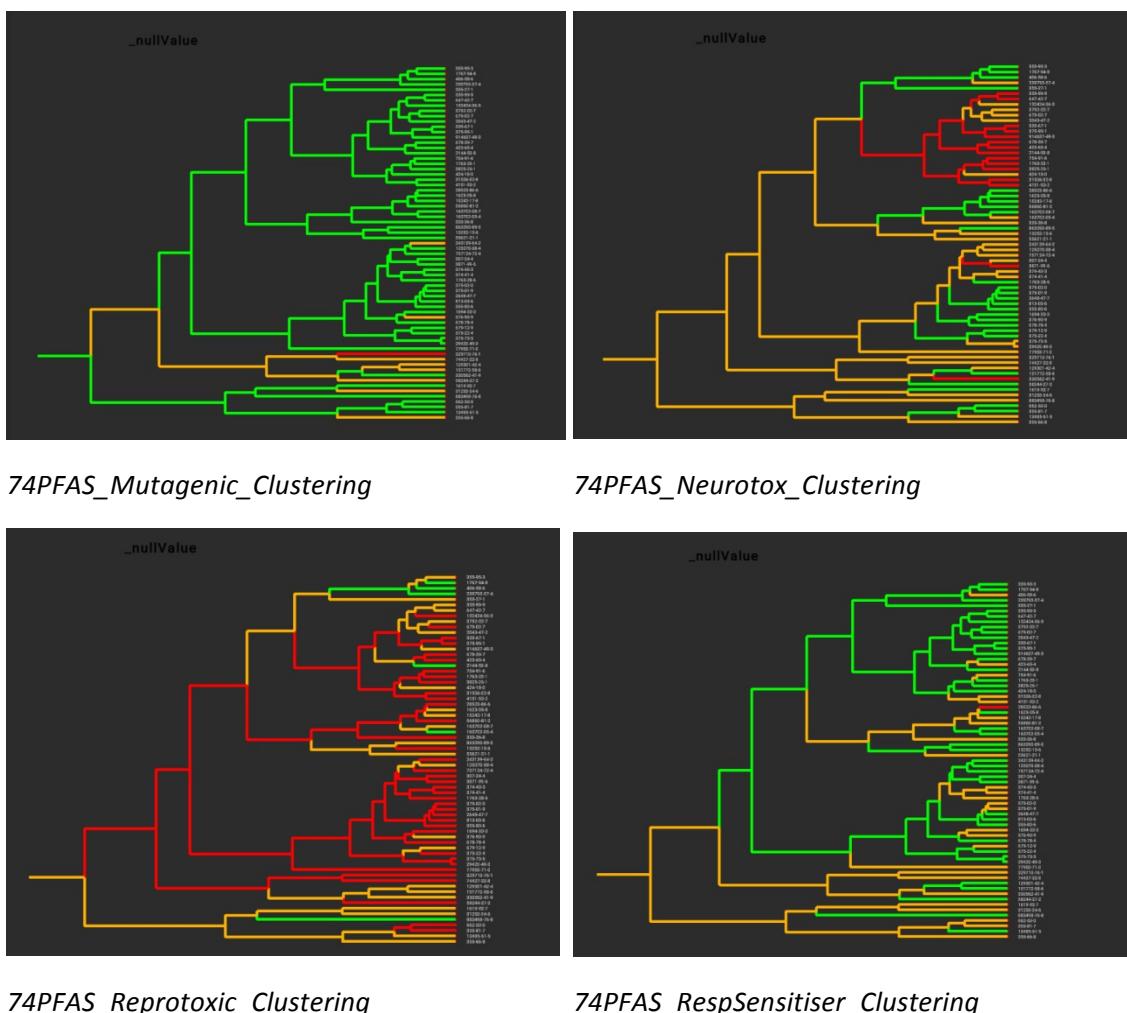
Finally, we mapped the known TRV value of 9 reference PFAS on the clustering. It appears that those reference compounds could be used for read-across approach in case of lack of experimental toxicology value.

Altogether, consistency and robustness of the QSARs combined with a clustering analysis support the suitability of this new approach methodology (NAM) to a prospective screening to estimate potential toxicity of PFA's groups and enable a read-across approach in case of lack of toxicological data.

The use of SAFETY BY DESIGN® software enables this new methodological approach.

This NAM is a part of SAFETY BY DESIGN®, the new solution of services and software for toxicity prediction and characterization of chemical substances.

Illustration



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Exposure to PFAS among Canadian Children and its Determinants

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Data from the Canadian Health Measures Survey (CHMS) indicates that the majority of Canadians have detectable blood levels of per- and polyfluoroalkyl substances (PFAS). Exposures to these toxic chemicals have been associated with many adverse health outcomes, including reduced antibody response to vaccination, increase in cholesterol, ulcerative colitis, and kidney and testicular cancer. Despite the fact that children are particularly vulnerable, there are limited data on children's exposure levels, and most studies focused mainly on sub-set of PFAS, such as perfluorooctanoate (PFOA) and perfluorooctane sulfonate (PFOS). Moreover, there is limited information on the main determinants of PFAS exposure, making it difficult to develop exposure mitigation strategies. The overall objectives of this study are to characterize PFAS exposure in Canadian children and identify the main determinants of children's plasma PFAS concentrations.

The CHMS is a cross-sectional population survey conducted over a decade. The survey includes comprehensive questionnaires as well as biospecimen collection which are used to generate biomonitoring data on environmental chemicals. The data used in this study combine the most recent CHMS cycles (cycle 5 and cycle 6 together) and include 981 children 3-5 years and 1033 children 6-11 years old. A total of 9 PFAS substances have been measured in plasma along with a rich database based on questionnaires, biochemical and physical measurements.

Among the 9 analysed substances, only 5 were detected (PFOS, PFOA, PFNA, PFHS, PFDA), PFOS and PFOA having the highest concentrations (GM±GMSD: 1.618±0.081 vs. 1.251±0.047 µg/L) and PFDA having the lowest concentration (0.123±0.006 µg/L). No differences were found for the age group (3-5 vs. 6-11 years old) and for the sex (males vs. females). Breastfeeding status (yes/no) and duration (>7 vs. <7 months) were significantly associated with highest levels of plasma PFAS. Also PFAS concentrations show a decreasing trend when BMI increase, and an increasing trend with increasing seafood consumption. Statistical analysis is still ongoing and will help to identify additional determinants and their relative contributions.

Ecotoxicité des mélanges de composés perfluorés sur la faune du sol et possibles solutions de traitement

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Abstract

Les composés per fluorés appartiennent à une grande famille de contaminants environnementaux apparus dans les années 50' et qui ont pris une place extrêmement importante dans les biens de consommation courants mais également dans les procédés industriels. Cette utilisation, notamment par le biais de mousses anti incendies, peut conduire à une contamination des sols et après une phase de lixiviation à une contamination des eaux souterraines, ce qui risque donc de générer une potentielle contamination des aires de captages et d'alimentation en eau potable.

Si la toxicité des PFAS chez l'Homme est désormais connue, on déplore un manque de connaissances sur l'écotoxicité des PFAS sur la faune du sol ainsi que sur les mécanismes responsables de ces effets délétères.

Cette présentation a donc pour objet premier de détailler l'impact des PFAS, seuls ou en mélange, sur la faune du sol à l'aide des résultats de l'étude IPANEMA. Cette étude d'écotoxicologie réalisée par VALGO en collaboration avec l'INRAE, l'Université Gustave Eiffel, l'Université de Bordeaux, et la société CASC4DE et financée par l'ADEME, a étudié l'impact des PFAS à l'échelle individuelle et sub individuelle sur des oligochètes représentatifs des sols européens (*Apporectodea caliginosa*). Pour ce faire, cette étude a exposé des vers juvéniles à des sols artificiels dopés avec des concentrations croissantes de PFAS, ainsi qu'à des sols prélevés sur un site contaminé. L'impact de ces polluants sur la survie, la croissance, la maturation sexuelle, ainsi que la stabilité du génome ont été évalués. Ces travaux ont notamment mis en évidence une écotoxicité des PFAS à de très faibles concentrations (dès 0,3 mg/kg de PFAS), avec un effet synergique dans les mélanges de perfluorés.

Dans un second temps, cette présentation va se concentrer sur les différentes méthodes adaptées aux traitement des sols contaminés par les PFAS à des concentrations induisant une écotoxicité ou un risque de lixiviation dans les eaux souterraines. Il sera notamment présenté le procédé breveté par VALGO de Bio Foam Fractionnation, visant à utiliser des amendements biomimétiques dans un procédé de Fractionnement de mousse afin d'optimiser le traitement des eaux usées, des eaux souterraines, mais surtout des eaux issues du lavage des sols contaminés. L'étude a montré que ce procédé augmentait de manière significative les taux d'abattements des principaux PFAS réglementés de quelques pourcents à 99 % pour le PFOS, le PFOA, le PFHxS, le PFHpS. Ce traitement a également montré une bonne efficacité pour des composés à chaînes plus longues comme le PFDA (94%) et le 6:2 FTAB (96%).

Evaluating the mobility and remediation of PFAS in contaminated soils

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Abstract

Poly- and perfluoroalkyl substances (PFAS), commonly referred to as “forever chemicals”, are emerging chemical risks and environmental crises. The existence of PFAS in soils around the globe indicates soil’s role as an important reservoir of PFAS contamination. Due to their high mobility and stability, PFAS present in soils can migrate to other environmental compartments (surface water, groundwater, air, plants) leading to their contamination. Despite the widespread existence of PFAS-contaminated soils, there exists limited data on the remediation of PFAS-contaminated soils. The present study has been performed to evaluate the mobility and soil washing-based remediation of PFAS in contaminated soils. For this, in the first step, the transport and sorption of a variety of PFAS (short-chained and long-chained compounds) were assessed in soil-packed columns under flow-through conditions. Obtained results indicate that the PFAS retention efficiency and adsorption mechanisms were dependent on the specific type of the PFAS and the nature of adsorbent. Tested PFAS showed limited affinity towards ferrihydrite, gibbsite, and manganese dioxide. However, the presence of carbonaceous materials in soils improved the retention of PFAS, thereby reducing their mobility. A newly developed reactive transport model accurately predicted the adsorption and transport of the tested PFAS in single and mixed contamination systems. Further experiments with soil washing have been conducted with water and/or green extracting agents in combination with innovative under various experimental setups (saturated vs. non-saturated flow-through conditions) to remove PFAS at various contamination levels. Soil washing owing to its simplicity and cost-efficiency has the potential to become a viable remediation strategy for PFAS-contaminated soils. The outcomes of this study have important implications to assess the fate and remediation of PFAS in soils.

Keywords: PFAS; Soil remediation; Adsorption; Mobility; Soil washing;



Addressing PFAS challenges in water: Veolia return of experience in North America and Australia

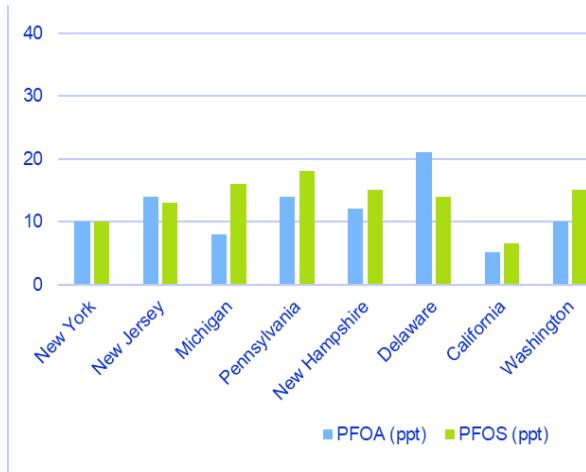
Relever les défis des PFAS dans l'eau : Retour d'expérience de Veolia en Amérique du Nord et en Australie

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The health dangers from per- and polyfluoroalkyl substances (PFAS)—a class of several thousand chemicals with a seven-decade legacy of military, industrial, and commercial use nationwide—have risen dramatically in public consciousness over the past few years. The vast extent of PFAS contamination throughout the environment has propelled legislators and the general public alike to crack down on the spread of these chemicals.

PFAS are gaining priority as a global environmental concern, with policy proposals in the US, in Australia and in Europe seeking to restrict the usage of these chemicals and implementing their own standards for PFAS in drinking water, groundwater, recreational water, biosolids, soil and wastes. In the US guidelines limit PFAS content in drinking water (<https://www.epa.gov>) and several states have moved forward with very stringent regulations. The following graph presents current and proposed PFOA & PFOS State Regulations.



Whereas in Australia, guidelines limit PFAS levels in drinking water and recreational water. Moreover, territories favour reduction at the source instead of treating produced drinking water.

In both of these countries, Veolia has multiple operations and had to adapt quickly and develop action plans to stay ahead of changing regulations.

In the US more than 30 PFAS drinking water treatments are currently at various stages of implementation. This paper will present various returns of experience on treatment optimization: treatment line (activated carbon, ion exchange resins, membranes), performance, CAPEX and OPEX,

residue management. For instance in the city of Rahway, NJ, the 4.85MGD (920 m³/h) surface water treatment plant, had PFOS and PFOA concentrations exceeding the state limits. The selected treatment option was to optimize the existing GAC treatment process by replacing the media with a more effective GAC product. This resulted in a 65% decrease in PFOA and PFOS and the plant is today compliant with NJ state limits (PFOA <14 ppt, PFOS<13 ppt, PFNA<13 ppt).

In Australia, there are numerous remote sites with contaminated water containing hydrocarbons, metals, suspended solids or PFAS. Contaminated water may be generated in emergency spill situations such as overflowing storage ponds or ongoing management of contaminated sites. To meet this challenge Veolia has implemented sustainable solutions for clients, with the development of a tailored onsite mobile water treatment plant.

Today, as more countries move forward with PFAS regulation, Veolia's return of experience in North America and Australia can help other entities implement their PFAS management strategies. As PFAS pose a unique challenge in terms of treatment it is key to identify the best technologies and also to take into account residues management in our strategy. A water and waste approach is essential.

DEVELOPING A SOCIETAL SUPPORTED REMEDIAL PLAN FOR AN IMPORTANT PFAS HOTSPOT IN BELGIUM: CHALLENGES FACED AND LESSONS LEARNED TO DATE?

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Introduction and Study Objectives

According to the European Chemicals Agency (ECHA), Per- and polyfluoroalkyl substances (PFAS) are a large class of thousands of synthetic chemicals that are used throughout society. They all contain carbon-fluorine bonds, which are one of the strongest chemical bonds in organic chemistry, and some are linked to negative effects on human health such as liver damage, thyroid disease, obesity, fertility issues and cancer. Given their unique physical and chemical properties, PFAS are highly stable and resistant to degradation when used and also in the environment. As a result, PFAS are being increasingly detected as environmental pollutants, not only close to the source of their release but also covering long distances away from its source.

With this case study, we take a critical look back in the rearview mirror at one of the most mediatised PFAS projects in Belgium. From the start it was important that the necessary priorities were correctly set, and a phased approach evaluating the potential human risk for PFAS exposure via the soil in the immediate vicinity of the PFAS manufacturing site was chosen as the first phase. This was not an obvious task, because both the policy and the reference framework were (and still are) evolutionary.

Materials and Methods

An initial descriptive soil investigation was carried out in a very short timeframe (<6 months) and included an area of >7 km². In addition to soil and groundwater samples, food-related samples (vegetables, fruit, milk, eggs, ...) were also analyzed. Additionally, a PFAS biomonitoring campaign (blood serum) performed in the immediate vicinity of the plant had to be integrated. Following the descriptive investigation and the PFAS biomonitoring campaign, a risk assessment was performed to assess the potential human risk for PFAS exposure at the Site and in its immediate vicinity.

Results and Discussion

Additional investigations are continuing, but the field data collected to date confirms that the PFAS problem is far more complex and widespread than initially anticipated. Other external sources have been identified, and the risk assessment indicates that there may be a human risk, with the risk being largely related to the consumption of free-range chicken eggs.

Only through transparent consultation it is possible to create the necessary societal support for the proposed remediation approach. In consultation with all the stakeholders involved (the environmental and public health agencies, the involved municipality, the NGOs, and the residents/farmers involved), the first phase remediation plan will address the identified human risk by means of a soil exchange remediation. The proposed remedial strategy has received approval by all involved stakeholders and is currently in the process of being organized.

PFAS sorption mechanisms – principles and issues

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Defining migration risks is an integral part of risk assessments. The general principle is that the solid-liquid distribution coefficient K_d is used to predict the migration of organic contaminants through groundwater. Another general principle is that K_d can be derived from K_{oc} , the organic carbon water distribution coefficient. In 2022, the Flemish soil authority OVAM defined threshold values for PFOS and PFOA in groundwater, based on their K_{oc} values.

In this presentation, we argue that calculating K_d from K_{oc} values for individual PFAS compounds is not correct. Furthermore, it is discussed that migration of PFAS in groundwater is influenced by many more factors than sorption.

The SFSE management guideline for PFAS : Contamination sources, Environmental Behavior, Investigations, Risk Assessments and Treatments

Le Guide de Gestion SFSE pour les PFAS : Sources de contamination, comportement environnemental, diagnostics de pollution, évaluations des risques et traitements

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1. Contexte

La SFSE : Société Francophone de Santé et Environnement (Belgique, Canada, France, Luxembourg, Suisse, Pays Africains, etc.), créée en France en 2008, se veut le lieu où se rencontrent savoirs, savoir-faire et faire-savoir sur le thème de la Santé Environnementale. A la croisée de domaines comme la Toxicologie, l'Epidémiologie, l'Expologie, la Chimie de l'Environnement, la Médecine, les Sciences Humaines et Sociales, la Réglementation et la Gestion, la Santé Environnementale ne peut être assurée sans une Approche Globale.

Les Per et Poly Fluoro Alkyl Substances (PFAS) présentent des propriétés de résistance à haute température, de stabilité chimique et de pouvoir imperméabilisant qui sont des atouts pour les industries et s'expliquent par la liaison carbone-fluor forte et stable. Les PFAS sont utilisés dans les mousses anti-incendie mais se retrouvent également dans les produits du quotidien : papier plastifié, emballages alimentaires (films anti-graissé), vêtements et textiles divers imperméabilisés (contre la pluie ou la saleté), équipements sportifs, tensioactifs (permet l'émulsion de liquides organiques à l'eau), outils de photolithographie (revêtement anti-reflet), semi-conducteurs (protection de surface), vernis, peintures, adhésifs, matériel médical, produits de beauté et d'hygiène (shampooing, crème solaire). L'ensemble de la population utilise ces produits au quotidien et est exposé aux PFAS, ce qui devient un enjeu de santé publique. La production de PFAS a commencé dans les années 1940, principalement par 8 industries internationales (par ordre alphabétique) : Arkema, Asahi, BASF Corporation, Clariant, Daikin, DuPont, 3M/Dyneon et Solvay/Solexis (ITRC, 2022). Jusqu'en 2000, la société américaine 3M était le plus grand fabricant mondial.

Au début, DuPont a inventé les produits chimiques PFAS (Poly & Perfluoro Alkyl Substances) sous forme de polymères de téflon, mais 3M est devenu son producteur le plus

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important. En 2001, un premier scandale de pollution aux PFAS éclate à Parkersburg /USA, après la découverte de monomères PFAS dans l'eau potable de dizaines de milliers de personnes à proximité d'une usine DuPont. Dès 2004, des PFAS ont été mesurés dans l'environnement en Scandinavie (Kärrman A., 2019). Depuis, les pollutions environnementales liées aux PFAS et révélées au grand public en Europe (en Allemagne en 2014, en Belgique en 2021, en France et en Italie en 2022. Au début 2023 des milliers de sites contaminés par les PFAS ont été publiés en Europe, comme dans France, Allemagne, Belgique, etc. (cf. Fig.1 - 3).

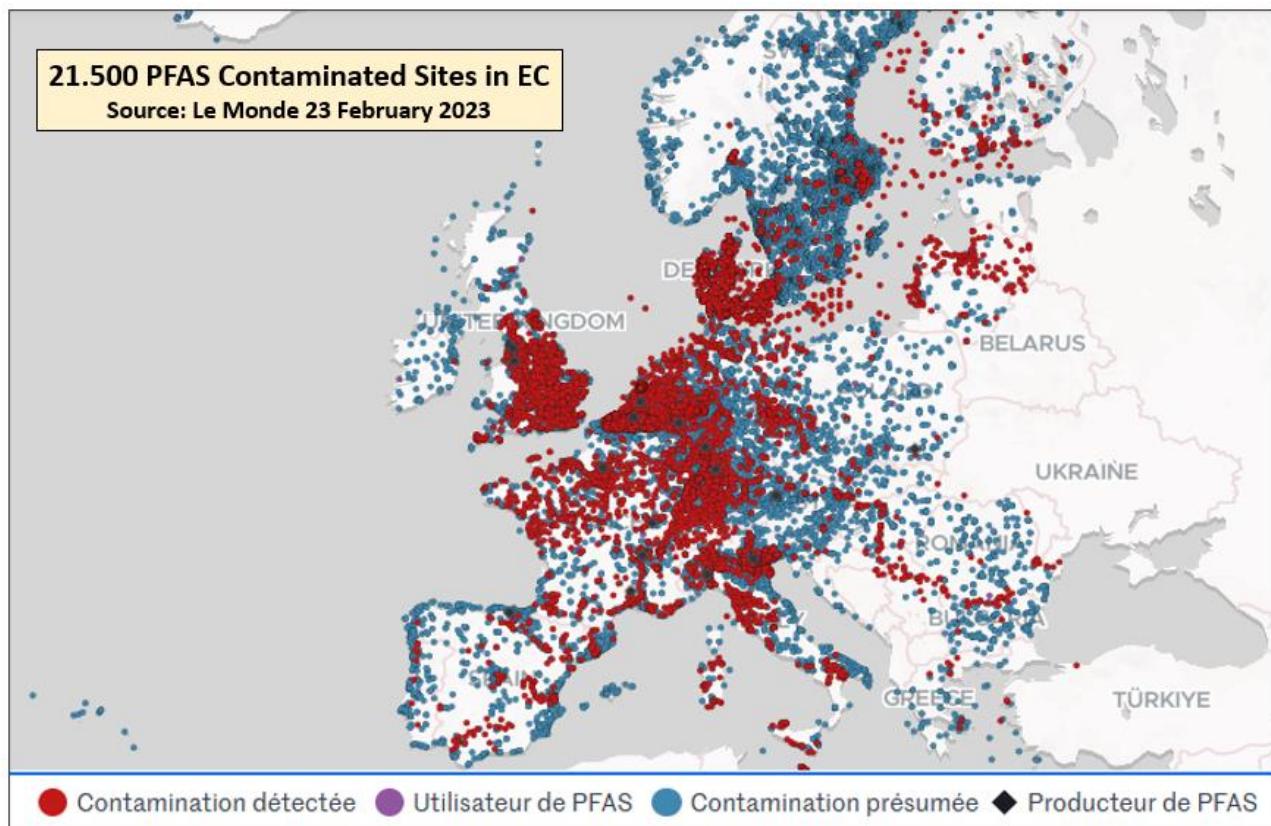


Fig. 1 : Au total, 21 500 sites contaminés par des PFAS (potentiels) dans la CE sont comptabilisés (Source : « Forever Pollution Projet », Le Monde 23 février 2023)

Tous ces événements ont conduit à une prise de conscience de l'impact de cette pollution sur l'Environnement et sur la Santé Publique. Des milliers d'ouvrages et de publications scientifiques existent sur la toxicologie, le comportement environnemental, l'épidémiologie, etc. des PFAS. De nombreuses publications scientifiques traitent des PFAS mais il subsiste des incertitudes majeures dues à des données éparses sur leur comportement ou leur toxicité, comme on le sait début 2023 sur un large nombre de 9 000 à 12 000 composés PFAS individuels existants. Dans ce contexte d'incertitudes et de besoins de recherche persistant et la gestion d'une crise environnementale et de santé publique liée aux PFAS est complexe et ce, pour les prochaines années.

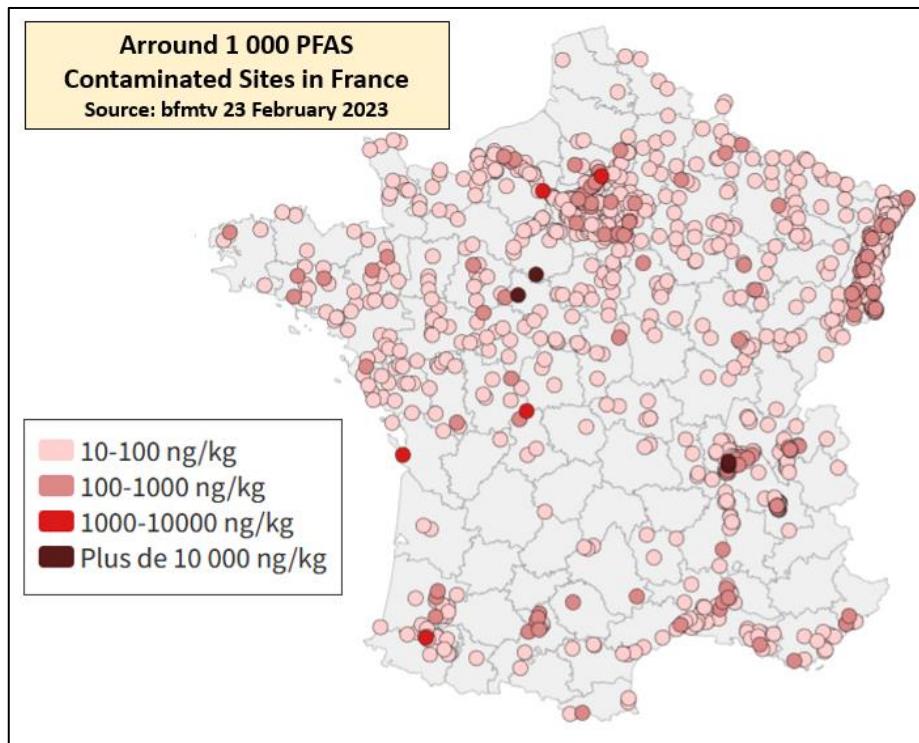


Fig. 2 : Environ 1 000 Sites contaminés par des PFAS en France sont comptabilisés (Source : bfmtv 23 février 2023)

En conséquence, les PFAS sont présents dans l'environnement, par ordre d'importance (ITRC, 2022, Merly C., 2020) :

- sur les zones d'entraînement à la lutte contre l'incendie (aéroports, bases militaires, sites d'entraînement à la lutte contre l'incendie, raffineries de pétrole) où sont utilisées des mousses anti-incendie (aussi appelées AFFF pour Aqueous Film-Forming Foam) contenant des PFAS ;
- dans les usines où ils sont produits ;
- dans les procédés industriels où ils sont utilisés : textile et cuir, papier/carton, métallisation et galvanisation (action anti-corrosion, utilisée pour limiter l'exposition des travailleurs aux aérosols de chrome VI), fabrication de câbles, fabrication de tensio-actifs, photolithographie, les semi-conducteurs, l'aéronautique (fluides hydrauliques) ainsi que les fabricants de mousses anti-incendie ;
- dans les stations d'épuration et les boues de stations d'épuration utilisées comme engrais sur les terres agricoles ;
- dans les installations de stockage de déchets ou les décharges et dans leurs lixiviats.

Le cas échéant, ajouter les PFAS à la liste des polluants à analyser dans les échantillons environnementaux comme le sol ou les eaux souterraines, l'eau potable ou les denrées alimentaires est un réflexe à adopter après la revue historique des activités, sans oublier les précurseurs (PFAS Polyfluorés).

En Europe, les sites de production de PFAS suivants ont indiqué une forte contamination pour la population locale : en France (Arkema à Lyon), Belgique (3M à Zwijndrecht), Italie (Solvay à Spinetta Marengo), Allemagne (Flughafen Düsseldorf, Site OTAN Bitburg, large site d'agriculture Rastadt) etc.

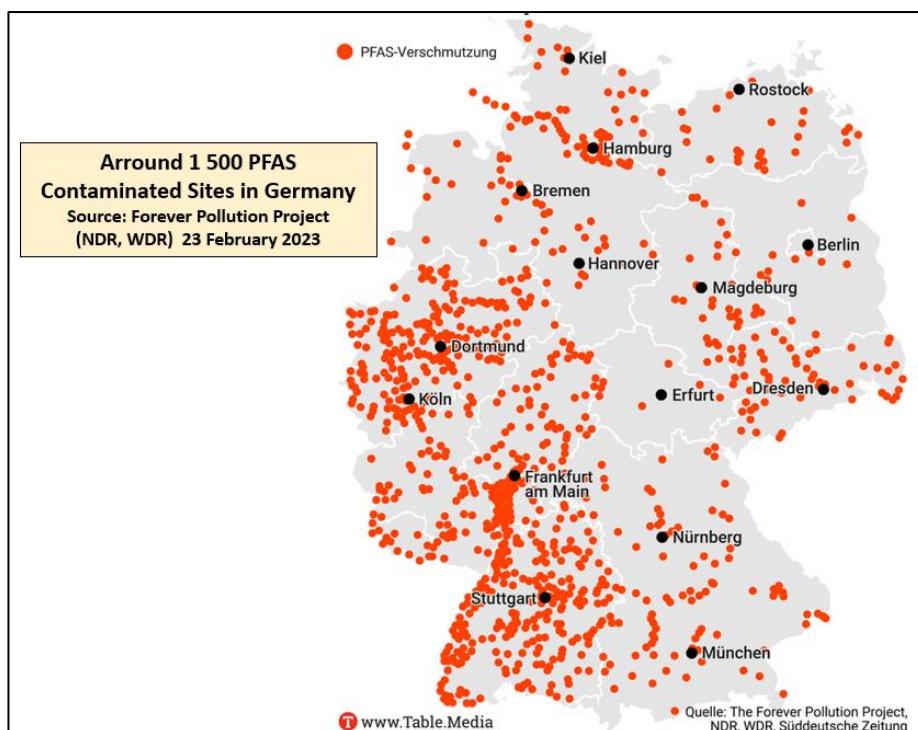


Fig. 3: Around 1 500 PFAS contaminated Sites in Germany are accounted (Source: Forever Pollution Project (NDR, WDR) 23 February 2023)

2. Chimie Environnementale : PFOS & PFOA, l'arbre qui cache la forêt :

Les PFAS sont des composés contenant une chaîne d'atomes hydrophobes de carbone et de fluor [$C_nF_{2n+1}-R$] avec une fonction hydrophile à son extrémité, comme un sulfonate (SO_3^-) ou un carboxylate (CO_3^{2-}). Ils peuvent être présents sous des formes ionisées et non ionisées. L'PFOA (acide perfluoro-octanonique) et le PFOS (acide sulfonique perfluoro-octanonique) cancérogènes sont classés comme POP (polluants organiques persistants) interdits par la convention de Stockholm. En raison de leur forte solubilité dans l'eau et de leur faible volatilité, la plupart des PFAS sont mesurés principalement dans les eaux souterraines (chaînes plutôt longues) et les eaux de surface (chaînes plutôt courtes). Dans l'environnement (pH compris entre 5 et 9), les PFAS sont normalement présents sous forme anionique, expliquant ainsi leur faible sorption sur les sols et les sédiments (qui sont chargés négativement) et leur forte solubilité dans l'eau. Certains PFAS, comme par ex. Les sulfo-bétaïnes telles que 6:2-FT(S)AB (= Capstone B) ou 6:2-FTS (= H4-PFOS), sont

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beaucoup plus solubles même dans des conditions légèrement alcalines (F. Karg 2022). Un point très important est que les PFAS polyfluorés (appelés précurseurs) sont biotransformés en PFAS perfluorés stables. Une biodégradation totale des PFAS n'a pas été prouvée aujourd'hui.

En raison de leur caractère hydrophile et hydrophobe, les PFAS ont tendance à s'accumuler à l'interface eau/air pour les eaux de surface. Leur tendance à s'accumuler au sommet d'une nappe phréatique est un sujet de recherche actuel. Avec un potentiel de dispersion en équivalent eau souterraine voire supérieur à celui du benzène et du trichloroéthane, les PFAS sont mesurés sur de longues distances, même au pôle Nord. Ainsi, la stratégie d'échantillonnage devrait inclure des points d'échantillonnage loin en aval de la source. Des données sur la pression de vapeur saturante (très différentes selon la forme ionisée ou non ionisée), la constante de Henry (H) et le coefficient de partage carbone organique/eau (Koc) disponibles dans la littérature. Certains PFAS, comme les alcools fluorotélomères (FTOH) sont volatils. Les données physico-chimiques, par ex. principalement pour le PFOS et le PFOA, sont très variables et n'existent souvent que pour la forme acide (et non pour les sels).

Afin de mieux estimer le comportement des PFAS, des données relatives à la forme appropriée doivent être sélectionnées et une mesure sur site du coefficient d'adsorption basée sur la teneur en carbone organique du sol (Koc) est recommandée. Des précautions particulières doivent être prises lors de la collecte des PFAS, en évitant l'utilisation de matériaux contenant du "fluor" comme le téflon, etc. pour éviter les résultats d'analyse faussement positifs. Les laboratoires d'analyse rencontrent des difficultés pour analyser les PFAS : seuls une quarantaine de composés peuvent être quantifiés normalement. Les meilleurs laboratoires ont pu analyser environ 500 composés PFAS, avec des limites de quantification parfois supérieures aux valeurs réglementaires et des suspicions d'interférences, augmentant les difficultés d'interprétation. Pour compenser ces incertitudes, il est recommandé d'augmenter le nombre d'analyses à blanc et répétées.

3. Toxicologie et Evaluation des Risques Sanitaires :

Comme certains PFAS se décomposent en acides perfluorés qui sont plus toxiques et ne se dégradent pas, les précurseurs doivent également être analysés. L'évaluation des risques des PFAS pour la santé humaine et l'environnement est complexe :environ 40 composés (uniquement) bien connus pour leurs effets sur la santé humaine et pour lesquels les connaissances évoluent. L'évaluation des risques ne peut pas considérer l'effet de mélange de ces 9 000 – 12 000 composés car il n'existe pas 9 000 Valeurs Toxicologiques de Référence (VTR). Des données de toxicité humaine sont principalement disponibles pour : PFOS, PFOA mais aussi pour PFHxA, PFDA et PFNA.

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Les principaux effets toxicologiques sont résumés dans le tableau suivant (US-EPA : 2021) :

	# of Carbons	Liver	Developmental	Reproductive	Immune	Hematologic	Thyroid	Neuro-behavioral	Tumors
Perfluoroalkyl Carboxylates									
PFBA	4	■	■	■	□	■	■	□	□
PFPeA	5	□	□	□	□	□	□	□	□
PFHxA	6	■	■	■	□	■	■	□	□ (Negative)
PFHpA	7	■	□	□	□	□	□	□	□
PFOA	8	■	■	■	■	■	■	■	■
PFNA	9	■	■	■	■	■	■	□	□
PFDA	10	■	■	■	■	■	■	■	□
PFUnA	11	■	■	□	■	□	□	□	□
PFDoA	12	■	■	■	■	■	□	■	□
Perfluoroalkyl Sulfonates									
PFBS	4	■	■	■	■	■	■	□	□
PFHxS	6	■	■	□	□	■	■	■	□
PFOS	8	■	■	■	■	■	■	■	■
Per- & Polyfluoroalkyl Ether Replacements									
ADONA	6	■	■	□	□	■	□	□	□
HFO-DA		■	■	■	■	■	■	□	
GenX	6	■	■	■	■	■	■	□	■

■ Effect reported in one or more laboratory animal study
 □ Effect was evaluated but not found, or effect has not been evaluated

Les principales voies d'exposition sont l'ingestion d'eau potable contaminée (eau du robinet) ou d'aliments contaminés. En 2020, l'EFSA a évalué l'exposition de la population européenne aux PFAS via les aliments (EFSA, 2020). Sur la base de 67 839 échantillons analysés dans 16 pays (dont 33 000 analysés chez les poissons), la principale exposition aux PFAS provient de la consommation de poisson (notamment carpe, anguille, gardon, perche et sardine) suivie des fruits et des œufs.

Les VTR évoluent chaque année avec une diminution drastique de la dose maximale admissible pour le PFOS et l'APFO, divisée par un facteur 10 en 10 ans. Concernant l'effet de mélange, le RIVM (RIVM, 2018) propose d'utiliser des Relative Potency Factors (RPF), selon W. Bil (2021) pour 12 composés, basés sur les effets de toxicité hépatique. L'EFSA (EFSA, 2020) propose une autre approche « indicateur d'exposition ». Malgré la volonté d'améliorer la situation sanitaire en produisant des PFAS à chaîne plus courte, les résultats des études toxicologiques ne permettent pas encore de se rassurer sur la réduction des effets toxiques de ces alternatives.

4. Le Guide et Fiches de la SFSE de Management des PFAS:

Afin d'aider les collectivités, industries, bureaux d'études, etc. concernés par cette problématique de gestion de la pollution des PFAS, la SFSE a décidé de synthétiser les informations disponibles et d'élaborer un Guide pratique avec des Fiches Thématiques pour les mettre à disposition du plus grand nombre possible, de manière pragmatique, comme par exemple :

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1. Fiche 1 : Note chapeau : Connaissances générales; identité et chimie des PFAS,
2. Fiche 2 : Cadre réglementaire & juridique des PFAS,
3. Fiche 3 : Sources de contamination des PFAS,
4. Fiche 4 : Devenir dans l'environnement et biotransformation des PFAS,
5. Fiche 5 : Méthodes de prélèvements (sols, eaux souterraines, eaux de surface, gaz du sol, air ambiant, Human Bio-Monitoring),
6. Fiche 6 : Méthodes d'analyses (sols/eaux souterraines/eaux de surface et autres),
7. Fiche 7 : Bruit de fond et recommandations pour la détermination du bruit de fond (base de données),
8. Fiche 8 : Comportement dans l'environnement – Paramètres physico-chimiques (base des données),
- 9a. Fiche 9a : Toxicité - Valeurs Toxicologiques de Référence (VTR),
- 9b. Fiche 9b : Evaluation des Risques Sanitaires (ERS),
10. Fiche 10 : Considération de l'effet toxique de mélange des PFAS,
11. Fiche 11 : Expositions et Données de bio-monitoring humain (HBM),
12. Fiche 12 : Méthodes d'assainissement - dépollution des sols,
13. Fiche 13 : Méthodes d'assainissement - dépollution des eaux souterraines,
14. Fiche 14 : Méthodes d'assainissement des gaz du sol et de l'air ambiant.

Ces fiches d'information PFAS de la SFSE seront régulièrement mises à jour et pourront être consultées sur <https://www.sfse.org/accueil>

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The SFSE management guideline for PFAS : Contamination sources, Environmental Behavior, Investigations, Risk Assessments and Treatments

Le Guide de Gestion SFSE pour les PFAS : Sources de contamination, comportement environnemental, diagnostics de pollution, évaluations des risques et traitements

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1. Context

The SFSE: Société Francophone de Santé et Environnement (Association of Health and Environment of the French speaking Hemisphere as in Belgium, Canada, France, Luxembourg, Switzerland, African Countries, etc.), created in France in 2008, aims to be the place where knowledge, know-how and let-know can be met on the topic of Environmental Health. At the crossroads of fields as Toxicology, Epidemiology, Expology, Environmental Chemistry, Medicine, Human and Social Sciences, Regulations and Management, Environmental Health cannot be ensured without a Global Approach.

The PFAS show properties of high temperature resistance, chemical stability and waterproofing power are assets for industries and are explained by the strong and stable carbon-fluorine bond. PFAS are used in fire-fighting foams but are also found in everyday products: plasticized paper, food packaging (grease resistant films), various waterproofed clothes and textiles (against rain or dirt), sports equipment, surfactants (allows the emulsion of organic liquids with water), photolithography tools (anti- reflection coating), semiconductors (surface protection), varnishes, paints, adhesives, medical equipment, beauty and hygiene products (shampoo, sun cream). The whole population uses these products on a daily basis and is exposed to PFAS, which is becoming a public health issue.

PFAS production started in the 1940s, mainly by 8 international industries (in alphabetical order): Arkema, Asahi, BASF Corporation, Clariant, Daikin, DuPont, 3M/Dyneon and Solvay/Solexis (ITRC, 2022). Until 2000, the US Company 3 M was the world's largest manufacturer. At the beginning DuPont invented the PFAS chemicals (Poly & Perfluoro Alkyl Substances) in form of Teflon polymers, but 3M became its most important producer. In 2001, a first PFAS pollution scandal

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occurred in Parkersburg /USA, after discovery of PFAS monomers in drinking water of tens of thousands of people near a DuPont plant. As early as 2004, Per and Poly Fluoro Alkyl Substances (PFAS) were measured in the environment in Scandinavia (Kärrman A., 2019). Since then, environmental pollution linked to PFAS and revealed to the general public in Europe (in Germany in 2014, in Belgium in 2021, in France and in Italy in 2022. In beginning 2023 thousands of FAS contaminated sites were published in Europe, as in France, Germany, Belgium, etc. (cf. Fig.1 - 3).

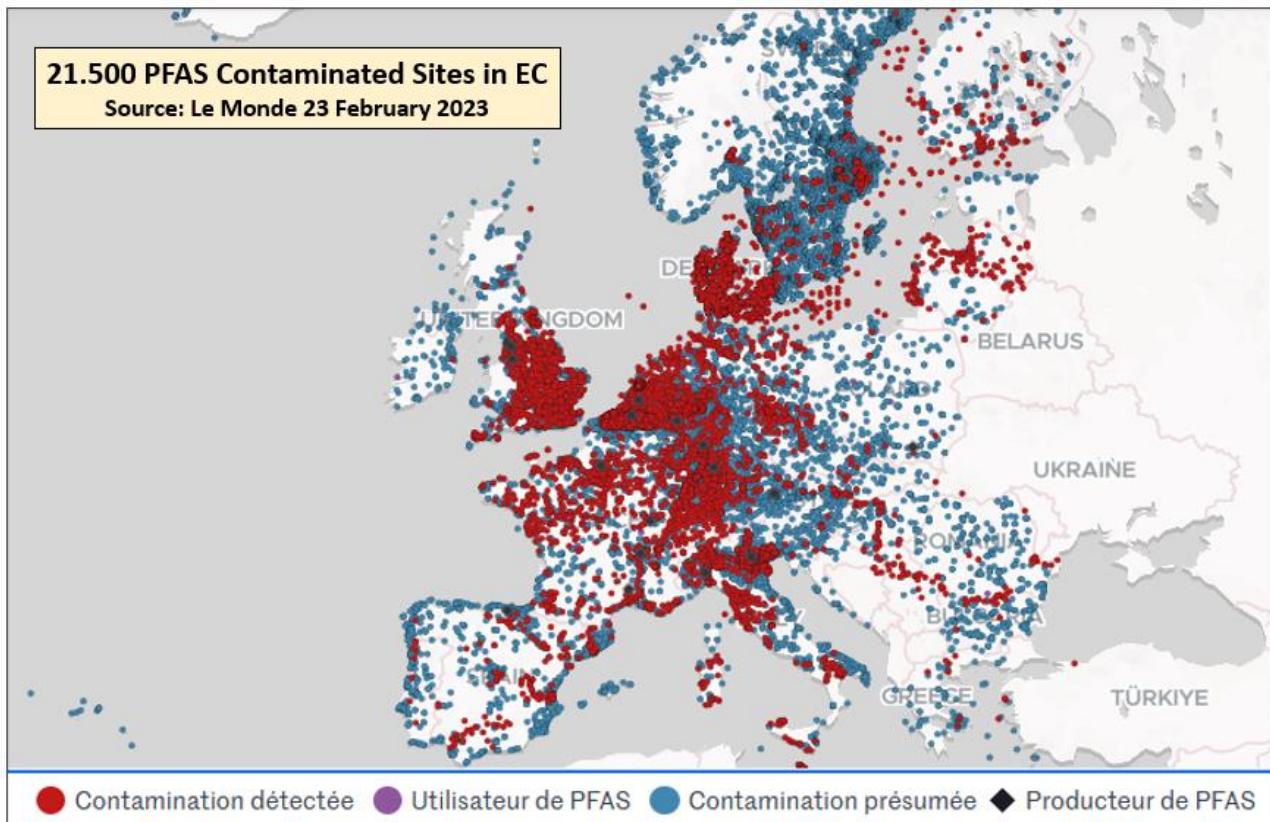


Fig. 1: In total 21.500 PFAS (potential) contaminated Sites in EC are accounted (Source: « Forever Pollution Project », Le Monde 23 February 2023)

All these events have led to an awareness of the impact of this pollution in Environment and impact to Public Health. Thousands of scientific works and publications on PFAS' toxicology, environmental behavior, epidemiology etc. Many scientific publications deal with PFAS but there are still major uncertainties due to scattered data on their behavior or toxicity, as we know in beginning 2023 about a large number of 9 000 – 12 000 existing individual PFAS compounds. In this context of uncertainties and research needs, the Management of an Environmental and Public Health crisis related to PFAS will be complex for the next years.

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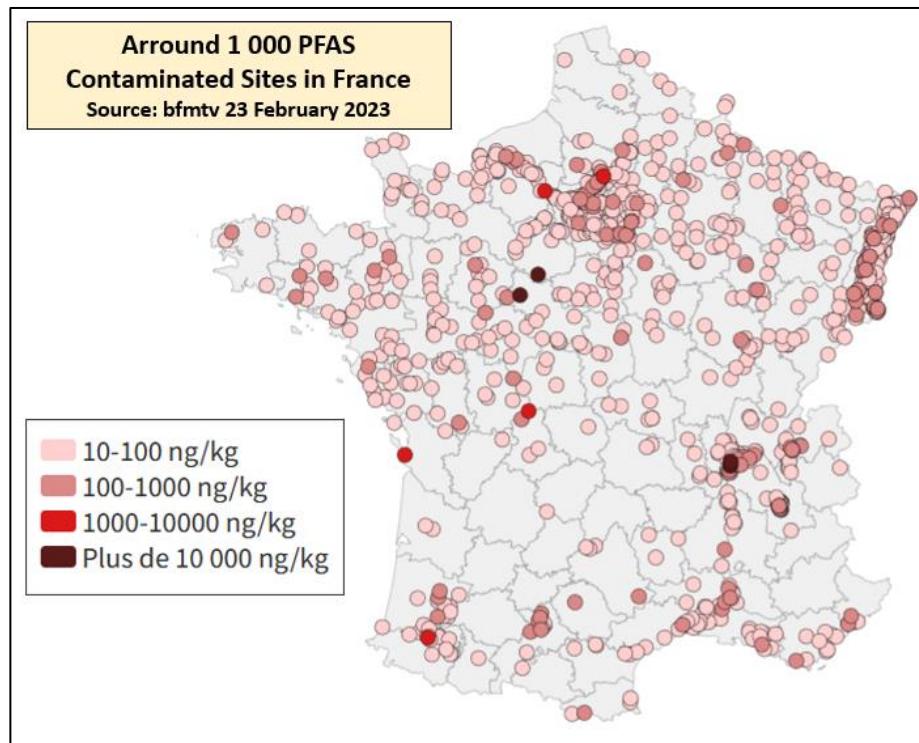


Fig. 2: Around 1 000 PFAS contaminated Sites in France are accounted (Source: bfmtv 23 February 2023)

As a result, PFAS are expected to be present in the environment, in order of importance (ITRC, 2022, Merly C., 2020):

- on fire-fighting training areas (airports, military bases, fire-fighting training sites, oil refineries) where fire-fighting foams (also called AFFF for Aqueous Film-Forming Foam) containing PFAS are used;
- in the factories where they are produced;
- in the industrial processes in which they are used: textiles and leather, paper/cardboard, metallization and galvanization (anti-corrosion action, used to limit workers' exposure to chromium VI aerosols), cable manufacturing, surfactant manufacturing, photolitho-graphy, semiconductors, aeronautics (hydraulic fluids) as well as fire-fighting foam manufacturers ;
- in waste water treatment plants and WWTP sludge used as fertilizer on agricultural land;
- in waste storage facilities or landfills and in their leachate.

Where appropriate, adding PFAS to the list of pollutants to be analyzed in Environmental Samples as soil or groundwater, drinking water or foodstuff is a reflex to be adopted after the historical review of activities, without forgetting the precursors (Poly fluorinated PFAS).

In Europe, the following PFAS production sites indicated high contamination for the local

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population: in France (Arkema in Lyon), Belgium (3M in Zwijndrecht), Italy (Solvay in Spinetta Marengo), Germany (Flughafen Düsseldorf, NATO Site Bitburg, Agriculture site Rastadt) etc. .

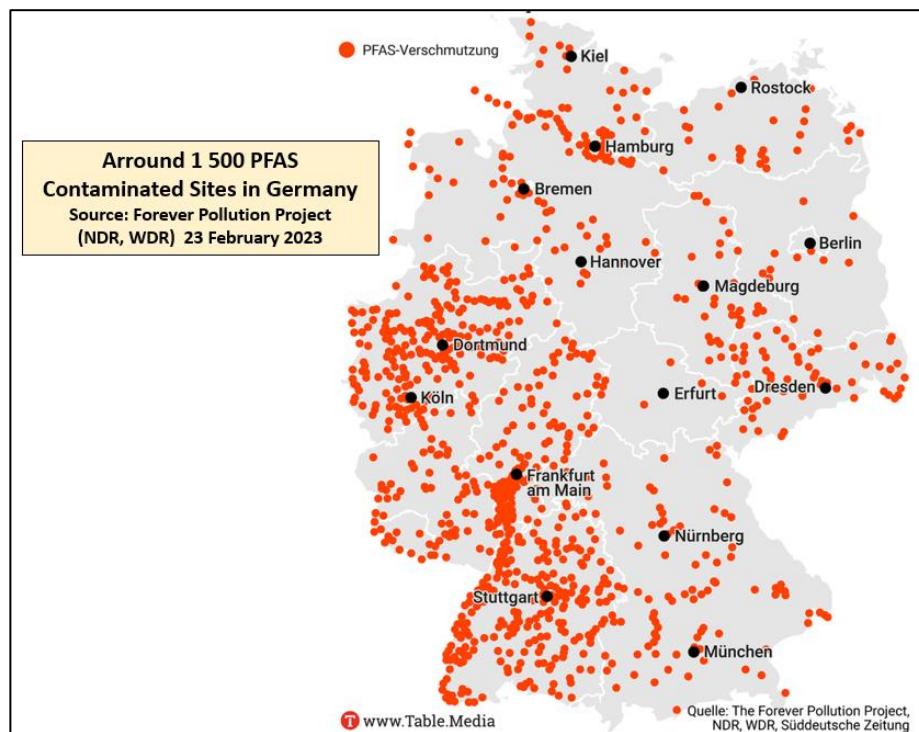


Fig. 3: Around 1 500 PFAS contaminated Sites in Germany are accounted (Source: Forever Pollution Project (NDR, WDR) 23 February 2023)

2. Environmental chemistry: PFOS and PFOA, the tree that hides the forest:

PFAS are compounds containing a chain of hydrophobic carbon and fluorine atoms $[C_nF_{2n+1}-R]$ with a hydrophilic functional group at its end, such as a sulphonate (SO_3^-) or carboxylate (CO_3^{2-}). They can be present in both ionised and non-ionised forms. Carcinogenic PFOA (Perfluoro-Octanonic-Acid) and PFOS (Perfluoro-Octanonic-Sulfone) are classified as banned POP (Persistent Organic Pollutants by the Stockholm Convention).

Due to their high water solubility and low volatility, most of PFAS are measured mainly in groundwater (rather long chains) and surface water (rather short chains). In the environment (pH between 5 and 9), PFAS are normally present in anionic form, thus explaining their low sorption to soils and sediments (which are negatively charged) and their high solubility in water. Some PFAS, as for ex. Sulfo-betaines as 6:2-FT(S)AB (= Capstone B) or 6:2-FTS (= H4-PFOS), are much more soluble under even only slight alkaline conditions (F. Karg 2022). A very important point is, that poly-fluorinated PFAS (called Precursors) are biotransformed to stable per-fluorinated PFAS. Further biodegradation was not proved up today.

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Due to their hydrophilic & hydrophobic character, PFAS tend to accumulate at the water/air interface for surface waters. Their tendency to accumulate at the top of a groundwater table is a current topic of research. With a dispersion potential in groundwater equivalent or even higher than that of benzene and trichloroethene, PFAS are measured over long distances, even at the North Pole. Thus, the sampling strategy should include sampling points far downstream of the source.

Data on saturation vapor pressure (very different depending on the ionised or non-ionised form), Henry's law constant (H) and organic carbon/water partition coefficient (K_{oc}) available in the literature. Some PFAS, as the Fluorotelomer alcohols (FTOH) are volatile.

Physical-chemical data, for ex. mainly for PFOS and PFOA, are very variable and often only exist for the acid form (and not for the salts). In order to better estimate the behavior of PFAS, data related to the appropriate form should be selected and on-site measurement of the adsorption coefficient based on the soil organic carbon content (K_{oc}) is recommended.

Particular care should be taken when collecting PFAS, avoiding the use of materials containing "fluorine" as Teflon, etc. to avoid false positive analysis results. Analytical laboratories encounter difficulties in analyzing PFAS: only about 40 compounds can be quantified normally. Best laboratories could analyze about 500 PFAS compounds, with limits of quantification sometimes higher than the regulatory values and suspicions of interference, increasing the difficulties of interpretation. To compensate for these uncertainties, it is recommended to increase the number of blank and replicate analyses.

3. Toxicology and Health Risk Assessment:

As some PFAS decompose into perfluorinated acids which are more toxic and do not degrade, the precursors should also be analyzed.

Risk assessment: about 40 compounds (only) well known for their effects on human health and for which knowledge is evolving

The risk assessment of PFAS for human health and the environment is complex: it cannot consider the mixing effect of these 9 000 – 12 000 compounds because there are not 9 000 Toxicological Reference Values (TRV). Human toxicity data are principally available for : PFOS, PFOA but also for PFHxA, PFDA and PFNA.

Major toxicological effects are summarized in the following Table (US-EPA: 2021):

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	# of Carbons	Liver	Developmental	Reproductive	Immune	Hematologic	Thyroid	Neuro-behavioral	Tumors
Perfluoroalkyl Carboxylates									
PFBA	4	■	■	■	□	■	■	□	□
PFPeA	5	□	□	□	□	□	□	□	□
PFHxA	6	■	■	■	□	■	■	□	□ (Negative)
PFHpA	7	■	□	□	□	□	□	□	□
PFOA	8	■	■	■	■	■	■	■	■
PFNA	9	■	■	■	■	■	■	□	□
PFDA	10	■	■	■	■	■	■	■	□
PFUnA	11	■	■	□	■	□	□	□	□
PFDoA	12	■	■	■	■	■	□	■	□
Perfluoroalkyl Sulfonates									
PFBS	4	■	■	■	■	■	■	□	□
PFHxS	6	■	■	□	□	■	■	■	□
PFOS	8	■	■	■	■	■	■	■	■
Per- & Polyfluoroalkyl Ether Replacements									
ADONA	6	■	■	□	□	■	□	□	□
HFPO-DA GenX	6	■	■	■	■	■	■	□	■

Effect reported in one or more laboratory animal study
 Effect was evaluated but not found, or effect has not been evaluated

The main routes of exposure are ingestion of contaminated drinking water (tap water) or contaminated food. In 2020, EFSA assessed the exposure of the European population to PFAS through food (EFSA, 2020).

Based on 67,839 samples analysed in 16 countries (of which 33,000 analysed in fish), the main exposure to PFAS comes from fish consumption (in particular carp, eel, roach, perch and sardine) followed by fruit and eggs.

The TRV evolve every year with a drastic decrease in the maximum permissible dose for PFOS and PFOA, divided by a factor of 10 in 10 years. Regarding the mixture effect, the RIVM (RIVM, 2018) proposes to use Relative Potency Factors (RPF), according W. Bil (2021) for 12 compounds, based on liver toxicity effects. EFSA (EFSA, 2020) proposes a different 'exposure indicator' approach.

Despite the to improve the health situation by producing shorter-chain PFAS, the results of toxicological studies do not yet provide reassurance that the toxic effects of these alternatives will be reduced.

4. SFSE's PFAS Management Guideline Fact Sheets:

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In order to help the authorities, industries, consultancy firms, etc. concerned by this PFAS Pollution Management issues, the SFSE has decided to summarize in a pragmatic manner the available information and to draw up a practical Guideline with Thematic Sheets made available to as many concerned people as possible, , as for ex.:

1. **Fact Sheet 1:** Generalities, identity and Chemistry of PFAS,
Fiche 1 : Note chapeau : Connaissances générales; identité et chimie des PFAS,
2. **Fact Sheet 2:** Regulations regulatory context of PFAS,
Fiche 2 : Cadre réglementaire & juridique des PFAS,
3. **Fact Sheet 3:** Contamination Sources of PFAS,
Fiche 3 : Sources de contamination des PFAS,
4. **Fact Sheet 4:** Environmental behavior and biotransformation of PFAS,
Fiche 4 : Devenir dans l'environnement et biotransformation des PFAS,
5. **Fact Sheet 5:** Sampling Procedures (soil, groundwater, surface water, soil vapor, ambient air, Human Bio-Monitoring),
Fiche 5 : Méthodes de prélèvements (sols, eaux souterraines, eaux de surface, gaz du sol, air ambiant, Human Bio-Monitoring),
6. **Fact Sheet 6:** Analysis Methods (soils, groundwater, surface water and others),
Fiche 6 : Méthodes d'analyses (sols/eaux souterraines/eaux de surface et autres),
7. **Fact Sheet 7:** Chemical Background and Recommendations for definition of chemical backgrounds (data banks),
Fiche 7 : Bruit de fond et recommandations pour la détermination du bruit de fond (base de données),
8. **Fact Sheet 8:** Environmental Behavior – Physical & chemical Parameters (data banks),
Fiche 8 : Comportement dans l'environnement – Paramètres physico-chimiques (base des données),
- 9a. **Fact Sheet 9a:** Toxicity – Toxicological Reference Values,
Fiche 9a : Toxicité - Valeurs Toxicologiques de Référence (VTR),
- 9b. **Fact Sheet 9b:** Health Risk Assessments,
Fiche 9b : Evaluation des Risques Sanitaires (ERS),
10. **Fact Sheet 10:** Consideration of Toxic Effects in case of PFAS Mixtures,
Fiche 10 : Considération de l'effet toxique de mélange des PFAS,
11. **Fact Sheet 11:** Exposure and Human Bio Monitoring Data (HBM),
Fiche 11 : Expositions et Données de bio-monitoring humain (HBM),
12. **Fact Sheet 12:** Soil Remediation for PFAS Contaminations,
Fiche 12 : Méthodes d'assainissement - dépollution des sols,
13. **Fact Sheet 13:** Groundwater Treatments for PFAS Contaminations,
Fiche 13 : Méthodes d'assainissement - dépollution des eaux souterraines,
14. **Fact Sheet 14:** Soil Vapor and Ambient Air Treatments for PFAS Contaminations,
Fiche 14 : Méthodes d'assainissement des gaz du sol et de l'air ambiant.

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These PFAS SFSE-Fact Sheets will be regularly up-dated and will be made available on <https://www.sfse.org/accueil>

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Screening de PFAS semi quantitatifs dans les eaux: REX

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Les substances per- et polyfluoroalkyles (PFAS) sont un groupe de composés synthétiques qui sont utilisés dans une variété d'applications industrielles et de produits de consommation, mais ce sont aussi des polluants persistant dans l'environnement.

Ces substances peuvent potentiellement avoir des effets néfastes sur l'environnement et la santé humaine ; au niveau local, la pollution pourrait concerter différents types de substances ainsi que des sous-produits ou des produits de réaction de sorte à nécessiter une évaluation des risques et cumulative tenant compte de la multiplicité des expositions.

En raison d'effets néfastes connus sur la santé, les deux composés PFAS les plus notoires, le PFOA et le PFOS, ont été interdits d'utilisation et de production dans la majeure partie du monde. Aujourd'hui, on estime qu'il y a eu 4 000 à 6 000 composés de type PFAS fabriqués depuis l'introduction de ces produits chimiques.

Le développement continu de composés émergents et l'identification consécutive d'espèces chimiques émergentes nécessitent des techniques de dépistage pour détecter de nouveaux contaminants préoccupants.

Les méthodes ciblées actuelles permettent une détection sensible de ces composés, mais ne se concentrent que sur une petite partie des PFAS.

Une technique non ciblée ou screening pourrait fournir une caractérisation plus complète de la contamination par les PFAS dans un échantillon. La découverte de PFAS non ciblée est également la façon dont l'EPA essaie de gérer les analyses de PFAS pour l'avenir.

Ce sujet correspond à une étude de cas spécifique, une recherche expérimentale menée et coordonnée par Aquasoil SRL pour le compte de la Région des Pouilles, avec la collaboration du Département des sciences et technologies biologiques et environnementales de l'Université du Salento, Mérieux NutriSciences et le Département de génie chimique et biochimique de l'ouest de l'Ontario.

De nouvelles données ont été apportées sur ce projet.



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Programme

Mercredi 14 juin 2023

Présidents :

- Dr. Frank Karg, SFSE & HPC International
- Dr. Christophe Rousselle / ARET & ANSES

08h45

Accueil des participants

Gestion – Dépollution – Traitements – Recherche

09h10

Utilisation d'outils de pointe pour déterminer la présence et le transport de PFAS dans une ancienne base de l'armée de l'air

Ryan Samuels, Géologue/Stratigraphe - AECOM (États-Unis)

09h30

PFAS dans les Biosolides : investigation, directives réglementaires et gestion - une perspective australienne

Fouad Abo, Directeur Technique Senior Environnement - GHD (Australie)

09h50

Essai pilote d'élimination des PFAS des eaux souterraines dans un aéroport suédois à l'aide de la résine échangeuse d'ions GAC et Sorbix™

Jeffrey Lewis, Directeur Général - ECT2, Montrose Environmental Group AB (Suède)

10h10

Solutions in-situ flexibles et durables pour les systèmes source-panache de PFAS mettant en œuvre diverses applications du charbon actif colloïdal

Kris Maerten, Responsable Technique Europe - Regenesis (Belgique)

10h30

Guide pratique du traitement des PFAS dans l'eau potable et polluée avec une résine échangeuse d'ions à usage unique

Cathy Swanson, Directrice du département Développement Durable - Purolite (États-Unis)

10h50

Pause-café / Thé

11h05

Démontrer l'efficacité du carbone Intraplex® pour l'immobilisation in situ des panaches de PFAS sur les sites aéroportuaires

Dr. Julian Bosch, CEO - Intrapore (Allemagne)

11h25

Rôle des résines échangeuses d'ions Lewatit dans l'assainissement des PFAS

Dr. Dirk Steinhilber, Directeur Marketing Technique - Lanxess Deutschland (Allemagne)

11h45

Nouvelles nanoparticules ferromagnétiques pour l'assainissement de l'eau à l'aide de cyclodextrines

Laura Gouriou, Doctorante - Université de Rouen Normandie

12h05

Immobilisation in situ des PFAS pour le contrôle du panache des eaux souterraines : défis pour l'applicabilité

Jeroen Vandenbruwane, Directeur - Injectis (Belgique)

12h25

Développement de méthodes appropriées pour la préparation d'échantillons de PFAS dans différentes matrices

Dr. Suman Kharel, Spécialiste des applications - LCTECH (Allemagne)

12h45

Pause déjeuner

14h00

Traitements durables in-situ des PFAS dans les sols et eaux souterraines contaminés, notamment par lavage via des Biopolymères protéiniques

Introduction

- *Dr. Frank Karg / Président HPC International SAS/ Directeur Scientifique Groupe HPC AG - ARET, SFSE et INOGEN & IBS - Expert Judiciaire (France & Allemagne)*

Technologie

- *Dr. Stephan Hüttmann, Directeur Général, FABEKO-group (Allemagne)*

14h30

Bref aperçu des méthodes d'assainissement des PFAS pour le sol et les eaux souterraines

Lars Van Passel, Directeur Technique - RSK Benelux (Belgique)

14h50

Traitement des sols contaminés par les PFAS : retour d'expérience sur deux projets d'assainissement de grande envergure

Nicolas Etard, Ingénieur Projets - Züblin Umwelttechnik GmbH

15h10

Gestion des sols contaminés par les PFAS : enseignements tirés des stratégies de traitement par lessivage des sols mises en œuvre en France et en Belgique

Olivier Sibourg, Directeur Technique Adjoint - SARPI Remediation France

15h30

Pause-café / Thé

15h45

Couplage OI basse pression/électro-oxydation anodique pour le traitement des eaux contaminées par des PFAS

Jean-Philippe Croué, Professeur - Université de Poitiers

16h05

Partager une approche de production responsable pour les fluoropolymères

Cédric Triquet, Directeur Stratégie et Plaidoyer EMEA - Materiaux de performance avancés - Chemours International Operations (Suisse)

16h25

Traitemet et valorisation des eaux contaminées par les PFAS au moyen de procédés atmosphériques eau-plasma

Pia Kronsbein, Chargée de projet - Hydr.o. Geologen und Ingenieure (Germany)

16h45

Lixivation de PFAS à partir d'un site contaminé par des AFFF : des expériences de simulation de précipitations sur des colonnes de sol non perturbé révèlent la contribution des particules de sol au transport de PFAS

Elisabeth Fries, Doctorante - UGE & INRAE

17h05

Life Capture : combiner de nouveaux protocoles analytiques avec des technologies efficaces pour une remédiation durable des PFAS

Axelle Mineur, Bioingénieur/Ingénieur agronome – Consultante - ABO (Belgique)

17h25

Discussion & Conclusions:

Dr. Frank Karg, HPC International - Inogen, Expert judiciaire (France & Allemagne)

17h45

Fin du congrès



Program

Wednesday, June 14, 2023

Chairs:

- Dr. Frank Karg, SFSE & HPC International
- Dr. Christophe Rousselle / ARET & ANSES

08:45

Welcoming participants

Management – Remediation – Treatments – Research

09:10

Applying state-of-the-practice tools to determine the presence and transport of PFAS at a former air force base

Ryan Samuels, Geologist/Stratigrapher - AECOM (United States)

09:30

PFAS in Biosolid, investigation, regulatory guidelines and management - An Australian perspective

Fouad Abo, Senior Technical Director Environment - GHD (Australia)

09:50

Pilot testing PFAS removal from groundwater using GAC and Sorbix™ ion exchange resin at a Swedish airport

Jeffrey Lewis, General Manager - ECT2, Montrose Environmental Group AB (Sweden)

10:10

Flexible and sustainable in situ solutions for PFAS source-plume systems using a variety of colloidal activated carbon applications

Kris Maerten, Technical Manager Europe - Regenesis (Belgium)

10:30

Practical guide to PFAS treatment for both potable and polluted water with single-use ion exchange Resin

Cathy Swanson, Environmental Sustainability Segment Manager - Purolite (United States)

10:50

Break for Coffee and Tea

11:05

Demonstrating the effectiveness of Intraplex® carbon for the in situ immobilization of PFAS plumes at airport sites

Dr. Julian Bosch, CEO - Intrapore (Germany)

11:25

Role of Lewatit Ion Exchange resins in PFAS remediation

Dr. Dirk Steinhilber, Technical Marketing Manager - Lanxess Deutschland (Germany)

11:45

New Ferromagnetic nanoparticles for water remediation using cyclodextrins

Laura Gouriou, PhD - Université de Rouen Normandie

12:05

In-situ PFAS immobilization for groundwater plume control: challenges for the applicability

Jeroen Vandenbruwane, Director - Injectis (Belgium)

12:25

Development of suitable methods for PFAS sample preparation in different matrices

Dr. Suman Kharel, Application Specialist - LCTECH (Germany)

12:45

Lunch Break

14:00

Sustainable In-situ Treatments of PFAS in contaminated Soil and Groundwater, Washing with Protein Bio-polymers

Introduction

- *Dr. Frank Karg / President HPC International SAS / Scientific Director HPC AG Group - ARET& SFSE and INOGEN & IBS, Judicial Expert (France & Germany)*

Technology

- *Dr. Stephan Hüttmann, Managing Director, FABEKO-group (Germany)*

14:30

Short overview of PFAS remediation methods for soil and groundwater

Lars Van Passel, Technical Director - RSK Benelux (Belgium)

14:50

Treatment of soil contaminated by PFAS: feedback from two large-scale remediation projects

Nicolas Etard, Project engineer - Züblin Umwelttechnik GmbH

15:10

PFAS-contaminated soil management: Learned lessons from soil-washing treatment strategies implemented in France and Belgium

Olivier Sibourg, Deputy Technical Director - SARPI Remediation France

15:30

Break for Coffee and Tea

15:45

Coupling low pressure RO with anodic electro-oxidation for the treatment of PFAS contaminated water

Jean-Philippe Croué, Professor - Université de Poitiers

16:05

Unpacking a responsible manufacturing approach for fluoropolymers

Cédric Triquet, EMEA Strategy and Advocacy Director - Advanced Performance Materials - Chemours International Operations (Swiss)

16:25

Treatment and recovery of PFAS-contaminated water by means of atmospheric water-plasma Processes

Pia Kronsbein, Project employee - Hydr.o. Geologen und Ingenieure (Germany)

16:45

PFAS leaching from an AFFF-contaminated site: rainfall simulation experiments on undisturbed soil columns reveal the contribution of soil particles to the transport of PFAS

Elisabeth Fries, PhD - UGE & INRAE

17:05

Life Capture: Combining novel analytical protocols with effective Technologies for sustainable PFAS Remediation

Axelle Mineur, Bioengineer/Agricultural Engineer – Consultant - ABO (Belgium)

17:25

Discussion & Conclusions:

Dr. Frank Karg, HPC International - Inogen, Judicial Expert (France & Germany)

17:45

End of the Congress

Applying State-of-the-Practice Tools to Determine the Presence and Transport of PFAS at a Former Air Force Base

Ryan C. Samuels - AECOM (United States)

Lead Geologist/Stratigrapher

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Since the late 1940s, the US military has released property once used to support war efforts. Since the late 1980s, over 350 installations have been closed under a formal US Department of Defense program (BRAC). As a requirement of returning property to the public, there are requirements to investigate, manage, and mitigate environmental issues that resulted from military use. The former Reese Air Force Base in Lubbock, TX, was founded in 1941 and closed under BRAC in 1997, and legacy chlorinated solvent groundwater contamination was cleaned up by the military based on hydrogeological understanding and Texas state environmental requirements. With the discovery of per- and polyfluoroalkyl substances (PFAS) releases at this former airfield, the need arose to assess these emerging contaminants and an Affected Property Assessment (APA) was undertaken.

The objective of the APA investigation was to define the nature and extent of PFAS in soil, sediment, surface water, and groundwater at the former Reese AFB. The project posed numerous challenges including 10+ release areas, highly heterogeneous geologic conditions, seasonal/irregular agricultural pumping, complex surface water/groundwater interactions, exceedances in private wells 2.5 miles (4 kilometers) downgradient of the former Base extent, and redistribution of PFAS via the historic Injection/Extraction well network.

To accomplish the APA, AECOM developed a first-of-its-kind 3-D numerical PFAS groundwater fate & transport (GW F&T) model. The GW F&T model integrated PRISM® - a state-of-the-art correlation technique that constructs a predictive, 3-D geologic framework and more accurately define subsurface heterogeneity – to provide a high-resolution, geologically-defensible basis for estimating measured hydraulic conductivity data and constraining groundwater flow estimates. The model provided unparalleled understanding of preferential contaminant migration pathways and was used to refine mass flux estimates, guide optimal placement/screen depths of 150+ monitoring wells, and streamline investigations to address data gaps. Over the next 3 years, the model will serve as an invaluable tool for evaluating remedy alternatives, designing an optimized remedy, and effectively communicating remedial decision-making to all stakeholders.

PFAS in Biosolids: investigation, regulatory guidelines, and management: An Australian perspective

PFAS dans les Biosolides: investigation, directives réglementaires et gestion : une perspective australienne

Author: Dr Fouad Abo, A GHD principal, a Senior Technical Director and an Environmental Auditor, GHD PTY LTD, Melbourne, Australia

Key words: PFAS, Biosolids, Investigation, Management, Regulations and Guidelines

Objectives: This paper aims at sharing the Australian experience on investigating, regulating and management of PFAS including in Biosolids. It aims at providing insight about how this emerging group of contaminants are investigated and regulated in Australia. Biosolids are a critical issue for water authorities around the country and are a legacy of water treatment plants that causing important exposure and liability issues.

Introduction

Biosolids are an ongoing and a legacy important problem resulting from all water treatment plants. There are millions of tons of biosolids as a legacy but also ongoing that are put in some big stockpiles. Water Authorities around Australia are struggling to manage them let alone treating or disposing of them. PFAS is a recent issue found in Biosolids and hence causing issues for the regulators and water authorities alike. PFAS are making the potential reuse of biosolids problematic, and authorities are rightly concerned about ongoing liability, which this paper is providing insight about what is the status quo of the investigation and management. We also discuss the National PFAS Environment Management Plan (EMP3) evolution over the last few years and how does it apply to PFAS in Biosolids.

This paper attempt to share an Australian perspective on how the PFAS are investigated in Biosolids but also how it is regulated and managed.

Methodology

A summary of how to investigate PFAS in Biosolids using existing guidelines to obtain a representative data of the PFAS content in Biosolids in order to classify them and also investigate the appropriate management and treatment is discussed. Short falls are exposed through the investigation experience and attempt to share our lesson learned.

We discuss a combination of a long-term management and reducing PFAS source together with a pragmatic approach.

Results & Discussions

Due to the huge volume of biosolids mount present, we share our opinion on using a statistical interpretation using polygon to maximise the interpretation and exploitation of sampling and analysis results. This also involves conducting a data analysis using digital info related to environmental sampling locations Treatment Capacity Increase. Sampling locations were mapped, and spatial data presented in ESRI Shapefile file format for the sampling locations in the figures. This also involved setting up GIS layer

maps to improve the statistical analysis and probabilistic analysis of the contamination map across the biosolids and land around it.

Conclusions

This study resulted on a high confidence estimation of volume that require treatment/management. It also provided a good mapping of the spatial and vertical distribution of contamination. We will also conclude by how the national EMP 3 of PFAS is evolving in Australia.

Technical Summary, Pilot testing PFAS removal from groundwater using GAC and Sorbix™ ion exchange resin at a Swedish airport

Jeffrey Lewis, Ph.D, General Manager ECT2 Sweden
jklewis@ect2.se

Abstract

A groundwater treatment pilot for PFAS contamination was run at Skövde airfield in Sweden to determine whether granular activated carbon (GAC) or Sorbix™ regenerable ion exchange resin (IX) system would provide a more cost effective solution. The pilot ran for seven months in 2022 and the measured breakthrough curves were used to calculate both capital costs and operational costs for a full size treatment system capable of handling a flow of 20 m³/hr. The conclusion of the test was that the Sorbix™ media was eleven times as effective as GAC at adsorbing PFAS, the annual operational costs of a regenerable Sorbix™ ion exchange system will be approximately 3% of a GAC system, and the capital costs to install the two systems would be roughly equivalent.

Background/Introduction

Skövde airport is a regional airport in central Sweden that is being converted by the municipality into an industrial park. Consultants reports showed PFAS concentrations in the groundwater up to 17 000 ng/L, much higher than the permissible level of 90 ng/L. ECT2 was asked to perform a pilot test on the groundwater to determine the most cost-effective solution for a pump-and-treat system.

Experimental

The process flow diagram of the pilot is shown in Figure 1. Two of these were installed, one for the GAC and one for the Sorbix™ IX. The experimental conditions are shown in table 1. Water samples of the influent, following column 1 and following column 2 were obtained weekly for seven months.

Results

The breakthrough curves for the GAC and Sorbix™ IX are shown in Figure 2. The breakthrough in the GAC reached 90 ng/L after 7000 bedvolumes while the breakthrough in the Sorbix reached 90 ng/L after 80 000 bedvolumes (projected not measured). The consequences of these breakthrough times, combined with the physical size of the full-size plant dictated by the experimental conditions in Table 1 were used to calculate the operational costs and capital costs of a plant. These details will be presented during the conference.

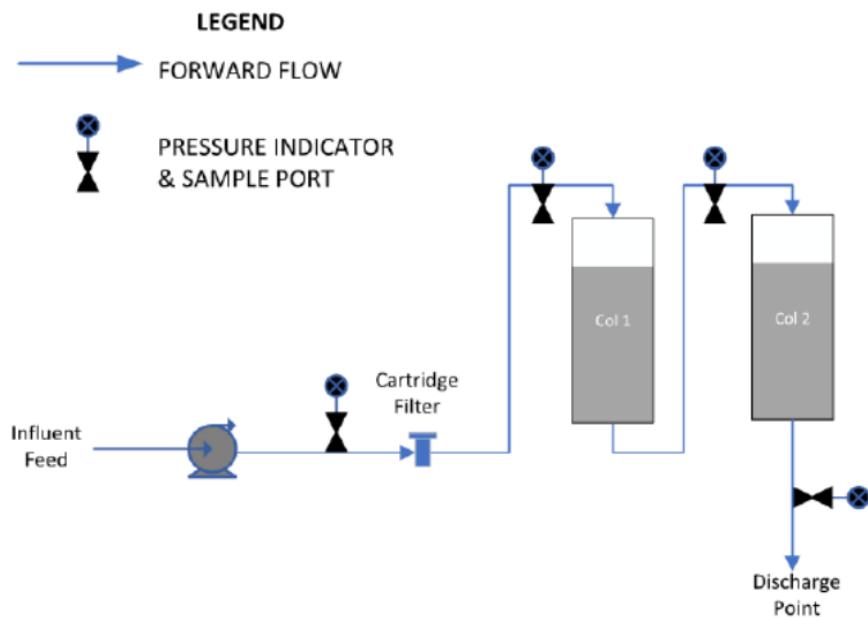


Figure 1. Process flow diagram of the pilot

Table 1.

	GAC	Sorbix™ IX
Flow	135 ml/min	300 ml/min
Bedvolumes (Lead/Lag)	1,98 L / 1,98 L	1,75 L / 1,75 L
Contact time (Lead/Lag)	14,4 min / 29 min	6 min / 12 min

Experimental conditions of the pilot

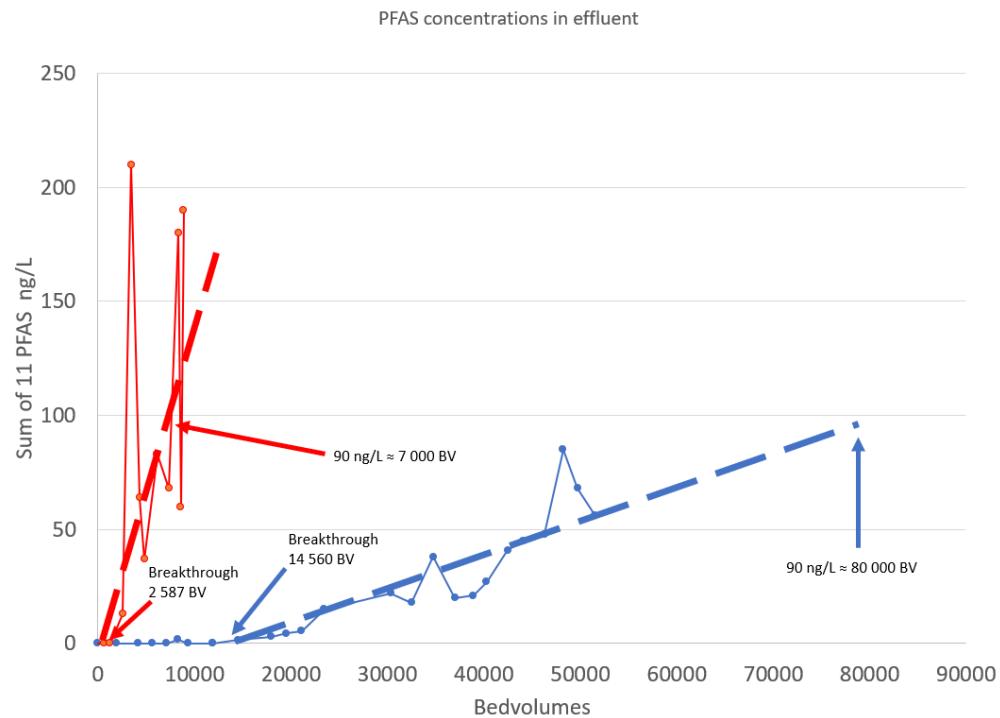


Figure 2. Breakthrough curves for the GAC (Red) and Sorbix™ IX (Blue)

Flexible and sustainable *in situ* solutions for PFAS source-plume systems using a variety of colloidal activated carbon applications

Solutions in situ flexibles et durables pour les systèmes source-panache de PFAS mettant en œuvre diverses applications du charbon actif colloidal

Kris Maerten (REGENESIS) and Mariangela Donati (REGENESIS)

Ir. Kris Maerten, Technical Manager, Europe

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Contamination of soils and groundwater with Per- and polyfluoroalkyl substances (PFAS) can result from the historic use of Aqueous Film Forming Foams (AFFF), particularly associated with fire training facilities. Accumulation of PFAS in the soil creates significant reservoirs of contaminant mass. Further vertical movement of PFAS may occur through leaching into infiltrating precipitation. Eventually this may result in discharge to groundwater, plume propagation and risk to downgradient environmental or human receptors. The mobility of the PFAS compounds, combined with the toxicity of longer chain compounds, means that these source areas can create and maintain widespread environmental risk long after the discharge event has occurred.

Treatment of both source and plume zones has been considered problematic due to the recalcitrant nature of PFAS compounds. Waste management facilities increasing do not want to receive PFAS impacted soils excavated from the site. The extraction and treatment of impacted groundwater is inefficient, needing a large volume of water to be removed and filtered for many years –requiring long-term energy use, constant waste production and an uncertain end-point to the treatment.

In situ sequestration of the PFAS compounds provides a sustainable and long-term alternative solution that has now been used on 40 sites worldwide. This comprises the application of Colloidal Activated Carbon ((CAC), a liquid comprising 1-2µm particles of activated carbon suspended in water) in different locations within the source-plume system. In the impacted vadose zone soils, CAC is physically mixed with the soil and other sorption modifiers to reduce the leachability of the PFAS. A horizontal barrier of CAC is emplaced just under the treated soils to remove residual PFAS infiltration and prevent further discharge to the groundwater. In the source area groundwater, injection of CAC in a grid pattern is used to target the lateral and vertical extent of high concentrations of dissolved PFAS. The CAC attaches to the aquifer material and removes the PFAS from the groundwater through sorption, preventing further migration into the body of the plume.

The combination of prevention of PFAS discharge into the groundwater with adsorption of contaminant mass already in the source area groundwater, will allow the enhanced attenuation of the residual plume. Site treatment can be further enhanced by combining these source-treatments with a plume application. This is achieved through the low-pressure injection of CAC into flux zones using Direct Push Technology (DPT) at the site boundary or upgradient of a receptor. This creates a treatment zone comprising a 1-2µm layer of activated carbon on the aquifer material, which does not affect the groundwater flow, while acting as a purifying filter, immediately removing the PFAS influx from the groundwater through adsorption, thereby eliminating the risk to downgradient receptors.

In situ treatment of PFAS using CAC is a sustainable solution as the installation will take only a few weeks of site activity, after which the treatment will continue without the need for further energy use or onsite machinery and no waste is created at the surface.

This presentation will introduce CAC technology and explain the processes through which it can be used to treat source and plume zones and how designs are tailored to each site. Laboratory data will demonstrate sorption efficacy and leachability reduction. Practical application and validation will be discussed. A case study including treatment results will be shown of source treatment at a military base in the USA and full-scale plume treatment at an airport site in the UK.

This presentation will be of interest to practitioners, site operators and regulators interested in how to practically and sustainably manage PFAS contamination.

Practical Guide to PFAS Treatment for both Potable and Polluted Water with Single-Use Ion Exchange Resin

Cathy Swanson, MBA, B.S. Ch.E.
Environmental Sustainability Segment Manager
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Objectives:

The objective of this presentation is to introduce participants to worldwide success stories using PFAS Selective Ion Exchange Resin for treatment of per - and polyfluoroalkyl substances, including why resin was chosen, how it works, and lessons learned.

Content:

- Show the process of going through the steps of design and modeling, piloting, and finally full-scale implementation for water sites with PFAS Selective Single Use Ion Exchange Resin.
- Review the accuracy of the throughput models
- Discuss why ion exchange was chosen over alternative technologies for each site.
- Discuss the cost modeling
- Show how resin deals with both short chain and long chain PFAS.
- Discuss how buffered resins can be provided where there are corrosivity and chloride to sulfate mass ratio (CSMR) concerns.
- Discuss best practices for disposal of spent resin

In this presentation, we will do a quick overview of the basics of PFAS. We'll talk about current regulations and the drivers behind those. We'll talk briefly about treatment options, and then dive in to why single use ion exchange resin has become one of the most common best available technologies (BAT). To illustrate the technology, design guidance, case studies, and lessons learned will all be reviewed. Finally, we'll end on how to dispose and destroy the "forever chemicals."

Demonstrating the effectiveness of Intraplex® carbon for the in situ immobilization of PFAS plumes at airport sites

Julian Bosch, Intrapore GmbH, Essen, Germany

Erwan Goulian, Evonik Authorized Technical Representative, France

Keywords: PFAS, in situ remediation

At Intrapore, over the last years, we put considerable effort into a new generation of colloidal activated carbon for the in situ remediation of PFAS (per- and polyfluoroalkyl substances) in groundwater. The new material, Intraplex® carbon, a novel generation of specialized colloidal carbon, was synthesized by a unique activation process and characterized for its specific surface area, pore size distribution, and adsorption capacity for PFAS. Results showed a significantly higher specific surface area and improved adsorption capacity compared to conventional activated carbon. The new material was then applied in real-world contaminated groundwater and showed significant reduction of PFAS concentrations in a short period of time. These studies demonstrate the potential of the new colloidal activated carbon for efficient and cost-effective in situ PFAS remediation. In particular, the material showed excellent performance at large airport sites where PFAS contamination from firefighting training areas is a major concern. Results showed a high removal efficiency of PFAS and a quick reduction of concentrations to below regulatory limits. This suggests that the new colloidal activated carbon is a promising solution for the remediation of PFAS contaminated groundwater at large airport, industry, infrastructure or military sites.

Efficient removal of per- and polyfluorinated alkyl substances from potable- and waste water with selective Lewatit® resins

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PFAS is a family of highly efficient surface-active agents used in various applications such as firefighting foams, water repellent textiles, cook ware, galvanics, and paints. However, when they are not handled and disposed of thoroughly they can leach into the groundwater where they persist as a result of their high chemical stability. Due to their hazardous potential, drinking water limits have been set very strictly. Standard technologies such as activated carbon usually cannot comply with the low effluent limits for short-chain PFAS that represent the biggest challenge in remediation. On the other hand, Reverse Osmosis, an alternative technology for PFAS removal, generates large amounts of aqueous concentrates that have to be handled. Therefore, LANXESS has developed a new type of selective ion exchange resin (IER) that reduces PFAS reliably below the drinking water limits and which can safely be disposed after use.

In order to face this challenge, LANXESS has developed the novel ion exchange resin Lewatit® TP 108 DW, which has a very high selectivity towards different PFAS types.

In this paper we will

Lewatit® TP 108 DW is particularly applicable for the removal of short- and long-chain PFAS to very low levels, including perfluorooctanoic acid (PFOA), perfluorooctanesulfonic acid (PFOS), perfluorononanoic acid (PFNA), perfluoroheptanoic acid (PFHpA), perfluorohexanoic acid (PFHxA), perfluorohexanesulfonic acid (PFHxS), perfluorobutanesulfonic acid (PFBS) and perfluorobutanoic acid (PFBA). GAC is a widely used method for removing organic contaminants (e.g. PFAS). It is a proven technology and works in special conditions (e.g. high oxidizer, high TDS), but has major drawbacks including low selectivity, low loading capacity, and slower absorption kinetics. RO can be used to treat PFAS water, but the reject is not concentrated enough to destroy PFAS economically. Therefore another PFAS concentration step and final treatment is required. It has a major disadvantage of resulting in a large volume of concentrated PFAS waste stream to be further treated. In terms of the PFAS removal efficiency, both IER and RO provide much better performance than GAC and UF, as shown by the following data (Figure 1). Based on overall consideration, IER is a better choice for treating PFAS than the other two alternative methods.

The excellent performance of PFAS selective ion exchange resin Lewatit® TP 108 DW leads to a substantial benefit when it comes to a cost comparison of different PFAS treatment technologies. Due to the lower selectivity towards PFAS, GAC breaks through about five times earlier than Lewatit® ion exchange resins. As a result, customers need to replace the Lewatit® ion exchange resin less frequently and achieve savings in investment costs. The much shorter EBCT of IER enables a smaller footprint while treating a larger volume of waste stream. A cost calculation for a plant with five years' operation to remove PFHpA (Figure 2), indicates Lewatit® TP 108 DW costs 58% less than GAC, and 29% less than a competitive resin.

New Ferromagnetic Nanoparticles for Water Remediation using Cyclodextrins

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Nowadays, the recovery of PFAS in contaminated media is provided by Granular Activated Carbon (GAC), Powdered Activated Carbon (PAC) and Anionic Resins. But, for instance, the low selectivity and low affinity of GAC for PFAS makes this technique expensive to reach the regulatory thresholds.^{1,2,3} Moreover, once the PFAS loaded on GAC, the adsorbent regeneration and the contaminant release is energy-consuming. Meanwhile, other conventional techniques such as oxidation, UV irradiation, biodegradation, free chlorine treatment or coagulation, are known to be inefficient for PFAS remediation.⁴ Thus, we aim to find new process, in other words, new adsorbents for PFAS removal able to recover PFAS and support *via* a low energy cost and a minimum of water-consuming suitable for industrial uses. Finally, to reach a circular economy a recycled starting material has been selected.

Recently, β -cyclodextrin (β -CD) has become a hot topic as potential adsorbent. CDs are non-toxic natural macrocycles resulting from the degradation of starch. Therefore, these water-soluble biomaterials are cheap and produced on industrial-scale for pharmaceutic, analytic, cosmetic, and food industries. β -CD is composed of 7 glucose units linked in α -1,4 and known for their host-guest inclusion properties due to their truncated cone shape promoting a hydrophobic inner cavity (Figure 1).^{5,6}

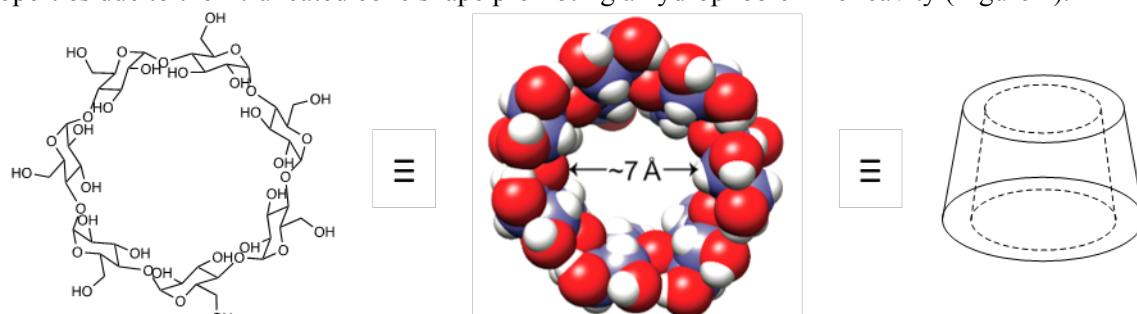


Figure 1: Different representations of β -cyclodextrin

Interestingly, it turns out that β -CD forms supramolecular inclusion complexes with the fluorinated hydrophobic chain of PFAS with good binding affinities (Figure 2).

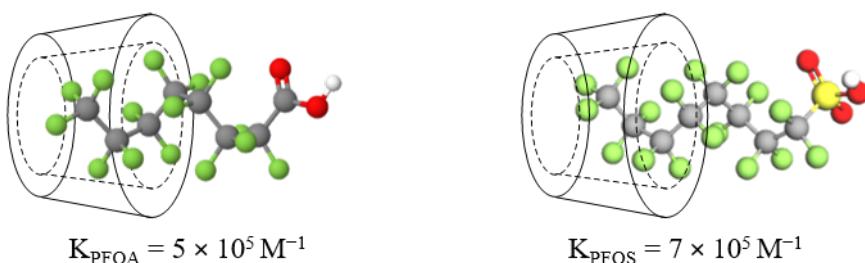


Figure 2: 1:1 Host-Guest inclusion complexes β -CD:PFOA and β -CD:PFOS

¹ Appleman, T. D.; Dickenson, E. R. V.; Bellona, C.; Higgins, C. P. *Journal of Hazardous Materials* **2013**, *260*, 740–746.

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³ Xiao, L.; Ching, C.; Ling, Y.; Nasiri, M.; Klemes, M. J.; Reineke, T. M.; Helbling, D. E.; Dichtel, W. R. *Macromolecules*. **2019**, *52*, 3747–3752.

⁴ Rahman, M. F.; Peldszus S.; Anderson W.B. *Water Research*. **2014**, *50*, 318–40.

⁵ Morin-Crini, N.; Fourmentin, S.; Fenyvesi, É.; Lichtfouse, E.; Torri, G.; Fourmentin, M.; Crini, G. *Environ. Chem. Lett.* **2021**, *19*, 2581–2617.

⁶ Szente, L.; Fenyvesi, É. *Struct. Chem.* **2017**, *28*, 479–492.

Thus, in 2019, Dichtel and al.³ reported the use of water-insoluble complex polymers based on CDs to extract PFAS by host-guest inclusion complex and electrostatic interactions. Variable capture rates were obtained, depending on the cross-linking density of the polymer, the pH, and the chemical functionalization on the β -CD. Moreover, the affinities are strongly dependent on pH with a capture rate of 0.9 mg/g at pH 3.9 versus 0.06 mg/g at pH 7.2, for instance.⁷

Our goal is to implement an industry-suitable alternative solution by taking advantage of the inclusion properties of CDs grafted on coated magnetite nanoparticles (MNP@coating) as a support. Magnetites are solid insoluble black iron oxide particles with formula $\text{Fe}^{\text{II}}\text{Fe}^{\text{III}}_2\text{O}_4$ crystallized in a spinel form with particle sizes ranging between 100 and 200 nm. They are surrounded by oxide and hydroxide functions at their surface. These iron magnetic nanoparticles have received a considerable attention during recent years because of their potential applications in different fields, such as the reprocessing of water including the removal of metals, radionuclides, organic contaminants, dyes and pathogens from water.^{8,9} Moreover, a synergic association of MNP and CD has been already reported to capture fluoroquinolones¹⁰, malachite green, crystal violet¹¹ or pyrethroids¹² from aqueous media. In addition, these nanoparticles can be easily separated from the solution by magnetic decantation which is a major advantage for industrial separation avoiding clogging and water/solvent uses but are all described at laboratory level.

The MNP selected in this project, are industrially produced¹³ from the recycling of steel industry wastes thanks to a green process promoting a circular economy. In order to improve the stability and the recycling in time of the support, the particles need to be coated. We first focused on a silica layer which has the advantage of immobilizing functionalized branched linkers but also of preserving the magnetic properties after multiple uses (Figure 3).

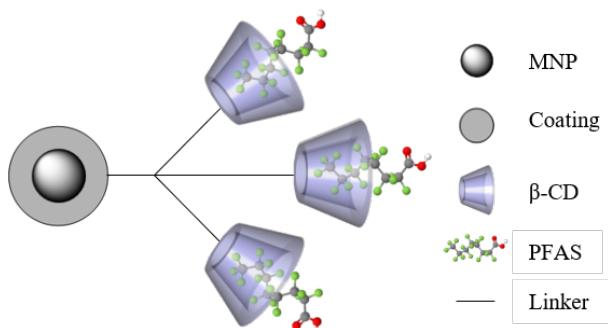


Figure 3: MNP model MNP@coating@ β -CD

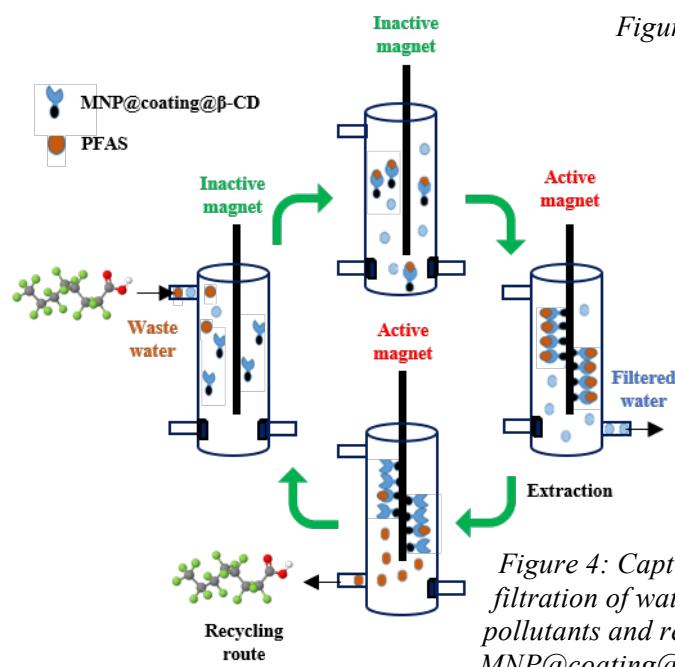


Figure 4: Capture of PFAS, filtration of water, release of pollutants and recycling of the MNP@coating@ β -CD support

For the application, pollutants will be sequestered onto the magnetite which, under the action of a magnetic field, will be extracted, thus depolluting the medium. Under specific stimulus, this magnetite will release the pollutant into a highly concentrated medium and be reused for other cycles (Figure 4).

⁷ Yang, A.; Ching, C.; Easler, M.; Helbling, D. E.; Dichtel, W. R. *ACS Mater. Lett.* **2020**, 2, 1240–1245.

⁸ Alfaro, I.; Molina, L.; González, P.; Gaete, J.; Valenzuela, F.; Marco, J. F.; Sáez, C.; Basualto, C. *J. Ind. Eng. Chem.* **2019**, 78, 271–283.

⁹ You, J.; Wang, L.; Zhao, Y.; Bao, W. *J. Clean. Prod.* **2021**, 281, 124668–124684

¹⁰ Li, F.; Wang, M.; Zhou, J.; Yang, M.; Wang, T. *Microchemical Journal* **2021**, 164, 105955–105967.

¹¹ Moradi Shahrebabak, S.; Saber-Tehrani, M.; Faraji, M.; Shabanian, M.; Aberoomand-Azar, P. *Environ Monit Assess* **2020**, 192, 262–275.

¹² Liu, F.; Yang, X.; Wu, X.; Xi, X.; Gao, H.; Zhang, S.; Zhou, W.; Lu, R. *Food Chemistry* **2018**, 268, 485–491

¹³ <https://www.hymagin.com/magnetite/>

In-situ PFAS immobilization for groundwater plume control: challenges for the applicability

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As PFAS plumes are typically characterized by large volumes and low concentrations, conventional pump-and-treat based groundwater remediation techniques are economically and ecologically not responsible. Additionally, PFAS show relatively low sorption affinity for granular activated carbon (GAC), which makes groundwater treatment costs unaffordable and non-sustainable. Last years, more and more research has been done on the development of new sorbent materials for in situ immobilization of PFAS compounds (Sleep and Juhasz, 2021). Sleep and Juhasz concluded that the immobilization process of PFAS is complex and that the efficiency varies with soil properties, amendment properties and molecular properties of the individual PFAS. They suggested that optimal remediation strategies need to be adjusted according to site specific soil properties and contamination profiles. In the June 2022 ITRC update on treatment technologies for PFAS, it was clearly mentioned that only limited field application have been demonstrated for in situ sorption and stabilization. A relatively large number of sorbents have been developed and tested in batch and column test in the laboratory, but only a few have been applied in the field during pilot test or full-scale applications. Fagerlund et al. (2023) reported on the results of a pilot test performed in a complex Swedish geology and stated that this complex geology made the injection of colloidal activated carbon (CAC) challenging, which also determined the distribution and hence the efficiency of the CAC. The efficiency of field applied sorbents is almost exclusively determined by the quality of contact between the sorbents and the pollution. To achieve this physical contact, the injection technique, the injection parameters and the colloid properties are the variables that can and must be adapted for a given geology. In this presentation, the main challenges to be expected in the transfer from the laboratory results to field application will be explained from the solid experience in reagents injection.

Development of Suitable Methods for PFAS Sample Preparation in Different Matrices

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Abstract. Per- and polyfluorinated alkyl substances (PFAS) get into the environment during their manufacturing process and further also during their use and disposal. Analyzing these PFAS compounds is challenging and dedicated lab equipment helps to avoid blind values and standardizes the processes in sample preparation. Currently, there are existing SPE cartridge solutions for extractions of PFAS from drinking water. However, solid samples (especially soil samples) have often the ability to strongly bind neutral and long chain PFAS. Further, more demanding sample matrices, for instance, environmental and food/feed matrices are handled with a more elaborate dual SPE or a combination of SPE and dispersive clean-up to remove matrix interferences. In spite of the tedious process, all of them lack the ability to enrich some specific PFAS in a sufficient way e.g. neutral sulfonamides or long chain PFAS. The challenge is to develop a SPE solution that is not cost intensive and at the same time effective for the clean-up and enrichment for all types of PFAS analytes. Additionally, some of the PFAS analytes are volatile and an appropriate method for sample concentration is needed. Analyte extraction of solid samples via manual shaking was compared with a pressurized liquid extraction system. For the enrichment and/or purification of PFAS compounds solid phase extraction (SPE) was applied. SPE cartridge solutions of different compositions were used. For the critical evaporation step a vacuum centrifuge with cold trap was used. The samples were subsequently analysed by LC-MS/MS. A pressurized liquid extraction method was successfully developed especially to extract neutral and long chain PFAS from soil. An important point in the development of the method was the establishment of the blind value free PLE system. SPE cartridge solutions of different compositions show high recoveries, low standard variations and reliable clean-up of PFAS analytes in different matrices. An evaporation method where no volatile PFAS compounds were lost has been successfully developed using a vacuum centrifuge in combination with a cold trap. The presented workflow shows a streamlined sample preparation process for the PFAS analysis, which can be used for different kinds of PFAS including neutral sulfonamides and long chain analytes in various matrices. Further, SPE cartridges with a superior performance for enrichment and clean-up of PFAS molecules from water, soil and food matrices were developed.

Développement de méthodes appropriées pour la préparation d'échantillons de PFAS dans différentes matrices

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Résumé. Les substances per- et polyfluoroalkylées (PFAS) pénètrent dans l'environnement pendant leur processus de fabrication et plus loin encore lors de leur utilisation et de leur élimination. L'analyse de ces composés PFAS est difficile et l'équipement de laboratoire dédié aide à éviter les mesures en aveugle et normalise les processus de préparation des échantillons. Actuellement, il existe des solutions de cartouches SPE pour les extractions de PFAS de l'eau potable. Cependant, les échantillons solides (en particulier les échantillons de sol) ont souvent la capacité de lier fortement les PFAS neutres et à longue chaîne. En outre, les matrices d'échantillons plus exigeantes, par exemple, les matrices environnementales et les matrices destinées à l'alimentation humaine et animale, sont traitées avec une double SPE plus élaborée ou une combinaison d'EPS et de nettoyage dispersif pour éliminer les interférences de la matrice. Malgré ce processus fastidieux, tous n'ont pas la capacité d'enrichir suffisamment certains PFAS spécifiques, par exemple des sulfamides neutres ou des PFAS à longue chaîne. Le défi consiste à développer une solution SPE qui ne soit pas coûteuse et

en même temps efficace pour le nettoyage et l'enrichissement de tous les types d'analytes PFAS. De plus, certains des analytes PFAS sont volatiles et une méthode appropriée pour la concentration de l'échantillon est nécessaire. L'extraction d'échantillons solides par agitation manuelle a été comparée à un système d'extraction de liquide sous pression. Pour l'enrichissement et/ou la purification des composés PFAS, l'extraction en phase solide (SPE) a été appliquée. Des solutions de cartouches SPE de différentes compositions ont été utilisées. Pour l'étape critique d'évaporation, une centrifugeuse à vide avec piège à froid a été utilisée. Les échantillons ont ensuite été analysés par LC-MS/MS. Une méthode d'extraction liquide sous pression a été développée avec succès spécialement pour extraire les PFAS neutres et à longue chaîne du sol. Un point important dans le développement de la méthode a été la mise en place du système PLE sans mesure en aveugle. Les solutions de cartouches SPE de différentes compositions présentent des taux de récupération élevés, de faibles variations standard et un nettoyage fiable des analytes PFAS dans différentes matrices. Une méthode d'évaporation où aucun composé PFAS volatile n'a été perdu, a été développée avec succès à l'aide d'une centrifugeuse sous vide combinée à un piège à froid. Le flux de travail présenté montre un processus simplifié de préparation des échantillons pour l'analyse des PFAS, qui peut être utilisé pour différents types de PFAS, y compris les sulfamides neutres et les analytes à longue chaîne dans diverses matrices. En outre, des cartouches SPE avec une performance supérieure pour l'enrichissement et le nettoyage des molécules PFAS à partir de matrices d'eau, de sol et d'aliments ont été développées.

Traitements in situ durables des PFAS dans les sols et les eaux souterraines contaminés, lavage avec des biopolymères protéiques et stabilisation par injection haute pression GAC

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En raison de leur stabilité chimique élevée, de leur mobilité dans l'environnement et de leur potentiel de bioaccumulation, les substances alkylées polyfluorées et perfluorées (PFAS) représentent un groupe de polluants très pertinent en ce qui concerne leur potentiel de danger. En Allemagne et dans d'autres pays, de nombreux sites contaminés par des PFAS ont été identifiés, dont certains couvrent de vastes zones, comme la contamination de la couche arable de plusieurs centaines d'hectares de terres arables dans la région de Rastatt (Baden-Württemberg, GER), où des PFAS contenant les boues de compostage du papier ont été utilisées comme engrais. À l'heure actuelle, aucune technologie de remédiation in situ pour les contaminations par les PFAS dans la zone vadose n'était disponible. Les stratégies d'assainissement de pointe comprennent principalement des méthodes traditionnelles et très coûteuses telles que l'excavation des sols contaminés (et le traitement thermique ou le déversement) ou le pompage et le traitement en tant que mesure de confinement hydraulique des eaux souterraines contaminées.

L'objectif du projet de recherche conjoint "Fabeko" était le développement et la mise en œuvre d'une technologie alternative in situ et sur site pour l'assainissement des PFAS de la zone vadose et/ou du sol excavé en utilisant un procédé de lavage du sol avec des composés polymères biodégradables (cf. figure 1). Des composés biopolymères tensioactifs spécialement développés sont utilisés pour séparer efficacement le PFAS du sol. Le lixiviat qui en résulte est transféré dans les eaux souterraines puis retiré hydrauliquement de l'aquifère au moyen d'un système de drainage. La technologie développée permet la mobilisation in situ/ et sur site des composés polyfluorés et perfluorés de la couche arable sans détruire la structure physique du sol et sans séparer le sol en différentes fractions granulométriques.

Au cours du développement approfondi du procédé, l'efficacité de cette approche technique de remédiation a d'abord été prouvée dans des tests de colonne de sol à petite échelle, puis dans des lysimètres de sol et finalement dans une application pilote de 3 semaines sur le terrain.

Dans le lysimètre ainsi que dans les essais à l'échelle pilote, de nouvelles méthodes d'analyse, telles que les analyses TOP, AOF et EOF, ont été utilisées en plus de l'analyse à paramètre unique PFAS pour quantifier de manière exhaustive l'efficacité de la technique d'assainissement développée.

Les résultats montrent que les acides carboxyliques perfluorés à chaîne courte peuvent être lessivés de la couche arable simplement en utilisant de l'eau comme percolat. Cependant, les acides carboxyliques avec une longueur de chaîne > C8 et les acides sulfoniques se sont avérés moins mobiles. L'étude a fourni des preuves solides que ces substances peuvent être mobilisées au moyen des biopolymères spécialement développés. Au cours des essais au lysimètre, une réduction de 95 % de la concentration de PFAS dans le sol a été obtenue en appliquant la méthode nouvellement développée. Dans le test sur le terrain, les concentrations des PFAS ont été réduites de > 80 % en une application de 3 semaines.

Dans le cadre du développement du procédé, des solutions techniques ont été créées pour inhiber les réactions géochimiques qui l'accompagnent (notamment en ce qui concerne la liaison non spécifique des cations polyvalents) et la croissance de la biomasse lors de l'utilisation des structures polymères facilement biodégradables. L'efficacité de ces mesures d'accompagnement a été testée avec succès au lysimètre.

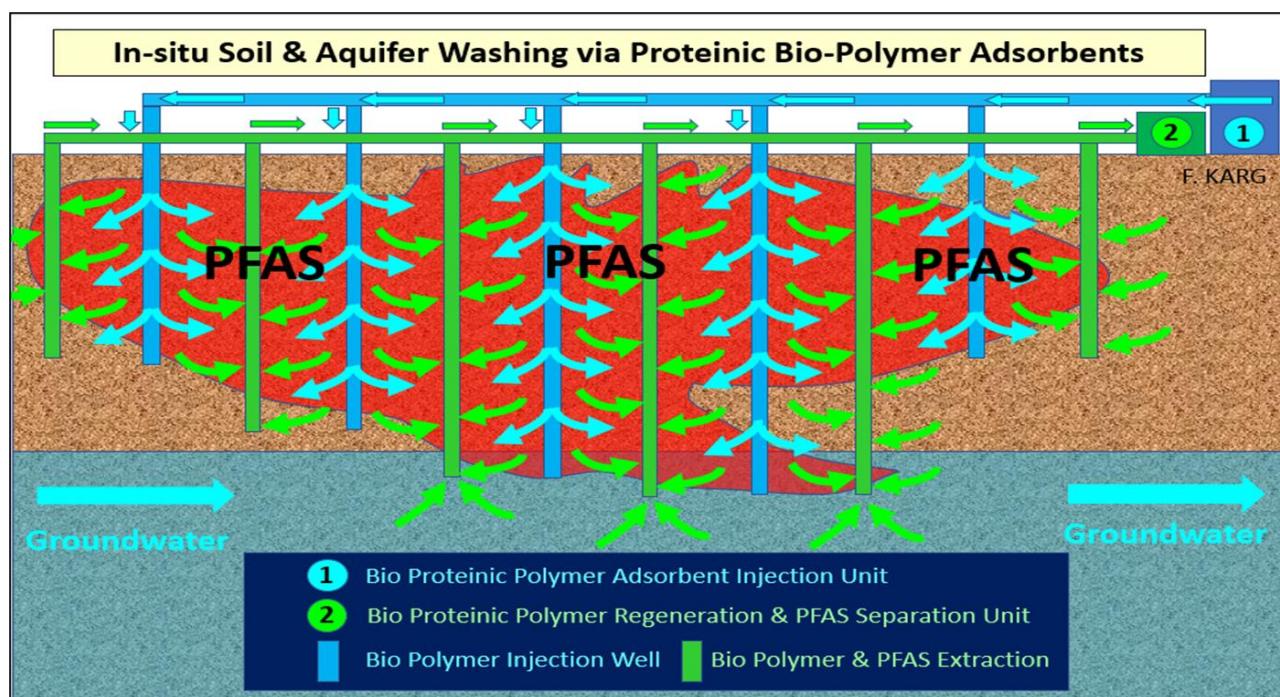


Fig. 1 : Lavage du sol in situ pour la décontamination des PFAS avec des biopolymères protéiques.

Le procédé développé offre la possibilité de traiter les sols contenant des PFAS in situ dans la zone vadose en mobilisant les PFAS et en traitant ensuite les eaux souterraines à l'aide d'une technologie de pointe. Une autre application possible de la méthode développée qui est actuellement appliquée sur le terrain est le traitement ex-situ de matériaux de sol excavés contenant des PFAS en appliquant le biopolymère dans des cycles de lavage. La technologie de lavage des sols peut très bien être associée à la technologie de stabilisation in-situ des PFAS par injection de matière solide de charbon actif. Les options d'application de cette approche technique sont discutées dans la présentation.

Sustainable In-situ Treatments of PFAS in contaminated Soil and Groundwater, Washing with Protein Bio-polymers and Stabilization by GAC high pressure Injection

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Due to their high chemical stability, mobility in the environment and bioaccumulation potential, poly- and perfluorinated alkylated substances (PFAS) represent a highly relevant pollutant group with regard to their hazard potential. Across Germany and other countries, numerous sites with PFAS contamination have been identified, some of which cover extensive areas, such as the contamination of topsoil on several hundred hectares of arable land in the Rastatt area (Baden-Württemberg, GER), where PFAS containing paper composting sludge was used as fertilizer. At present, no in-situ remediation technology for PFAS-contaminations in the vadose zone was available. State of the art remediation strategies mostly comprise traditional and very expensive methods like excavation of contaminated soil (& thermal treatment or dumping) or pump-and-treat as a hydraulic containment measure for contaminated groundwater.

The objective of the joint research project "Fabeko" was the development and implementation of an alternative in-situ and on-site technology for the remediation of PFAS from the vadose zone and/or excavated soil using a soil washing process with biodegradable polymer compounds (cf. Fig. 1). Specially developed surface active biopolymer compounds are used to effectively separate the PFAS from the soil. The resulting leachate is transferred to the groundwater and subsequently hydraulically removed from the aquifer by means of a drainage system. The developed technology enables in-situ- and on-site mobilization of poly- and perfluorinated compounds from topsoil without destroying the physical structure of the soil and without separating the soil into different grain size fractions.

During extensive process development, the effectiveness of this technical remediation approach was proven first in small scale soil column tests, then in soil lysimeters and eventually in a 3 week pilot application in the field.

In the lysimeter as well as pilot-scale test new analytical methods, such as TOP, AOF and EOF analyses were used in addition to PFAS single parameter analysis to comprehensively quantify the effectiveness of the developed remediation technique.

The results show that short-chain perfluorinated carboxylic acids can be leached from the topsoil simply by using water as percolate. However, carboxylic acids with a chain length > C8 and sulfonic acids proved to be less mobile. The study provided strong evidence that these substances can be mobilized by means of the specially developed biopolymers. During the lysimeter tests a 95%-reduction of the PFAS concentration of the soil was achieved by applying the newly developed method. In the field test PFAS concentrations were reduced by > 80% within a 3-week application.

As part of the process development, technical solutions have been created to inhibit the accompanying geochemical reactions (especially with regard to the non-specific binding of polyvalent cations) and the biomass growth when using the readily biodegradable polymer structures. The effectiveness of these accompanying measures has successfully been tested in the lysimeter.

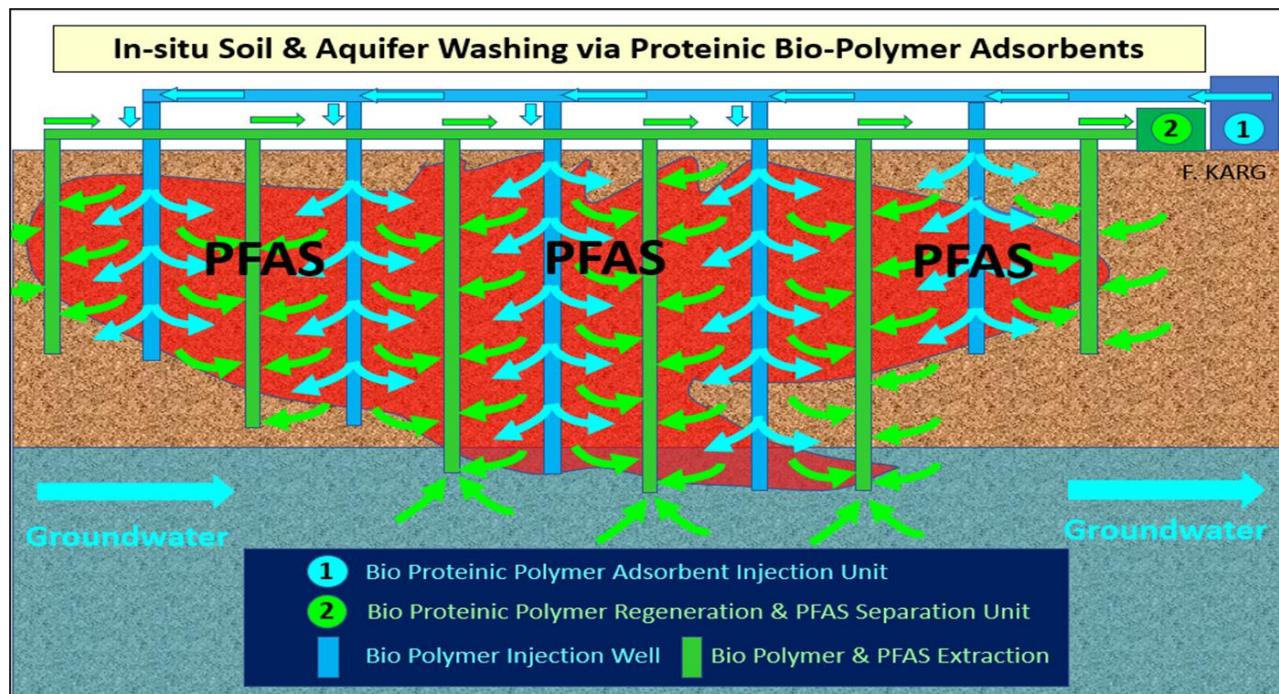


Fig. 1: In-situ soil washing for PFAS decontamination with proteinic Bio-polymers.

The developed process offers the possibility to treat PFAS-containing soils in-situ within the vadose zone by mobilizing PFAS and subsequently treating the groundwater using state-of-the-art technology. A further possible application of the developed method that is currently applied in the field is the ex-situ treatment of excavated PFAS-containing soil materials by applying the biopolymer in washing cycles.

The soil washing technology may well be combined with the in-situ PFAS stabilization technology by injection of solid activated carbon matter. The application options for this technical approach are discussed in the presentation.

Short overview of pfas remediation methods for soil and water

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In 2021, RSK Benelux was tasked by OVAM to perform a limited literature study and to conduct a market research in Flanders (Belgium) and surrounding regions. The market research was done by interviewing land management organisations, soil treatment centres and specialists in Flanders, the Netherlands and Germany.

The presentation gives a short overview of treatment technologies, promising and proven.

Then the results of the contacts with soil treatment centres are presented: soil washing in Flanders, thermal soil treatment in the Netherlands and soil treatment in Germany.

Treatment of soil contaminated by PFAS: feedback from two large-scale remediation projects

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The method of soil washing is currently the only economically feasible cleaning method for PFAS-contaminated soils on large scale.

A total of about 430,000 t of soil contaminated with PFAS and HCs were washed from August 2018 to October 2021 at the site of a former refinery in Bavaria. The cleaned soil could then be reinstalled. The site-specific costs for the complete soil treatment including excavation and reinstallation were about 50 EUR per ton.



Fig. 1: Soil washing plant for PFAS and HC contaminated soil at the site of a former oil refinery in Southern Germany

Since September 2022, Züblin Umwelttechnik GmbH has been operating another soil washing plant in Northern Germany, which was specially designed to treat soils contaminated with PFAS. Up to 800,000 tons of sandy soil with around 10% fines will be washed and refilled on site in the next years.

Data obtained from these two large-scale remediation projects are presented. The technical concept and challenges to treat 1,000 to 3,000 t of material per day are discussed, along with regulatory obstacles and the boundary conditions to process these quantities while limiting the emissions. A special focus lies on the heterogeneous nature of input material, with varying contaminant load, soil quality and soil structure as well as affiliated contaminants. In both attempts, the washing water is circulated in a closed water cycle, therefore intensive efficient treatment of sludge and polluted water is required.



Fig. 2: Soil washing plant for PFAS contaminated soil in Northern Germany.

One of the critical issues in soil washing is to optimise the washing process so that the washing liquid serves as the predominant contaminant sink. In this way, consistently good washing results are achieved, and all output fractions can be successfully processed while water consumption is kept to a minimum.

The complex nature of the contaminant itself, the multitude of singular PFAS substances and precursors, further complicate operation and controlling. Progress shows that soil washing can even be a viable tool to treat the fine fraction, and various strategies in different treatment steps have been tested on multiple scales, including several washing agents. Since after treatment, the washed soils can be safely reused on site, large emissions from transportation and properly disposing the PFAS-materials is avoided while increasingly limited landfill space is maintained.

Future development is required for expanding the applicability of the method to highly challenging materials such as high fines and/or high organic content. Further projects for washing of PFAS contaminated soils in Europe are already in the planning stage.

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**PFAS-contaminated soil management :
Learned lessons from soil-washing treatment strategies implemented in France and Belgium**

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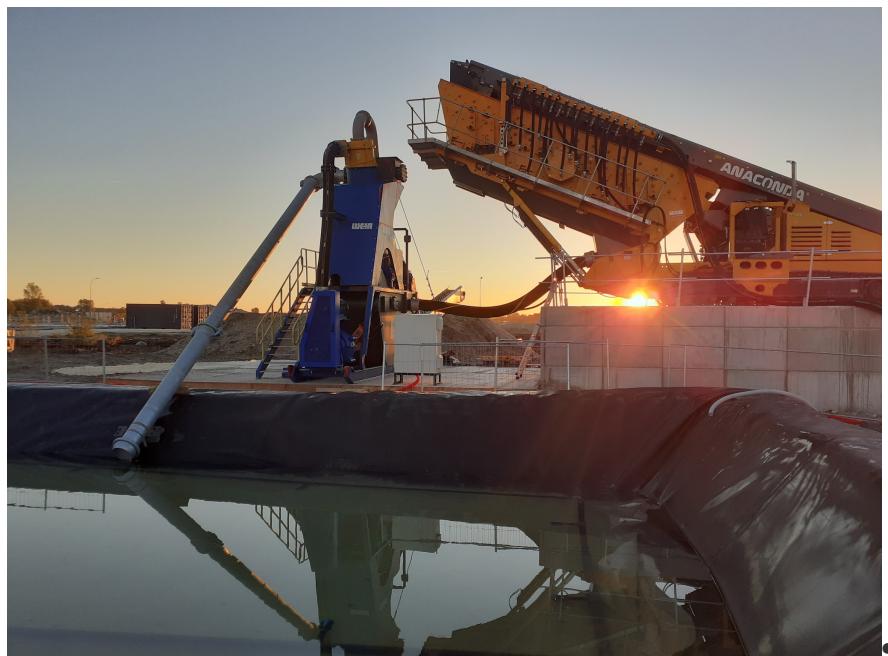
1. Introduction and study objectives

Perfluorinated alkyls constitute a large chemical class characterized by the total or partial substitution of hydrogen atoms with fluorine atoms around the carbon. It is estimated that there are over 9,000 substances belonging to this chemical class, all of which are of anthropogenic origin. These chemicals have been synthesized since the end of the 1940s and are used in more than 200 industrial and domestic applications (waterproofing of textiles, leather and packaging, fire-retardant foams, electronics industry, synthesis of fluorinated polymers, etc.).

Among perfluorinated alkyls, two sub-families have recently attracted the scientific community's attention: carboxylates (PFCA) and sulfonates (PFAS). These substances have been found in many environmental compartments (biota, sediments, soils, atmosphere, rainfall, water, etc.). This *ubiquitous* presence arises from 1) the direct use of these compounds; 2) the biological or chemical degradation of other perfluorinated alkyls (also called precursors), which are their ultimate metabolites.

As part of a French contaminated site rehabilitation project, SARPI Remediation implemented a soil-washing treatment for more than 6,000 m³ of soil impacted by PFOS.

A feasibility and dimensioning step was conducted by the internal laboratory and the on-site implementation was scheduled immediately after that.



At its soil treatment platform in Grimbergen, Belgium, in 2022, SARPI Remediation treated over 20,000 m³ of soil contaminated with PFAS by soil washing."

The aim of this presentation is to share our feedback and experience on the soil washing of PFAS and particularly on the washing water treatment. Our presentation will show the results of a study conducted on 30 different media (GAC, resins, clay..) and the behavior of long and short chain PFAS for each media tested. On top of this, we also assess the impact of TOC (Total Organic Carbon) concentration and origin on the sorption efficiency of selected media.

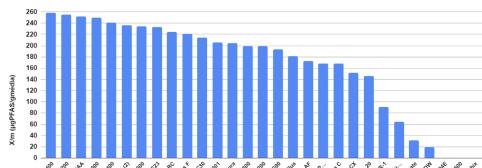
2. Results & discussion

The presentation will show the different trials carried out for the on-site implementation:

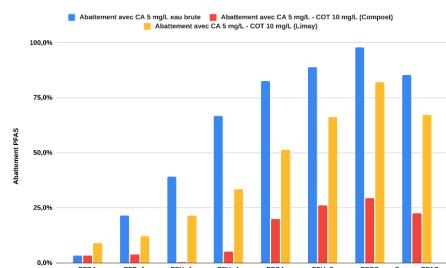
- Laboratory feasibility test: soil washing, stabilization, solidification
- Test and comparison of 30 different supports (activated carbon, clay, resins) in order to evaluate the best support for wash water loop treatment.
- The Total Organic Carbon (TOC) effect on the activated carbon adsorption capacity in the wash water.



Absorbents comparison : Sum of PFAS



Compounds comparison (% reduction)
GAC = 5 mg/L - TOC = 10 mg/L



Couplage OI basse pression/electro-oxydation anodique pour le traitement des eaux contaminées par des PFAS

Coupling low pressure RO with anodic electro-oxidation for the treatment of PFAS contaminated water

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Les substances per- et polyfluoroalkylées connues sous le nom de PFAS sont présentent dans des produits de la vie courante tels que les emballages alimentaires ou les mousses anti-incendie. Du fait de leur grande stabilité chimique, ces substances s'accumulent dans les écosystèmes. La présence de PFAS dans les eaux de surface est de plus en plus relevée, et des traces de ces composés ont été détectées dans l'eau potable et le sang humain. La présence de PFAS dans les eaux naturelles constitue donc un enjeu environnemental majeur, en raison de leur impact potentiellement nocif sur la santé humaine. L'électro-oxydation anodique (EOA) est une technique de dépollution (réaction radicalaire à la surface des électrodes) prometteuse pour les eaux contaminées par les PFAS. Toutefois les études antérieures à base d'électrodes en diamant dopé au bore (BDD) ont montré que l'efficacité de EOA augmente avec la teneur en PFAS des eaux (importance du phénomène de diffusion, i.e. diffusion-limited electrochemical oxidation). La décontamination des eaux polluées, notamment par les PFAS, peut être envisagée en mettant en place des filières de traitement incluant une phase de pré-concentration suivie d'une étape de dégradation. Dans ce contexte, le couplage membrane osmose inverse (OI)/électro-oxydation anodique apparaît comme une approche viable. Les membranes OI dites basse pression présentent un intérêt grandissant, car elles peuvent être mises en œuvre à un coût énergétique plus faible que les membranes classiques OI, tout en offrant une efficacité attendue supérieure à celle des membranes NF. Il est donc important de vérifier leur efficacité, en particulier pour les PFAS de faible masse moléculaire. De plus, le rôle du colmatage organique associé aux matières organiques dissoutes (MOD) et au biofouling sur le pouvoir de réjection des membranes OI reste partiellement expliqué, ce qui nécessite une approche de recherche fondamentale pour mieux le cerner.

Comme rapporté dans la littérature nos travaux réalisés sur des coupons membranaires (module plan) à partir de solutions synthétiques et d'eaux réelles ont confirmé la très bonne réjection des PFAS par les membranes d'OI vierges. De l'ordre de 98% pour les chaînes longues, ainsi que pour les chaînes plus courtes. Ces propriétés de rétention sont maintenues avec des membranes OI présentant un biofilm issues de modules prélevés sur des unités pilotes installées dans des usines de production d'eau potable. La présence d'un biofilm ne semble pas avoir d'incidence significative sur l'efficacité de rétention des PFAS par les membranes d'OI dans nos conditions expérimentales. Le biofilm a été caractérisé au moyen de techniques analytiques telles que l'analyse infrarouge, la pyrolyse couplée à la spectrométrie de masse (GC-MS), l'analyse élémentaire par spectrométrie de plasma à couplage inductif (ICP) ainsi que l'analyse des PFAS.

Les expériences conduites avec une cellule d'électro-oxydation BDD (anode et cathode identiques) en mode recirculation à partir de solutions synthétiques de PFAS (acides perfluoroalkyls carboxyliques allant de C4 à C9 et 3 acides perfluoroalkyls sulfoniques C4, C6 et C8) en présence de sulfate de sodium, ont montré que la cinétique de dégradation de ces molécules augmente avec la longueur de la chaîne carbonée. Pas de différence significative n'a été relevée entre les composés portant un groupement carboxylique ou sulfonique. La dégradation des PFAS conduit à la formation de sous-produits résultant de la perte successive de -CF₂, e.g. la dégradation du PFOA (C8) conduisant à la formation successive du C7, du C6, du C5 et du C4. Dans nos conditions expérimentales (8 mA/cm², 20V), la dégradation du C4 s'est avérée négligeable. Le traitement de solutions de mélange de PFAS peut donc conduire à une augmentation des teneurs en molécules à petite chaîne suite à la dégradation des composés à plus longue chaîne. La présence

en solution de matière organique dissoute (substances humiques isolées d'eaux de surface) avec des concentrations comprises entre 5 et 20 mg/L, a peu d'impact sur la cinétique de dégradation du PFHxA (C6, 50 µg/L), quel que soit le caractère aromatique de la MOD (SUVA = 1 à 5).

Différents facteurs tels que la densité de courant, l'électrolyte et les concentrations initiales de PFAS ont été modifiés et étudiés pour déterminer les conditions de fonctionnement optimales pour l'élimination des PFAS. Les résultats ont montré que des électrolytes sous forme de sels d'hydroxyde (NH_4OH , KOH et NaOH) ont la même efficacité vis-à-vis de la dégradation du PFOA (C8) que des sels de sulfate ($(\text{NH}_4)_2\text{SO}_4$, K_2SO_4 et Na_2SO_4) et des sels de carbonate ($(\text{NH}_4)_2\text{CO}_3$, K_2CO_3 et Na_2CO_3). De plus, le type de cation dans l'électrolyte n'a aucun impact sur la dégradation et la défluoration du PFOA. Des travaux complémentaires sont en cours sur des concentrats d'osmose inverse.

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Unpacking a responsible manufacturing approach for Fluoropolymers

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TECHNICAL SUMMARY

Fluoropolymers are high-performance plastics material with a unique combination of properties that makes them the building blocks for an extraordinary range of product and industry applications. Their unique properties—including durability, mechanical strength, inertness, thermal stability, and resistance to chemical, biological, and physical degradation—mean they are used across many sectors where they can withstand the most challenging and high-stress conditions where failure is not an option.¹

Fluoropolymers are critical to a range of industries, have no suitable alternatives in the majority of uses, do not pose risks to human health and the environment when used for their intended purposes, and are manufactured responsibly.

As high molecular weight molecules with unique properties attributable to the strong Carbon–Fluorine bonds, the strongest bond in organic chemistry, fluoropolymers are unlikely to be replaced by an alternative material, regardless of the resources invested. Applications requiring a combination of these properties, fluoropolymers are the superior and, often, only solution.

Fluoropolymers' health and safety profiles help ensure that they are safe, non-toxic, stable, and inert, among other characteristics, throughout their intended use. To supplement these characteristics, responsible manufacturing and reliable emissions control ensure that fluoropolymers are safe for their intended use for both people and the planet.

Every stage of fluoropolymer manufacturing—from the earliest stages of raw materials and monomers to the creation of polymers—is completed responsibly, with thorough management of raw materials, polymerization aids, and the resulting polymers that are used in various product applications. The vast majority of fluoropolymers are manufactured in a manner that enables them to meet the OECD's criteria for polymers of low concern.²

¹ [A critical review of the application of polymer of low concern and regulatory criteria to fluoropolymers](#)

² [A critical review of the application of polymer of low concern and regulatory criteria to fluoropolymers](#)

Guided by existing regulation, industry already employs robust responsible manufacturing practices, including control of gaseous, aqueous, and solid emissions throughout manufacturing and processing, stringent emissions abatement technology and waste disposal methods, and end of life management. Leading fluoropolymer manufacturers have invested in state-of-the-art emissions control facilities and have already demonstrated significant emission reduction, including fluorinated polymerization aid recovery for reuse, 99% reduction of air/water process emissions, and 99% plant emission reductions.³

Moreover, industry groups are currently engaged in rigorous research around further efforts to ensure responsible manufacturing and mitigating emissions. R&D investments have increased recyclability and re-use, core to the goal of a circular economy. Fluoropolymers are safe for their intended, essential uses, and it has been proven by good science, readily available data, and a demonstrable, measurable effort from industry leaders.

In this presentation, Cedric Triquet will unpack Chemours' approach to responsible manufacturing sharing advances in analytical and abatement technologies that support meaningful reductions in emissions.

³ [A critical review of the application of polymer of low concern regulatory criteria to fluoropolymers II: Fluoroplastics and fluoroelastomers](#)

Treatment and recovery of PFAS-contaminated water by means of atmospheric water-plasma processes

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The research project "Treatment and recovery of PFAS-contaminated water by means of atmospheric water-plasma processes" is processed by the cooperation between the Fraunhofer Institute for Interfacial Engineering and Biotechnology IGB (FhIGB) and the engineering office HYDR.O. GEOLOGEN UND INGENIEURE. The project is currently sponsored by the Federal Ministry of Education and Research of Germany, as part of the KMUi-fund "Resource efficiency and Climate Protection" for small and medium-sized enterprises (promotional reference: 02WQ1601A/B). The project started in July 2020 with the aim to develop an innovative treatment method to efficiently degrade PFAS compounds from contaminated ground- and seepage water.

Several reasons have led to the research approach of treating PFAS-contaminated water with atmospheric water plasma. First, atmospheric water plasma has proven to be a successful treatment method for the remediation of pollutants such as heavy metals or pharmaceuticals. Second, its application on PFAS already showed effectiveness in previous studies on standardized PFAS samples and on groundwater samples, especially when looking at PFOA and PFOS [1]. And third, the degradation mechanisms of PFAS caused by plasma processes are known and can lead to the mineralization of the substances [2]. The last reason is especially remarkable because remediation would not include residual PFAS as an end-product, which is the case for most alternative treatment methods. Looking at examples of contaminated sites, the deposition is currently a major issue, as many landfills do not accept PFAS-contaminated waste anymore. This is due to the challenging treatment of seepage water and hinders remediation procedures. Further, the treatment has the potential to degrade more pollutants, such as heavy metals or pharmaceuticals, from the groundwater matrix without an additional expense.

The technical features of the methodology are the DBD reactor, consisting of an outer high voltage electrode, an inner steel electrode, a cylindric glass column and the plasma generator. Additionally, monitoring devices for pH, ozone and temperature parameters and a surrounding set-up for sample circulation are installed. The water is pumped inside of the hollow steel cylinder and then runs down at its side. This is where the contact with the discharged plasma is taking place and where the PFAS substances are degrading through interaction with the reactive species of the plasma. The methodology was tested with standardized PFOA samples first ($1 \mu\text{g L}^{-1}$ and $10 \mu\text{g L}^{-1}$) and then applied to groundwater and seepage water samples from seven existing contaminated sites. These sampling locations comprised of varying additional contaminants and turbidities as well as a high range of PFAS-spectra and -concentrations.

The treatment methodology applied to standardized samples showed 70 % of degradation to the original substance PFOA and 60 % degradation of total PFAS. Best results were achieved with helium as a process gas, 250 V (intermediate circuit voltage) and a treatment duration of 120 min. The observed degradation products were in line with expectations based on Singh et al. (2019), who performed research on PFAS degradation pathways in non-thermal plasma (NTP) processes specifically. In contrast, samples from existing PFAS contaminated sites showed a high variation in results. Especially the sum of PFAS concentrations after the treatment were higher than in the initial sample for almost all treatment runs. Measurements were performed in the same institute as before,

which had reliable results in synthesized samples and were double-checked in another, accredited laboratory.

Main hypothesis for potential problem sources relate to analytical limitations, sorption effects and the sample matrix. First, it is possible that the samples entail a high amount of non-detectable PFAS. Even though there is a very large number of PFAS substances, in practice only 13 or a maximum of 27 are included in standardized measurements. These polymeric PFAS, precursors or simply non-detectable PFAS, may degrade to detectable PFAS compounds during the plasma treatment. Additionally, low-concentrated samples imply lower analytical certainties, which may lead to variations in the initial analysis. Second, sorption of PFAS on organic particles or at test vessels and parts of the treatment reactor may represent a yet unquantified and dynamic influence on the subsequent treatment and the analysis. And third, the sample matrix, including other contaminants, may cause substance interactions which are hardly comprehensible.

As a response, different adaptions of analysis and methodology were tested. First, ESR Spectroscopy was performed and proved that the effect of the plasma treatment is existent and measurable. Second, AOF, TOP-Assay and F-NMR measurements were tested to assess the amount of potential undetected PFAS-substances. Unfortunately, these parameters did not provide reasonable results for the sample locations. In terms of sorption effects, F-NMR results suggested a reduction of CF2 by 18.6 % and CF3 by 16.3 % as a consequence of sorption to the treatment plant, highlighting the high importance of its consideration. In order to deepen knowledge on the degradation pathways and confirm mineralization of PFAS during the process, F-NMR spectroscopy was applied again. Analysis showed peaks which can be associated with NaF, increasing with a longer duration of the plasma process and thus indicating increasing mineralization.

Conclusively, three key messages can be derived from the results of treating PFAS-contaminated samples with atmospheric water plasma processes. First, degradation and mineralization of PFAS are possible with this treatment method. Second, mineralization is verifiable but not yet quantifiable. Third, site-specific matrix effects lead to substantially different results in treatment efficiency than before-tested standardized samples.

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**PFAS leaching from an AFFF-contaminated site:
rainfall simulation experiments on undisturbed soil columns reveal the contribution of
soil particles to the transport of PFAS**

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The topic of per- and polyfluoroalkyl substances (PFAS) is a late-breaking issue due to its high environmental relevance (toxicity, persistence and bio accumulation) and due to the detection of PFAS as contaminants in various environmental compartments including groundwater, surface water and soil. PFAS enter the environment via different pathways such as industrial, agricultural, and house-hold activities but as well through firefighting activities using aqueous fire-fighting foams (AFFF) which contain a high variety of PFAS molecules [1]. Especially firefighting training zones are therefore highly exposed to PFAS contamination in soil and ground water and can even pollute other regions since PFAS can be carried away from the source within the soil pore water [2] [3].

Previous studies have shown that the transport and retention of PFAS in soils depend on various factors such as the PFAS type (e.g. molecular structure and carbon chain length), the soil properties (e.g. amount of organic carbon, surface charge), and the pore water (e.g. degree of saturation, pH). Still, the retention and release mechanisms of PFAS on soil constituents are not completely elucidated, and an official, universal protocol for the determination of PFAS concentrations in the subsoil is not existing yet.

PFAS are commonly analyzed in the ground water or/and in the soil assuming that PFAS transport only occurs directly by desorption into solution and not considering the co-transport due to strong adsorption of PFAS to eroded particles as a possible solid carrier phase next to the liquid phase [4] [5]. Still, this assumption might not be correct including the already observed spatial and temporal interactions between the soil and the soil pore water like e.g. soil particle mobilization caused by wetting and drying cycles or macro pore flow [6] [7] [8] [9]. Especially with regard to the surfactant characteristic of PFAS, particle mediated transport of PFAS should be further investigated since only few projects on this topic have been carried out so far [10]. Based on the findings of enhanced leaching of one PFAS molecule because of its adsorption to released particles in artificially contaminated, homogenously compacted soil columns [11], we now increased the complexity to more real-life conditions working with undisturbed soil columns from a firefighting training site contaminated by various PFAS through the use of AFFF. We expose these columns to simulated rain falls of different rain intensities and dry periods and afterwards analyze the drained effluent water for PFAS once in the liquid phase and once adsorbed to eroded soil particles. In preliminary tests we were already able to detect even long-chain PFAS in the collected leachate (although their mobility in soil is expected to be low), which could be attributed to colloid leaching and colloidal particles as a carrier phase for long-chain PFAS.

The separated analysis of PFAS in the liquid and the solid phase will help us to better understand the contribution of soil particles to the transport of PFAS in the soil pore water and therefore possibly

point out the risk of higher pollutions of remote areas, and whether this contribution should also be taken into account in the determination of PFAS in the subsoil in the future.

key words: PFAS transport, particle mobilization, leaching experiments

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LIFE CAPTURE: COMBINING NOVEL ANALYTICAL PROTOCOLS WITH EFFECTIVE TECHNOLOGIES FOR SUSTAINABLE PFAS REMEDIATION

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Per- and polyfluoroalkyl substances (PFAS) are a family of man-made chemicals that have been widely used since the 1950s. PFAS is a class of thousands of substances (>6,000 compounds) characterized by extreme persistence and different behaviors/mobility in the environment. The efforts towards the substitution of old generation PFAS with short-chain and new generation PFAS has brought new concerns as these substitutes are once again poorly known, with similar mobility in the environment, and could lead to relevant health effects.

The overarching objective of LIFE CAPTURE is to develop sustainable management methods for dealing with Per- and polyfluoroalkyl substances (PFAS) in soil and groundwater. This targets concrete issues and challenges that are currently encountered when dealing with PFAS-contamination. These relate to characterization, assessment and mitigation of the contamination.

Up to now, far too little attention has been paid to the methodology for the PFAS analysis. Nowadays, only a very limited number is taken in account (approx. 40) during quantification, leading to sometimes severe underestimations of the PFAS presence in soil, water and other matrices. Therefore, LIFE CAPTURE intends to develop a protocol for any type of PFAS contamination by enlarging the spectrum of the PFAS family that can be identified and analyzed. LIFE CAPTURE proposes a robust protocol that follows a staggered approach to: quantify whether any kind of PFAS is present; quantify the concentrations of a set group of well researched PFAS; determine if there are significant concentrations of other PFAS present; quantify and qualify those other PFAS.

During site investigation it is currently considered best practice to only look at concentrations. The use of flux-measurements would be a valuable addition, allowing more direct measurement of exposure and risks. It can also be used to drastically enhance the efficiency of mitigation measures such as remediation. Within this project, passive flux samplers that can be integrated in the new analytical protocol will be demonstrated.

Due to PFAS's chemical properties, most existing remediation technologies don't provide adequate sustainable treatment solutions. We propose a toolkit of promising innovative remediation technologies for PFAS. Technology trains, including technologies such as foam fractioning, biological treatment, advanced oxidative practices and ventilated activated carbon, will be tested to optimize the remediation approaches in order to eliminate the contaminants from the environment. Notably, the LIFE CAPTURE project will pilot test the (cost-) effectiveness of tailored Advanced Oxidation Processes in eliminating PFAS from groundwaters in one of the most impacted aquifer in north-east Italy. This aquifer is currently showing sum of PFAS concentrations of thousands of ng/l .

The assessment whether a certain PFAS-contamination is harmful is typically governed by a regulatory framework. We propose to develop a pragmatic risk and effect assessment approach for varying soil types. This will be based on existing standardized ecotoxicity tests. Special attention will be given to risk assessment for mixtures of different known or unknown PFAS.



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